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Tree biomass and biodiversity relationship in a mixed Mediterranean forest in Spain

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RESUMEN

La biomasa de árboles y su relación con la diversidad en bosques mixtos se ha convertido en uno de los temas de investigación más interesantes para los ecólogos en las últimas décadas debido a la importancia de los bosques mixtos para una mejor provisión de servicios ecosistémicos. La pregunta “¿Producen los bosques mixtos más a medida que aumenta la diversidad de árboles?” Ha sido objeto de muchos estudios que llevan a resultados no concluyentes. Este estudio se realizó para contrastar el resultado de estudios anteriores mediante la investigación de la biomasa de árboles y la relación de diversidad en un rodal mixto de bosque mediterráneo (Llano de San Marugan, España), tanto a nivel de masa como de árbol individual. Se analizaron diversos modelos que se ajustaron a partir de ecuaciones de regresión lineal y no lineal para determinar la relación entre la biomasa de los árboles y la diversidad. Se utilizaron 10 índices de diversidad que se pueden clasificar en 3 categorías: índices de riqueza de especies (Sm, Sn, D, E); Índices de composición / mezcla de especies (Mi, MS, S); Los índices estructurales verticales (W, A, TH) como variables predictoras de los modelos con el objeto de caracterizar diferentes estructuras de diversidad en el rodal. Nuestro resultado reveló que la relación entre la biomasa y la diversidad de árboles varía entre las especies. Una combinación de la relación negativa del índice D-Berker-Parker (abundancia de especies dominantes) y la relación positiva de TH (heterogeneidad de la altura) explica la variación de la biomasa a nivel rodal y para *Pinus pinea*. La biomasa de las especies de *Quercus* (*Quercus faginea* y *Quercus ilex*) se relaciona positivamente con la proporción de especies en área basimétrica (Gp); los índices de diversidad probados no mostraron ninguna relación con la biomasa de las especies del género *Quercus*.

Palabras clave: diversidad de especies arbóreas, índices de diversidad, riqueza específica, composición específica, estructura vertical del rodal.

ABSTRACT

Tree biomass and diversity relationship in mixed forest has become one of the attractive research subjects for ecologist in recent decades due to an importance of multicultural mixed forest for better provision of goods and services than monoculture. The questions “Does mixed forest produce more productive and the productivity increase as tree diversity increases?” have been subject of many researches that lead to two contrast results. This study was conducted to contrast the result of previous studies by investigating the tree biomass and diversity relation in Mediterranean multicultural mixed stand, Llano de San Marugan, Spain, at stand and individual species level. A variety of models that developed from linear and nonlinear regression equations were employed to reveal tree biomass and diversity relation. 10 diversity indices that falls in 3 categories: species richness indices (Sm, Sn, D, E); species compositional/mingling indices (Mi, MS, S); vertical structural indices (W, A, TH) were used as predictor variables for the models to characterize different structure of diversity in the stand. Our result revealed that tree biomass and diversity relation varies among species. A combination of negative relation of D- Berker-Parker index (abundance of dominant species) and positive relation of TH (height heterogeneity) explains the variation of biomass at community level and for *Pinus pinea*. Biomass of *Quercus* species (*Quercus faginea* and *Quercus ilex*) was positively related with basal area proportion of species (Gp); the tested diversity indices didn't show any relation with biomass of *Quercus* species and *Juniperus thurifera* as concerned by metrics and models in this study.

Key words: tree biomass, tree diversity, mixed forest, diversity indices, species richness, species composition, stand vertical structure.

1. - INTRODUCTION

Mediterranean forests are characterized by a remarkable set of features that make them naturally and aesthetically attractive, but also quite fragile (Scarascia-Mugnozza *et al.*, 2000). Mediterranean forest is a multi-functional, providing a wide range of goods and services for society ranging from products with high market value (fuelwood, cork, mushroom, pinecones etc) and non-market value ecosystem services (soil and landscape protection, water regulation, biodiversity conservation, carbon dioxide fixation, recreation, aesthetic view etc). The latter is more significant than their productive value, especially their significant role for carbon sequestration (del Río *et al.*, 2017). One of notable characteristics of Mediterranean forests is its rich biodiversity, reflected by high genetic variability, exemplified by the large number of tree species in comparison to Nordic forests resulting from the survival of many conifer and broadleaf species during the glacial periods. Long-term exploitation (manipulation) of trees and forestland since ancient times is another feature of Mediterranean forest which results in the dispersion of species as *Pinus pinea*, *Castanea sativa*, and *Quercus suber* all over the Mediterranean basin (Scarascia-Mugnozza *et al.*, 2000). Dry, hot, harsh climate along with long lasting and frequent droughts, pest and disease, increasing the risk of large-scale fires and severe water scarcity are main challenges for the Mediterranean forests which largely impact on forest health, growth and productivity. The role of mixed forest for promoting forest productivity while coping with these challenges has been increased in Mediterranean region in recent decades.

Multicultural mixed forest have been taken a great attention in recent decades due to its greater provision of goods and services, high ecological value in comparison to monoculture forest (Pretzsch & Schütze, 2014; Riofrío, *et al.*, 2017). Mixed forest is defined as a forest unit of at least 0.5 ha that composes at least two tree species at any developmental stage, shares common resources (water, light, soil nutrients) and its structure and component species are altered over the time (Bravo-Oviedo *et al.*, 2014). Main characterizations of mixed forests are described not only by better protection, preservation, maintaining and monitoring of biodiversity but also have high resistance capacity against both natural and anthropogenic disturbances such as climate change, storm, pest and disease, air pollution and its consequences. Economic importance of mixed forest is un-negligible because of its multi-use, multi-source than pure stands (Knoke *et al.*, 2005).

The loss of biodiversity imposed by anthropogenic and climatic change has brought the importance of diversity under the control worldwide over the past 25 years (Hooper *et al.*, 2012; J. Liang *et al.*, 2016; Szwagrzyk *et al.*, 2007) after the Earth Summit of world governments in Rio de Janeiro in 1992 (Evans, 2016). Since that time, scientists have been taken into account the importance of biodiversity on many ecosystem functioning and service such as productivity, stability, sustainability, sinking carbon dioxide, preserving soil fertility, controlling pest outbreaks, retaining water, and so on (Baskin, 1994). Among them, the importance of tree species diversity on biomass productivity has been studied based on the variety of genes, species, or functional traits of organisms in hundreds of types of ecological communities (Fraser *et al.*, 2015; Jingjing Liang *et al.*, 2016). A series of biodiversity-ecosystem functioning studies have revealed that biodiversity (including taxonomic, functional and phylogenetic diversity) promotes the functionality of ecosystems such as primary production, decomposition, nutrient cycling, trophic interactions and so on) and consequently supports a broad

range of ecosystem services (e.g. food production, climate regulation, pest control, pollination (Gamfeldt *et al.*, 2013; Mori *et al.*, 2017). However, contradictory results have been documented in the findings of previous researchers focused on relationship between species diversity and biomass: biomass decrease (Szwagrzyk *et al.*, 2007) or doesn't change (Grace *et al.*, 2016) with species diversity. In addition, numerous researches justified on loss of biodiversity ranks among the most pronounced changes to environment (Sala *et al.*, 2000), reduction of diversity along with species composition changes alter fluxes of energy and essential services that ecosystem provide to human such as production of food, pest and disease control, water purification and so on (Daily, 1997). Biodiversity are largely and irreversibly being degraded and lost globally due to direct drivers; i.e. habitat disturbance, habitat fragmentation, land use change, over-exploitation and the spread of alien species and indirect drivers; i.e. climate change, population growth, economic growth and increasing demand for food, materials, water and energy (Iranah *et al.*, 2018). The loss of biodiversity weakens species connections and impairs the ecosystems, leading to extinction of species and local populations, which will disrupt the capacity of ecosystem to contribute to human well-being and sustain future generations.

Tree diversity plays a fundamental role for forest diversity because it is often linked with major properties of forest ecosystem, leading to the possible enhancement of diversity of other forest assembles (Mori *et al.*, 2017) and providing required resources and suitable habitat for other forest species (Ozcelik, 2009). Diversity is generally defined by the variety of organisms including micro-organisms, plants, and animals in different ecosystems, i.e. deserts, grassland, forests etc. The most commonly used representation of ecological diversity is species diversity, which is defined by the number of species and abundance of each species living within a certain area (Liu *et al.*, 2018). The species coexisting in a certain area are interconnected and dependent on one another for survival, while doing so; they perform important ecosystem functions and offer different ecosystem services for human life and society: provisional service (products obtained from ecosystem: many different type of food, fresh water etc); regulating services (the benefits obtained from the regulation of ecosystem processes: air quality and pollination); cultural service (the non-material benefits that people obtain: spiritual enrichment, recreation and aesthetic experiences); supporting service needed to maintain other services (i.e photosynthesis and nutrient recycling). The provision of ecosystem for such goods and service depends basically on functions performed by living plants (Tilman *et al.*, 1997).

Two main mechanisms explain the reasons that biodiversity influence on productivity: selection effects and complementary effects. Different plant species in a mixture have different physiologies, morphologies and life history traits might allow them to fully utilize limiting resources at different space and time than a monoculture of any species (Tilman 1997). For instance, some tree species have more ability to adapt and grow better in cooler and wetter environmental condition while others grow better in hotter and drier environment. If these species grow together in a mixture, these complementary characteristics of both species lead to greater productivity across the whole grown season than either species grows alone. Similarly, tree species that have different root morphologies occupy different soil profile which potentially allowing them to exploit soil resources from different soil depth. However it should be noted that these complementary occur solely when co-existing species exhibits various forms of niche differentiation that allow them to capture resources in different space or time (Cardinale

et al., 2007; David Tilman, 1999). Another mechanism that diversity effects on productivity is selection effect (sampling effect) which describes species specific effect on biomass: a greater productivity in more diverse communities is due to the most productive species which become dominant in the community due to competition. The likelihood of becoming a high productive species increases as diversity increases. Thus this causes in the increment of the total productivity of the community. Such considerations have led to the general perception of having higher productivity in an area where more plant species co-exist.

Forest is 3-dimensional system whose structure is a key element in ecosystem functioning and biological diversity by regulating resource related forest functioning (light, water, soil nutrients supply, capture, use), intra and inter specific interactions (Brockerhoff *et al.*, 2017a; del Río *et al.*, 2018), regeneration pattern, consequent self-thinning and past and present disturbance events (Bohn & Huth, 2017; Zhang *et al.*, 2018). Stand structural diversity leads to increase species richness and contributes to forest stability and integrity (Wang *et al.*, 2016). Stand structural diversity combines the concepts of species richness (diversity), species composition (mixture), and spatial diversity (tree positioning) and size differentiation (Bravo & Guerra, 2002). Accordingly three distinct types of stand structural indices and methods have frequently been purposed in preceding literature for explaining the influence of stand structural diversity on productivity and functioning of forest stand: i) species richness - Simpson index (1949), Shannon index (1948), Berger-Parker index (Berger *et al.*, 1970) and Evenness index (Kohn, 1977); ii) species composition indices – Mingling index (Füldner, 1995), Spatial diversity status (Gadow & Hui, 2002) and Segregation index (Pielou, 1977); iii) tree distributional indices including horizontal and vertical patterns and size differentiation - Aggregation Index (Clark *et al.*, 1954), Uniform Angle Index (Gadow *et al.*, 1998), Vertical Species Profile (Pretzch 1995b), Height differentiation index (Gadow 1993). Since forest structure is determined in 3 dimensions, it is appropriate to analyze the effect of tree diversity on biomass by the metrics that can fully address 3-dimensionality of mixed forest structures.

2. - OBJECTIVES

In this study, we addressed the question “Does stand diversity impact on biomass?” by examining the relationship between tree biomass and diversity indices (species richness, species compositional, horizontal and vertical structural indices) in the Mediterranean multi-species mixed forest, Llano de San Marugan, Valladolid, Spain. The specific aims were:

- To calculate biomass of individual trees by different compartments (stem, thin branches + needles, medium and thick branches) and in total
- To compute the tree diversity indices to represent the diversity of the stand
- To determine the relationship between tree biomass and diversity.

3. - MATERIAL AND METHODS

3.1. Study site

The study was conducted in a Mediterranean mixed forest stand, located in Llano de San Marugan, Valladolid, Castile and Leon, Spain. Valladolid has a continental

Mediterranean climate with cold winters and hot summers. The average annual temperature ranges between 11 and 12°C. Fog is frequent in the long, cold winter, while summers are dry and hot with average temperature around 30°C. Precipitation falls irregularly throughout the year with a minimum in the summer and a maximum in spring and autumn, with maximum of 400 mm. In a mixed forest stand, a marteloscope was installed in 2015 covering 1 hectare (ha). The marteloscope (a square of 100 by 100 m) was divided into 16 subplots (hereafter referred as quadrants), 25 x 25 m length, as shown in Figure 1. Within each quadrant, locations of all trees were recorded in 4.55728° W and 41.43948° N geographic coordinates and their species were identified and corresponding diameter at breast height (dbh, in cm) and, total height (in m). The diameter at breast height and total height were measured from trees whose diameters were greater than 5 cm. The study workflow is shown in Figure 2 and detailed explanations are given in following sub-sections.

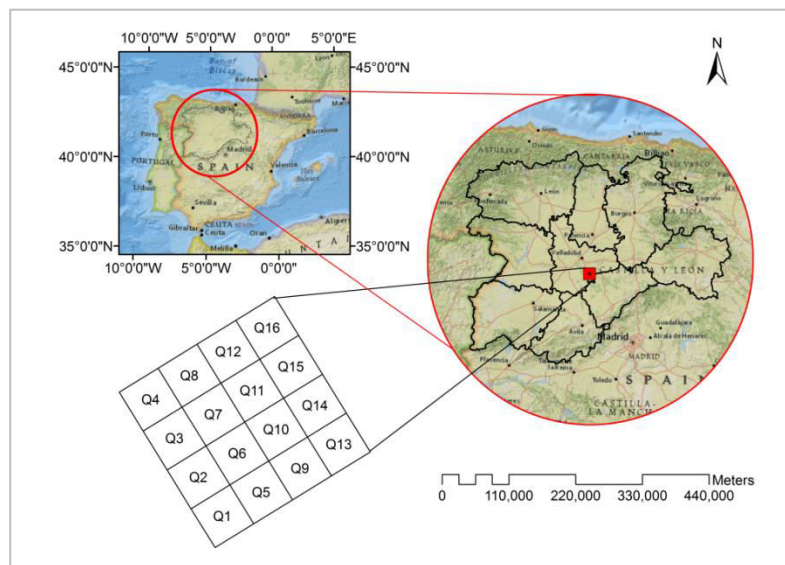


Figure 1. Study area location of Marteloscope of Llano de San Marugan (Valladolid, Spain)

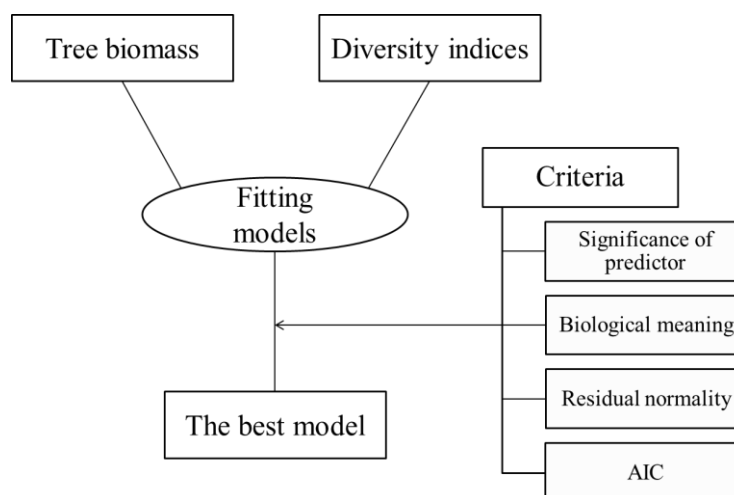


Figure 2. Workflow for the study of tree biomass and diversity relationship

3.2. Biomass estimation

Biomass in the different components of the tree e.g., stem and branches (thick, medium and thin+needles) and roots were calculated from dbh and height using existing relative allometric equations (Table 1) developed by (Ruiz-Peinado Gertrudix, Montero, & del Rio, 2012; Ruiz-Peinado, del Rio, & Montero, 2011). Tree component biomass values were computed for individual tree within each quadrat, and summed up to derive a summary of tree biomass for each quadrat. Total biomass obtained from sum of the biomass of all components.

Table 1. Biomass allometric equations by species

Species	Components	Model
Juniperus thurifera L.	Stem	$W_s = 0.0132 \times d^2 \times h + 0.217 \times d \times h$
	Thick branches	$if \begin{cases} d \leq 22.5 \text{ cm then } Z = 0, \\ d \geq 22.5 \text{ cm then } Z = 1' \end{cases}$ $W_{b7} = [0.107 \times (d - 22.5)^2] \times Z$
	Medium branches	$W_{b2-7} = 0.00792 \times d^2 \times h$
	Thin branches +needles	$W_{b2+n} = 0.273 \times d \times h$
	Roots	$W_r = 0.0767 \times d^2$
Pinus pinea	Stem	$W_s = 0.0224 \times d^{1.923} \times h^{1.0193}$
	Thick branches	$if \begin{cases} d \leq 22.5 \text{ cm then } Z = 0, \\ d \geq 22.5 \text{ cm then } Z = 1' \end{cases}$ $W_{b7} = [0.247 \times (d - 22.5)^2] \times Z$
	Medium branches	$W_{b2-7} = 0.0525 \times d^2$
	Thin branches +needles	$W_{b2+n} = 21.927 + 0.0707 \times d^2 - 2.827 \times h$
	Roots	$W_r = 0.117 \times d^2$
Quercus faginea	Stem	$W_s = 0.154 \times d^2$
	Thick branches	$W_{b7} = 0.0861 \times d^2$
	Medium branches	$W_{b2-7} = 0.127 \times d^2 - 0.00598 \times d^2 \times h$
	Thin branches + leaves	$W_{b2+1} = 0.0726 \times d^2 - 0.00275 \times d^2 \times h$
	Roots	$W_r = 0.169 \times d^2$
Quercus ilex	Stem	$W_s = 0.143 \times d^2$
	Thick branches	$if \begin{cases} d \leq 12.5 \text{ cm then } Z = 0, \\ d \geq 12.5 \text{ cm then } Z = 1' \end{cases}$ $W_{b7} = [0.0684 \times (d - 12.5)^2 \times h] \times Z$
	Medium branches	$W_{b2-7} = 0.0898 \times d^2$
	Thin branches + leaves	$W_{b2+1} = 0.0824 \times d^2$
	Roots	$W_r = 0.254 \times d^2$

Source: (Ruiz-Peinado Gertrudix *et al.*, 2012; Ruiz-Peinado *et al.*, 2011).

3.3. Tree species diversity estimation

The diversity indices used in this study were classified into 3 categories (Table 2). The basic idea of a diversity index is to obtain a quantitative estimate of biological variability that can be used to compare biological entities, composed of discrete components, in space or in time (Morris *et al.*, 2014).

Table 2. Categorization of indices

Richness & diversity	Species composition	Distribution pattern	
		Vertical	Horizontal
Simpson index (Sm)	Mingling (Mi)	Vertical Profile index (A)	Aggregation index (R)
Shannon index (Sn)	Spatial Diversity Status (MS)	Height Differentiation index (TH)	Uniform Angle Index (W)
Evenness index (E)	Segregation index (S)		
Berker-Parker index (D)			

3.3.1. Species richness indices

Two different aspects are generally used to conceptualize the diversity in a community: species richness and evenness. Species richness represents the number of species or attributes present in a community which is the simplest and most commonly applied metric. The distribution of individuals over species is called evenness. Additionally, species or trait abundance is also important for diversity, and the proportional abundance of species can be incorporated into indices that represent diversity.

3.3.1.1. Simpson Diversity index (1-D)

The Simpson diversity index (Eq. 1) was introduced by Edward H. Simpson (Simpson, 1949) to measure species diversity in a community by taking into account the number of species present and the abundance of each species. The index represents the probability that two individuals that are randomly selected from a sample will belong to different species.

$$1 - D = 1 - \frac{\sum_{i=1}^R n_i(n_i - 1)}{N(N - 1)} \quad \text{Eq. 1}$$

where n_i is the number of individuals belonging to i -th type, N is total number of individuals in the dataset, R – richness (total number of species types in dataset).

It ranges $0 \leq D \leq 1$. The value increases with species diversity. The higher the diversity, the greater the value of D .

3.3.1.2. Shannon index

Shannon index (H') by Shannon and Weaver (Shannon, 1948) is distance independent index to characterize the species diversity in a given stand. It takes into account both abundance and evenness of the existing species (Eq. 2).

$$H' = - \sum_{i=1}^S p_i \ln p_i \quad \text{Eq. 2}$$

where S is the number of species, p_i – proportion of i -th species in the total number of individuals of all species and calculated from individuals of i -th species divided by total number of individuals present (n/N), $\ln p_i$ is natural logarithm of this proportion. Its value

ranges from 0 to $\ln(S)$. When all species in the dataset are equally common, all p_i values equal $1/S$, and the Shannon index hence takes the value $\ln(S)$. When all abundance is concentrated in one species, and the other species are very rare (even if there are many of them), its value reduces to 0. The value is 0 when only one species in the dataset.

3.3.1.3. Berker-Parker index (*D*)

Berker-Parker index (Berger & Parker, 1970) is a measure of the numerical importance of the most abundant species in the population (Eq. 3). It has an analytical relationship with the geometric series of the species abundance model and represents the proportional abundance of only the most abundant species in the population (Morris *et al.*, 2014).

$$D = \frac{N_{max}}{N} \quad \text{Eq. 3}$$

where N_{max} is the number of individuals in the most abundant species and N is the total number of individuals in the sample. The reciprocal of the index, $1/D$, is often used, so that an increase in the value of the index corresponds an increase in diversity and a reduction in dominance.

3.3.1.4. Evenness index (*E*)

Species evenness (*E*) (Pielou, 1975) refers to how species are close to each other in numbers (del Río *et al.*, 2018). It represents the degree to which individuals disturb closely among species in terms of number. *E* is not calculated independently, but rather derived from compound diversity measures such as H' indices, as they inherently contain richness and evenness components. In Eq. 3, H' is the number derived from the Shannon diversity index and H'_{max} is the maximum possible value of H' (if every species was equally likely). *E* is supposed to be independent of a measure of species richness.

$$E = \frac{H'}{\ln S} \quad \text{Eq. 4}$$

where H' is a value of Shannon diversity index, $\ln S$ is natural logarithm of the number of species which equals to H'_{max} . Its value falls between 0 and 1 (1 demonstrates complete evenness). Low values indicate that one or a few species dominate, and high values indicate that relatively equal numbers of individuals belong to each species.

3.3.2. Species intermingling

The spatial relationships between two groups of individuals play important role for many components of a species' population biology. A numerous different types of tests indices have been designed to seek for an answer to the question whether two species are spatially segregated (individuals occur near the same species), associated (individuals occur near the other species), or neither.

3.3.2.1. Species segregation (*S*) index

Segregation index (*S*) developed by Pielou (1977) describes the degree of intermingling of two species groups based on nearest-neighbor method. *S* considers the ratio of the observed probability (p_{ij}) that reference tree i and its nearest-neighbor j belong to different species along with the same probability for completely randomly distributed or independent species attributes (del Río *et al.*, 2018) (Eq. 5). There are 2 main procedures to calculate *S* index: 1) calculation between distances between reference

trees i to every tree in the plot which derived from Euclidean distance calculation. Once the distances were computed, trees were ranked from nearest to farthest to reference tree and the first n -th number of neighboring trees (which are user dependent) were selected, 2) computation of S index: which was computed based on the nearest-neighbor tree distances calculated in 1st step. S is originally designed for being applied to a two-species mixture (Biber & Weyerhaeuser, 1998). In Pielou's approach, a contingency table is constructed in form described in Table 3.

$$S = 1 - \frac{p_{ij}}{E(p_{ij})} = 1 - \frac{\text{Observed number of mixed pairs}}{\text{Expected number of mixed pairs}} \quad \text{Eq. 5}$$

P_{ij} and $E(p_{ij})$ can be solved by Eq. 6:

$$S = 1 - \frac{N * (b + c)}{mw + nv} \quad \text{Eq. 6}$$

where: m and n are the observed number of individual trees of species 1 and 2 respectively. N can easily be extracted from sum of m and n as described in the table. The v and w are the number of individual trees of species 1 and 2 that are found as the nearest-neighbors of a reference tree. These variables are clearly described in a contingency table (Table 3).

Table 3. Descriptions of variables for S index calculation

		Nearest-neighbor species (j)		
		species.1	species.2	Total (i)
Reference species (i)	species.1	a	b	$m=a+b$
	species.2	c	d	$n=c+d$
	Total (j)	$v=a+c$	$w=b+d$	$N=m+n$

If the nearest-neighbors are always the same species as the reference trees, then $S=1$ which implies that the reference tree is associated with itself. There is a segregation of reference species from others. If all neighbors are different species, $S=-1$ which indicates that the reference tree is associated with other species. There is association between 2 species. Independent distribution of species is indicated by value near to 0.

$$S = \begin{cases} 1, & \text{all nearest neighbors (j) are the same species with reference tree (i)} \\ -1, & \text{all nearest neighbors are the different than reference tree (i)} \end{cases} \quad -1 \leq S \leq 1$$

3.3.2.2. Mingling (M_i) index

Species spatial mingling (M_i) index is a measure of species diversity within a structure unit (neighbor trees plus reference tree) which describes the proportion of neighbor trees which don't belong to same species as the reference tree. The M_i by Fuldner (1995) is defined as in Eq. 7 :

$$M_i = \frac{1}{n} \sum_i^n V_{ij} \quad \text{Eq. 7}$$

where n is the number of nearest neighbor trees considered, V_{ij} produces binary output which equals to 1 if the j -th neighbouring tree is not the same species as the i -th reference tree and $V_{ij} = 0$ otherwise.

$$V_{ij} = \begin{cases} 0, & \text{neighbour}(j) \text{ belong to the same species as reference tree } (i) \\ 1, & \text{neighbour}(j) \text{ belong to different species from reference tree } (i) \end{cases} \quad 0 \leq M_i \leq 1$$

A low degree of mingling indicates that trees of a particular species occur together with few or no trees of different species in the same area. High degree of mingling means that trees are surrounded by different species. Assume there are 4 neighbor trees to a reference tree, 5 different outputs are possible to derive as shown in Figure 3. The distribution of the M_i values, in conjunction with the species proportions within a given tree population, allows a detailed study of the spatial diversity within a forest. However, the number of different species in the structure unit was not taken into account, and this was a shortcoming of the M_i index. This shortcoming has fulfilled in spatial diversity status index.

3.3.2.3. Spatial diversity status (MS)

MS_i is improvement of mingling index. It considers not only the spatial mingling, but also the number of tree species. MS_i is determined by the relative species richness within the structure unit i and the degree of mingling of the reference tree and expressed by Eq. 8 (Gadow & Hui, 2002). The structural unit is defined by the neighborhoods that consisting a reference tree and its nearest neighbors (Zhang *et al.*, 2018).

$$MS_i = \frac{S_i}{n_{max}} \times M_i \quad \text{Eq. 8.}$$

Where S_i is the number of tree species in the neighborhood of the reference tree i , including tree i , and n_{max} is the maximum number of species in this structure unit. M_i is the species mingling value. MS measures the tree species richness as well as an important species characteristic within a structure unit. Reference tree of a common species is more likely to have the neighbors of the same species, reflecting low MS value. Rare species have less probability to have same neighbor species, resulting in high value of MS. Thus, MS is considered as an index that sensitive to rare species.

Determinations of S_i and n_{max} for MS are given in Figure 3 where explanation are based on example images of M_i . For 4 nearest neighbors, the structural unit is considered to be 5 (4 neighbors plus reference tree). There are 4 different species (a, b, c, d) in the structure unit. So n_{max} is defined by 4. S_i can be calculated as a number of species within the structure unit as exemplified in Figure 3.

Mi index					
	$M_i = 0/4 = 0$	$M_i = 1/4 = 0.25$	$M_i = 2/4 = 0.5$	$M_i = 3/4 = 0.75$	$M_i = 04/4 = 1$
	None of the neighbor tree is different species than reference tree	One of the neighbor tree is different species than reference tree	Two of the neighbor tree are different species than reference tree	Three of the neighbor trees are different species than reference tree	All 4 neighbor trees are different species than reference tree
	No mixture	Low mixture	Medium mixture	High mixture	Complete mixture
MS index	$S_i/n_{max} = 1/4$	$S_i = 2/4$	$S_i = 2/4$	$S_i = 4/4$	$S_i = 4/4$
	There is 1 species (a) in the structure unit/max number of species in this structure unit is 4.	There are 2 species (a, b) in the structure unit/ max number of species in this structure unit is 4.	There are 2 species (a, b) in the structure unit/ max number of species in this structure unit is 4.	There are 4 species (a, b, c, d) in the structure unit/ max number of species in this structure unit is 4.	There are 4 species (a, b, c, d) in the structure unit/ max number of species in this structure unit is 4.
Wi index					
	$W_i = 0/4 = 0$	$W_i = 1/4 = 0.25$	$W_i = 2/4 = 0.5$	$W_i = 3/4 = 0.75$	$W_i = 04/4 = 1$
	None of the angles (θ) is smaller than standard angle (θ_s)	One of the angles (θ) is smaller than standard angle (θ_s)	Two of the angles (θ) are smaller than standard angle (θ_s)	Three of the angles (θ) are smaller than standard angle (θ_s)	All of the angles (θ) is smaller than standard angle (θ_s)
	Very regular	Regular	Random	Clumped	Very clumped

Figure 3. Description and calculation of Mingling index (M_i), Spatial diversity status (MS) and Uniform Angle index (W_i) and corresponding likelihood values for structure unit of 4 neighbors around the reference i tree: a, b, c, d are the tree species types ; θ are angles between adjacent neighbor trees; θ_s is a standard angle (which is equal to $360/4$ for 4 neighbor trees) (Adapted from Gadow & Hui, 2002)

3.3.3. Spatial structural indices

The concept of spatial distribution consist vertical and horizontal spatial distributions which refer to the spatial arrangements (positioning) of different tree species along the

vertical or horizontal axis (del Río et al., 2018). The horizontal spatial distribution gives an idea of the variation of tree positioning (Bravo & Guerra, 2002). The indices that measure the horizontal spatial distribution quantifies the degree of regularity of the trees which are typically classified into regular, random, and clustered patterns and linked to processes of tree mortality, competitive interaction, regeneration and gap creation and so on. Vertical spatial distribution is most commonly described in terms of layers that refer to distinct classes or stratification of the canopy corresponding to height-related differentiation between trees.

3.3.3.1. Uniform Angle Index (W)

The Uniform Angle Index (W) formulates the degree of spatial dispersion of nearest neighbors around the reference tree based on angles between adjusting nearest neighbor trees defined as vectors from reference tree to each neighbors as shown in Figure 3 (Gadow et al., 1998). W is determined as the proportion of the angles that are smaller than the standard angle α_0 ($360/n$) and calculated as (Eq. 9):

$$W_i = \frac{1}{n} \sum_{j=1}^n v_j \quad \text{where } v_j = \begin{cases} 1, \alpha_j < \alpha_0 \\ 0, \text{otherwise} \end{cases} \quad \text{Eq. 9}$$

where n is number of nearest neighbours

$$0 \leq W \leq 1; \quad \text{If } \begin{cases} W < 0.5, \text{ regular tree distribution} \\ 0.5 < W < 0.6, \text{ random distribution} \\ W > 0.6, \text{ clumped} \end{cases}$$

The value of W ranges from 0 to 1. The value of W increases from regular to clumped pattern (regular < random < clumped).

3.3.3.2. Aggregation index R

Aggregation index (R) by Clark & Evans (1954) is a single value index that is designed to describe aspects of variability of tree locations in forest stands (Eq. 10).

$$R = \frac{\bar{r}_{observed}}{E(r)} \quad \text{where } E(r) = \sqrt{\frac{A}{N}} \quad \text{Eq. 10}$$

Where $\bar{r}_{observed}$ is an average distance to their nearest neighbours in a given forest stand while $E(r)$ is an average nearest neighbor distance when trees completely random distributed, A is area of the plot, N is the total number of trees in the plot. The edge effect arising from the spatial limitations of experiment plots has minimized by applying the boundary correction factor by Donnelly (1978). Interpretation of R values is as follows:

$$0 \leq R \leq 2.149: \quad R \begin{cases} < 1; \text{ then pattern shows clustering} \\ \approx 1; \text{ then pattern is random distribution} \\ > 1; \text{ then pattern is regular distribution} \end{cases}$$

3.3.3.3. Height Differentiation index (TH)

Height differentiation index (TH) is size differentiation index, developed by Gadow (1993) which measures the variability in height between i -th reference tree to each neighboring trees ($j-1 \dots n$) and describes vertical distribution of tree height (Eq. 11).

$$TH_{ij} = 1 - \frac{MIN(H_i, H_j)}{MAX(H_i, H_j)} \quad \text{Eq. 11}$$

where H_i , & H_j are the height of reference tree and neighbor tree respectively.

$$0 \leq TH \leq 1. \text{ If } \begin{cases} TH = 1, & \text{neighbour trees have high differentiation in height} \\ TH = 0, & \text{neighbour trees have equal height} \end{cases}$$

3.3.3.4. Vertical species profile (A)

Vertical species profile (A) (Pretzsch, 1995) is outlined in Eq. 12. Calculation is based on the Shannon and Weaver (Shannon, 1948) diversity index. A considers both proportion of the species within a stand and the presence of each species in different height zones (Eq. 12). Height zones were determined as a same way as Pretzsch (2009).

$$A = - \sum_{i=1}^S \sum_{j=1}^Z p_{ij} \times \ln(p_{ij}) \quad \text{Eq. 12}$$

where S represents the number of species present, Z is the number of height zones (three in this case), p_{ij} is the proportion of a species in the height zone $p_{ij} = \frac{n_{ij}}{N}$, N is the total number of individuals, n_{ij} is the number of individuals of the species i in the zone j . Standardization of A can be done by dividing A value by the maximum value of the A index, i.e. $A_{max} = \ln(S \times Z)$. Its value is greater than 0. For a pure stand with single layer, A equals to 0. Its value is increases as heterogeneous the vertical profile increases.

3.4. Statistical analysis

We used multiple linear regressions to evaluate the relationship between tree biomass and diversity indices in the stand, where total biomass (B) per tree was the dependent variable, and species richness, composition and species distribution indices were covariates of interest. Following previous researches that have developed a various regression models for estimating total-tree and tree compartment biomass, we utilized following three general forms of linear and non-linear regression equations (Eq. 13 to Eq. 15) for development of different forms of prediction models.

$$Y = \beta_0 + \beta_1 x_1 + \dots + \beta_k x_j + \varepsilon \quad \text{Eq. 13}$$

$$Y = \beta_0 x_1^{\beta_1} x_2^{\beta_2} \dots x_j^{\beta_k} + \varepsilon \quad \text{Eq. 14}$$

$$Y = \beta_0 \pi(x_i x_j)^{\beta_k} \quad \text{Eq. 15}$$

Two different approaches of regression analysis were used to find the coefficients: the first approach was to apply, whenever possible, multiple linear regressions to the original equations; the second approach was to transform the above equations to the logarithmical form and then apply multiple linear regressions to the transformed equations. Eq. 13 used to develop multiple linear regressions that can be fitted by standard least squares estimation. Non-linear models (Eq. 14 and Eq. 15) were transformed into linear models (Eq. 16 and Eq. 17) by taking the logarithm of both sides of the equation. In this form, the equation parameters can easily be estimated by least squares procedures.

$$\ln Y = \ln \beta_0 + \beta_1 \ln x_i + \dots + \beta_2 \ln x_j + \varepsilon \quad \text{Eq. 16}$$

$$\ln Y = \ln \beta_0 + \beta_1 \ln(x_i * x_j) + \varepsilon \quad \text{Eq. 17}$$

where \ln is the natural logarithm. ε is the random error term which is assumed to be normally distributed with mean zero and variance constant.

The models structures and their corresponding predictor variables are given in Table 4. 11 predictor variables: four species richness indices (Sm, Sn, D, E), three species composition indices (Mi, MS, S), four spatial distribution indices: A, TH, W and G were used as predictor variables for fitting models. Several different ways were implemented in variable selection process for the fitting models in order to avoid a problem of collinearity. First, all the variables were used as a single predictor variable for the models with single term. Second, basal area, species richness indices and species composition indices were utilized as state variables individually and each one of the spatial distribution indices were added into the multivariable models with two terms as secondary predictor variable. Third, we tested G, TH or A as a second, other individual indices as a third predictor variable for the multivariable models with three terms. Finally all possible combinations of indices are examined for multivariable models as well. In total, 537 alternative models were examined for each individual species and community level. For the community level analyze, basal area per quadrant G (m²/ha), for the species level analyze, species proportion of basal area G_p per quadrant were explored.

For the best models selection, four criteria were employed: i) significance of variables (in the ANOVA analysis, an effect is concerned to be significant when its coefficients have a probability less than or equal to the significant probability ($P < 0.05$), ii) biological meaning of parameters, iii) the normality of the residuals with Q-Q plots and iv) Akaike Information Criteria (AIC).

Table 4. Fitting models and their predictor variable reference

#	Fitted models	Predictor variable		
		x_1	x_2	x_3
Single predictor (33 alternative models)				
1.	$B = \beta_0 + \beta_1 x_1$	G, Sm, Sn, D, E, Mi, MS, S, A, TH, W		
2.	$\ln B = \beta_0 + \beta_1 \ln x_1$			
3.	$\ln B = \beta_0 + \beta_1 \ln x_1 + \beta_2 \ln(x_1^2)$			
Multivariate models with 2 predictors (360 alternative models)				
4.	$B = \beta_0 + \beta_1 x_1 + \beta_2 x_2$	G, G _p	Mi, MS, S, A, TH, W	
5.	$B = \beta_0 + \beta_1 x_1^2 + \beta_2 x_2^2$	Sm		
6.	$B = \beta_0 + \beta_1 x_1 + \beta_2 x_2 + \beta_3 x_3$	Sn		
7.	$B = \beta_0 + \beta_1 (x_1 * x_2)$	D		
8.	$\ln B = \beta_0 + \beta_1 \ln x_1 + \beta_2 \ln x_2$	E		
9.	$\ln B = \beta_0 + \beta_1 \ln(x_1^2) + \beta_2 \ln(x_2^2)$	Mi	A, TH, W	
10.	$\ln B = \beta_0 + \beta_1 \ln(x_1 * x_2)$	MS		
11.	$\ln B = \beta_0 + \beta_1 \ln(x_1^2 * x_2)$	S		
Multivariate models with 3 predictors (144 alternative models)				
12.	$B = \beta_0 + \beta_1 x_1 + \beta_2 x_2 + \beta_3 x_3$	G	TH	Sm, Sn, D, E, Mi, MS, S, A, W
13.	$\ln B = \beta_0 + \beta_1 \ln x_1 + \beta_2 \ln x_2 + \beta_3 \ln x_3$		A	
14.	$\ln B = \beta_0 + \beta_1 \ln(x_1 * x_2 * x_3)$	D, E, A, Sm, S, TH,	MS, S, TH, S, W,	

where B represents total biomass; x_1, x_2, x_3 represent the predictor variables; $\beta_0, \beta_1, \beta_2, \beta_3$ are the parameters of the models; For the species level analyze, basal area proportion of each species in each quadrant, for the whole community level analyze, basal area of each quadrant (m²/ha) were explored along with other indices.

4. - RESULTS

From the result of 537 investigated models, 11 significant models for community level are summarized in Table 5. Parameters of all models were statistically significant at ($p < 0.01$). TH alone or in combination with species richness or composition indices i.e Sm+TH, Sm*TH, Sn*TH, D+TH, E+TH, E*TH, Mi+TH, MS+TH, S+TH, G+TH+Sm, G+TH+MS was a main explanatory variable in all models except model 10 demonstrating the importance of vertical structural on tree biomass. According to parameter estimation, negative parameters were associated to species richness and compositional indices (Sm in model 1 and 4; MS in model 2 and 10; Mi in model 9; S in model 11 in logarithmic form) while positive parameters were correspond to TH in all the models excluding models 5 and 8 which implies that biomass increases as species richness or species composition decreases and height heterogeneity increases. From the selected 11 models, model 6 (Eq. 18) was found to be the best model for predicting community level tree biomass, showing negatively influence of D and positively influence of TH on tree biomass of the stand.

$$\ln B = 6.594 - 1.349 \ln D + 0.841 \ln TH \quad \text{Eq. 18}$$

Where B is total biomass per tree (kg), D is Berker-Parker index, TH is height differentiation index.

Table 6 showed that the parameter estimates of selected models by species based on models in Table 4. Similar trend that had observed at community level occur for *Pinus pinea*: TH alone or in combination with species richness or composition indices such as Sm+TH; Sm*TH; Sn+TH; Sn*TH; D+TH; D*TH; E*TH, Mi*TH was a main explanatory variable for all the models. The strength of the relationship ranges from 0.02 to 0.16 ($0.16 < R^2 < 0.2$). The highest statistically significant and the lowest AIC value belong to Eq. 19 (model 6, Table 6). The same form of model as community level analysis has been chosen as a best model for *Pinus pinea* as well. The reciprocal interactive effect of D and TH had the best prediction power on biomass. G_p has been shown to be the best explanatory variable for the biomass for both *Quercus faginea* (Eq. 20) and *Quercus ilex* (Eq. 21). None of the model was significant for *Juniperus thurifera* as concerned by models in this study.

$$\ln B = 7.249 - 0.935 \ln D + 0.988 \ln TH \quad \text{Eq. 19}$$

$$\ln B = 4.278 + 0.156 \ln G_p \quad \text{Eq. 20}$$

$$\ln B = 4.241 + 0.168 \ln G_p \quad \text{Eq. 21}$$

Table 5. Parameter estimates for biomass from selected models at community level

Model	Explanatory variable											N	Adj.R2	AIC	RSE	F.value
	Intercept	logG	logD	logTH	logSm	logMS	log(Sm*TH)	logE	logMi	logS	log(E*TH)					
1	2.929*** (0.638)	0.812*** (0.217)		0.763*** (0.098)	-11.142*** (2.953)							437	0.20	1264.2	1.02	37.809***
2	4.426*** (0.594)	0.706*** (0.213)		0.990*** (0.097)		-2.028*** (0.465)						437	0.21	1259.6	1.02	39.746***
3	5.684*** (0.137)			0.883*** (0.096)								437	0.16	1283.9	1.05	84.958***
4	5.191*** (0.207)			0.797*** (0.099)	-9.264*** (2.954)							437	0.18	1276.1	1.04	48.260***
5	5.717*** (0.142)					0.881*** (0.097)						437	0.16	1285.8	1.05	82.659***
6	6.594*** (0.218)		-1.349*** (0.257)	0.841*** (0.093)								437	0.21	1258.9	1.02	58.877***
7	6.149*** (0.185)			0.894*** (0.095)			32.906*** (8.965)					437	0.19	1272.5	1.03	50.434***
8	5.701*** (0.138)									0.887*** (0.096)		437	0.16	1283.2	1.05	85.791***
9	6.466*** (0.217)			1.038*** (0.100)					-1.319*** (0.288)			437	0.20	1265.2	1.02	54.926***
10	6.288*** (0.197)			1.001*** (0.098)		-1.973*** (0.470)						437	0.19	1268.5	1.03	52.898***
11	5.963*** (0.146)			0.710*** (0.101)						-2.125*** (0.452)		437	0.20	1264.2	1.02	55.575***

*** $p < 0.01$

Table 6. Parameter estimates for biomass from selected models at species level

Model	Explanatory variable														N	Adj.R2	AIC	RSE	F.value	
	Intercept	logTH	logSm	log(Sm*TH)	logSn	log(Sn*TH)	logD	log(D*TH)	log(E*TH)	log(Mi*TH+1)	logG	log(E+1)	log(E+1)^2	log(D+1)						log(D+1)^2
<i>Pinus pinea</i>																				
1	6.575*** (0.191)	0.937*** (0.163)														177	0.16	488.5	0.95	33.191***
2	6.158*** (0.249)	0.926*** (0.160)	-8.738** (3.391)													177	0.18	483.8	0.94	20.452***
3	6.596*** (0.199)			0.917*** (0.163)												177	0.15	489.9	0.96	31.494***
4	7.986*** (0.662)	0.923*** (0.161)			-1.230** (0.553)											177	0.17	485.5	0.94	19.444***
5	5.500*** (0.074)					0.774*** (0.162)										177	0.11	497.5	0.98	22.873***
6	7.249*** (0.289)	0.988*** (0.160)					-0.935*** (0.306)									177	0.19	481.2	0.93	22.060***
7	5.782*** (0.098)							0.532*** (0.145)								177	0.07	506.2	1.00	13.398***
8	6.594*** (0.193)								0.943*** (0.163)							177	0.16	488.1	0.95	33.613***
9	5.004*** (0.129)									2.864*** (0.556)					177	0.13	494.3	0.97	26.489***	
<i>Quercus faginea</i>																				
1	4.278*** (0.170)									0.156** (0.072)						157	0.023	183.2	0.43	4.632**
<i>Quercus ilex</i>																				
1	4.241*** (0.126)									0.168*** (0.045)						69	0.157	74.3	0.403	13.702***

p < 0.05; *p < 0.01

5. - DISCUSSION

Our finding revealed that the variation of tree biomass can be accounted by negative influence of D and positive influence of TH indicating that specific dominance in the stand influenced negatively and height heterogeneity influenced positively on biomass at community level. For the species level analysis, the variation of biomass of *Pinus pinea* can be explained by the same model as community level analysis which highlighted the negative impact of D and the positive impact of TH on biomass. This result is agreement with Bohn & Huth, (2017) who examined an influence of species diversity and forest structure on aboveground biomass over a broad range of forest stands and found out a positive relation between forest structural diversity and forest productivity. The same result was found in Riofrío *et al.*, (2017). Size heterogeneity enables bigger trees to obtain greater amount of a certain resource and use them more efficiently than small trees (Brockhoff *et al.*, 2017b). In our stand, mixture of the *Pine*, *Quercus faginea*, *Quercus ilex* and *Juniperus thurifera* might create a different canopy strata; top layer occupied by *Pinus pinea*, enabling the light-demanding species – *Pine* - to capture more light and grow better than other species and become dominant in the stand.

The variation of biomass for *Quercus faginea* and *Quercus ilex* were predicted by basal area proportion of species (Gp). The examined diversity indices such as Sm, Sn, D, E, Mi, MS, S, W, A, TH were found not significant relation with biomass of *Quercus*. This might be explained by abundance of *Pinus pinea*. The abundance of *Pine* have an strong inhibitory effect on the abundance and richness of understory species through light, water, and soil nutrients and thus reversely influences on biomass productivity of understory species (Laughlin & Grace, 2006). Moreover, biomass of individual tree in a given stand is not only the reflection of diversity (species richness, composition and structure) but also various internal and external factors such as age, stand density, site productivity, competition at the tree level, climate, soil (texture, moisture content), geographical location, and length of grown season (Con *et al.*, 2013; Poudel & Hailemariam, 2015) which might not be reflected by the available metrics or models that we are considered in this study.

One notable thing is that almost all the predictor variables of significant models (except model 1 in Table 6) for *Pinus pinea* were identical with the community level models in Table 5. This might be fact that community level analysis may be influenced by characteristic of *Pinus pinea* due to its dominance in the stand. This can be explained by “selection effect” which describes the impact of the most productive species on relationship between species richness and productivity. Positive relation between productivity (biomass accumulation) and tree diversity largely depends on presenting highly productivity species in multi-cultural communities (Tilman, 1999). However, in our case the proportion of the most productive, dominant species (*Pinus pinea*) showed a negative effect on biomass stand as represented by D in Eq. 18 & Eq. 19. As stress-tolerant and pioneer species, *Pinus* has the ability to become a dominant tree species in the mixed forest and accumulate biomass in a short time and it is considered as a strong competitor species with relatively high production due to its prolonged photosynthetic activity (coniferous evergreen tree) and high nutrient uptake through the rapid turnover of nutrients (Li, Su, Lang, Liu, & Ou, 2018). In old growth stand, *Pinus* with large diameter have a greater contribution to the stand biomass than small diameter trees (Baishya & Barik, 2011). In terms of complementarity of the species in this stand, *Pinus pinea* may facilitate development of *Quercus* by increasing seed protections, enhancing habitat condition which promote the recruitment of *Quercus* (Sheffer, 2012). Nevertheless, the facilitation of *Pine* for *Quercus* colonization depends on many factors such as stand density, development stage of *Quercus*, and environmental conditions. *Pine* forests with intermediate density enhance the site conditions for successful colonization of *Quercus* by reducing light intensity, creating partial shading and

improving the soil moisture status for *Quercus* seedlings. However, there is an opposition between suitable condition for recruitment and suitable condition for further growth and development of *Quercus* species that after seedlings and saplings stages (Puerta-Piñero, Gómez, & Valladares, 2007). In dense forest with poor environmental condition, although *Pinus* improved soil properties during a short period (a decade), it causes in decrement of soil moisture which may reduce recruitment of *Quercus* in their subcanopy. Low light interception levels and competition for water with *Pinus* reduces *Quercus* colonization in poor environmental condition. Maestre & Cortina, (2004) emphasized that *Pine* plantation in semi-arid environment don't facilitate the establishment of *Quercus* and causes in reduction of species richness and all plant cover. Therefore, based on these reviews, we can say that *Pinus pinea* might be a main contributor for biomass of the stand whereas its abundance has negatively impact on biomass of coexisting species i.e *Quercus faginea*, *Quercus ilex* and *Juniperus thurifera*.

6. - CONCLUSIONS

In this study, we examined the relationship between tree biomass and diversity in Llano de San Marugan, Valladolid, Spain at stand and individual species level by a various linear and non-linear models. After thoughtful examination of 537 models with 11 predictor tree diversity indices, we found out that there is a relation between tree biomass and diversity although this relation varies among the species. A stand level analyze revealed that tree biomass-diversity relation can be explained by an interaction of negative affect of abundance of dominant species and positive effect of tree' height heterogeneity which indicates that tree biomass increases as abundance of dominant species decreases and height heterogeneity increases. This result was identical for *Pinus pinea*. For *Quercus faginea* and *Quercus ilex*, only the species proportion of basal area (G_p) has a positive relation with biomass. The examined diversity indices were found not to have significant explanatory power for the explanation of biomass variation for *Quercus* species and *Juniperus thurifera*.

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ANNEX

Annex 1. Descriptive statistics of stand variables (437 trees in the one hectare plot)

Variables	Mean	St.Dev	Min	Pctl(25)	Pctl(75)	Max
B	137.4	196.7	2.9	26.2	128.4	1.067.5
d	16.6	11.7	3.0	8.6	23.1	50.5
h	5.8	2.5	2.0	4.0	7.6	13.0
V	0.1	0.2	0.001	0.01	0.1	1.2
Gi (m ² /ha)	14.3	3.1	8.8	12.5	16.8	19.4
Sm	1.0	0.02	0.9	1.0	1.0	1.0
Sn	3.3	0.4	2.4	3.2	3.5	3.9
E	1.0	0.01	1.0	1.0	1.0	1.0
D	2.1	0.4	1.2	1.9	2.3	2.6
Mi	0.6	0.3	0.0	0.5	0.8	1.0
MS	0.3	0.1	0.0	0.2	0.4	0.6
W	0.5	0.2	0.0	0.2	0.8	1.0
S	0.3	0.2	0.1	0.2	0.3	0.9
A	2.5	0.8	0.0	1.9	3.1	3.7
TH	0.3	0.1	0.0	0.2	0.4	0.7

B – total biomass (kg); d – breast height diameter (m); h – total height (m), V – volume (m³), Sm - simson index; Sn – shannon index; E – evenness index; G- basal area (m²/ha), D – barker Parker index, Mi – mingling index; MS – spatial diversity status; W – uniform angle index; S – segregation index; A – vertical profile index; TH – height differentiation index.

Annex 2. Tree attributes by species

#	Species	N	d (cm)	h (m)	v (m3)	g (m2)	B (kg)	Sm	Sn	D	E	Mi	MS	S	W	A	TH
1	<i>Pinus pinea</i>	177	27.512	8.163	0.256	0.070	290.110	0.955	3.215	1.985	0.987	0.549	0.251	0.194	0.489	1.900	0.366
2	<i>Quercus faginea</i>	157	9.604	4.183	0.019	0.008	38.757	0.963	3.390	2.141	0.987	0.562	0.260	0.239	0.497	2.965	0.263
3	<i>Quercus ilex</i>	69	9.100	4.122	0.013	0.007	27.122	0.969	3.548	2.130	0.987	0.558	0.270	0.612	0.446	3.137	0.226
4	<i>Juniperus thurifera</i>	34	7.626	4.331	0.271	0.005	21.764	0.956	3.202	2.189	0.981	0.765	0.312	0.241	0.507	2.683	0.251

Average values are presented for the tree species variables: B- above-ground biomass (kg); d –tree diameter at breast height (cm); h- total height (m); v – volume (m³); g – basal area per tree (m²), Mi - Mingling index; MS – Spatial diversity status; A – vertical profile Index; W – Uniform angle index; S – segregation index; TH – height differentiate index; G – basal area per quadrant (m²/ha); Sm – simpson index; Sn - shannon index; D- Berker- Parker index; E – evenness index; R – aggregation index, respectively.

Annex 3. Main attributes of tree by quadrant.

Quadrant number	Number of trees	Tree variables											Stand variables					
		<i>B</i>	<i>d</i>	<i>h</i>	<i>v</i>	<i>g</i>	<i>Mi</i>	<i>MS</i>	<i>A</i>	<i>W</i>	<i>S</i>	<i>TH</i>	<i>G</i>	<i>Sm</i>	<i>Sn</i>	<i>D</i>	<i>E</i>	<i>R</i>
1	26	110.4	14.1	5.4	0.18	0.027	0.55	0.25	2.29	0.47	0.21	0.29	11.06	0.96	3.18	2.60	0.98	1.02
2	31	127.5	16.0	5.7	0.14	0.030	0.65	0.27	2.56	0.52	0.21	0.29	14.73	0.97	3.40	1.94	0.99	1.27
3	21	152.0	16.5	6.0	0.19	0.034	0.73	0.35	2.25	0.48	0.22	0.36	11.36	0.95	2.97	2.33	0.97	1.00
4	34	132.8	16.4	5.6	0.18	0.032	0.71	0.34	2.57	0.54	0.22	0.30	16.81	0.97	3.41	2.29	0.98	0.92
5	12	272.9	25.8	7.5	0.38	0.065	0.50	0.18	1.59	0.52	0.17	0.41	12.55	0.91	2.47	1.33	0.99	0.99
6	29	130.3	16.5	5.6	0.25	0.031	0.59	0.26	2.62	0.45	0.18	0.33	14.27	0.96	3.33	1.93	0.99	1.15
7	28	140.6	17.5	5.9	0.11	0.033	0.65	0.29	2.41	0.54	0.24	0.33	14.80	0.96	3.29	2.15	0.99	1.00
8	53	83.4	13.1	5.0	0.07	0.020	0.53	0.24	3.35	0.53	0.39	0.21	16.63	0.98	3.94	2.21	0.99	1.06
9	26	93.5	12.9	5.0	0.17	0.021	0.64	0.29	2.52	0.49	0.21	0.28	8.76	0.96	3.20	2.17	0.98	1.12
10	13	353.7	28.6	8.0	0.21	0.081	0.31	0.12	1.56	0.38	0.18	0.36	16.89	0.92	2.54	1.18	0.99	1.08
11	23	173.9	17.8	6.0	0.04	0.038	0.60	0.30	2.42	0.38	0.29	0.33	14.14	0.95	3.08	2.30	0.98	0.89
12	36	66.7	12.6	5.0	0.09	0.016	0.39	0.18	2.81	0.53	0.53	0.22	9.46	0.97	3.52	1.89	0.98	0.97
13	11	365.9	29.9	8.3	0.36	0.085	0.55	0.22	1.54	0.45	0.16	0.37	14.93	0.91	2.38	1.38	0.99	1.21
14	25	175.8	20.8	7.1	0.18	0.044	0.41	0.16	2.13	0.49	0.29	0.31	17.48	0.96	3.18	1.47	0.99	1.03
15	31	105.5	14.8	5.3	0.01	0.025	0.60	0.29	2.71	0.36	0.34	0.33	12.46	0.97	3.40	2.58	0.99	1.02
16	38	131.3	17.1	6.2	0.02	0.032	0.62	0.32	2.61	0.51	0.28	0.30	19.44	0.97	3.60	2.11	0.99	1.22

Average values are presented for the tree stand variables: B- above-ground biomass (kg); d – tree diameter at breast height (cm); h- total height (m); v – volume (m³); g – Basal area per tree (m²), Mi - mingling index; MS – spatial diversity status; A – vertical profile Index; W – uniform angle index; S – segregation index; TH – height differentiate index; G – basal area per quadrant (m²/ha); Sm – simpson index; Sn - shannon index; D- Berker-Parker index; E – evenness index; R – aggregation index, respectively.

Annex 4. R scripts

```
1 ##### Relationship between biomass and biodiversity in mixed forest #####
2
3
4 # The main question to address in this work is "Is the stand with more biodiversity holds
5 # more biomass? so main objective of this work is to explore the relationship between
6 # biodiversity and biomass in mixed forest.
7
8 # In order to achieve this objective:
9 # > 1. we computed biomass at different compartment and other indices that characterise a vertical and horizontal
10 # diversity of a stand at different levels (tree, stand, species) and
11 # > 2. we used different types of statistical analysis to find out the relationship between biomass and
12 # the indices
13
14 # CONTENTS of the script:
15
16 # BIOMASS at different compartments
17 # > stem
18 # > thick branch
19 # > medium branch
20 # > thin branch
21 # DIVERSITY INDICES (stand level indices)
22 # > Simpson index (D)
23 # > Shannon index (H)
24 # > Evenness index (E)
25 # > Berger-Parker index (D)
26 # SPATIAL POINT PATTERN ANALYZES (stand level indices)
27 # > L function (L)
28 # > - at whole stand level
29 # > - at quadrat level
30 # > - at species level (within and between interactions of species) within each quadrat
31 # > Aggregation index (R)
32 # > - at whole stand level
33 # > - at quadrat level
34 # > - at species level (within and between interactions of species) within each quadrat
35
```

```
29 # > - at quadrat level
30 # > - at species level (within and between interactions of species) within each quadrat
31 # > Aggregation index (R)
32 # > - at whole stand level
33 # > - at quadrat level
34 # > - at species level (within and between interactions of species) within each quadrat
35 # SPECIES INTERMINGLING INDICES (tree level indices)
36 # > Mingling index (MI)
37 # > spatial diversity status (MS)
38 # > uniform angle index (w)
39 # > Segregation index (s)
40 # VERTICAL SPATIAL PATTERN INDICES (stand level indices)
41 # > Vertical species profile (A)
42 # > height differentiation index (TH)
43 # CORRELATION ANALYSIS BETWEEN BIOMASS AND DIVERSITY INDICES
44
45 # Set working directory
46 setwd("D:/MEDFOR_classes/MScThesis_MEDFOR_UVA/data/processed")
47
48 # open the data to be used
49 data <- read.csv("data_martelloscopes_Marugan.csv", header = TRUE)
50
51 ##### BIOMASS CALCULATION #####
52 # https://www.r-bloggers.com/a-wrapper-around-nested-ifelse/
53 ##### Use ifelse to calculate biomass for each compartment of tree
54
55 # define variables that are used for calculation
56
57 d <- data$DBH_cm
58 h <- data$total_height_m
59 species <- data$species
60
61 ## biomass weight (kg) for stem (G) according to different diameter of different species
62
63
```

```

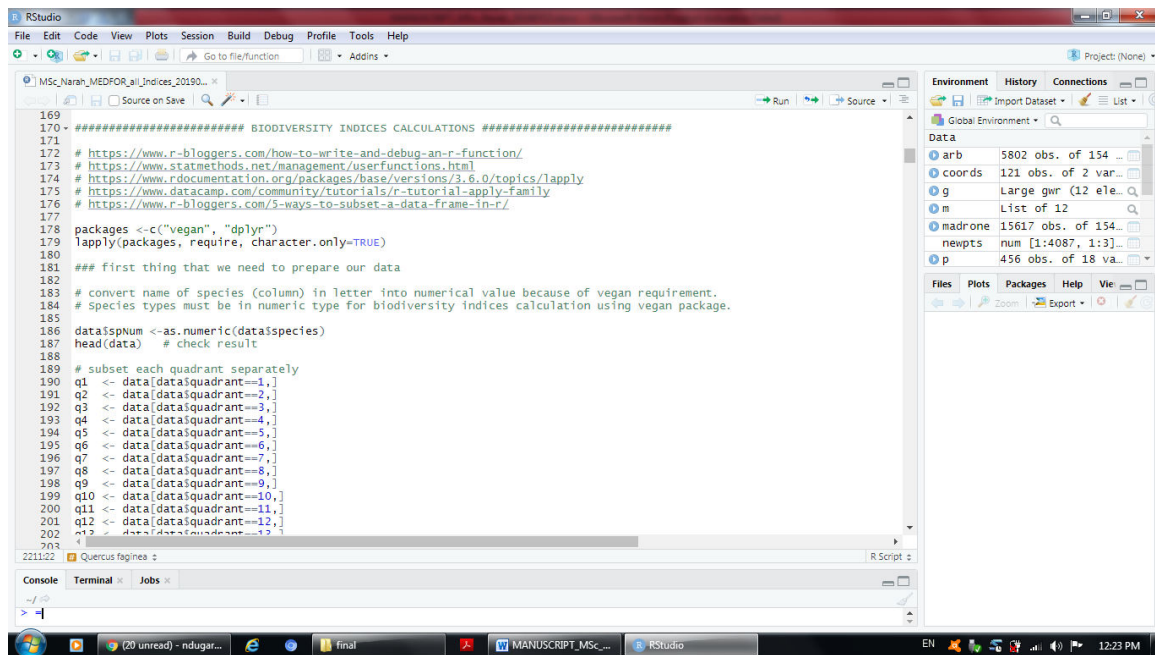
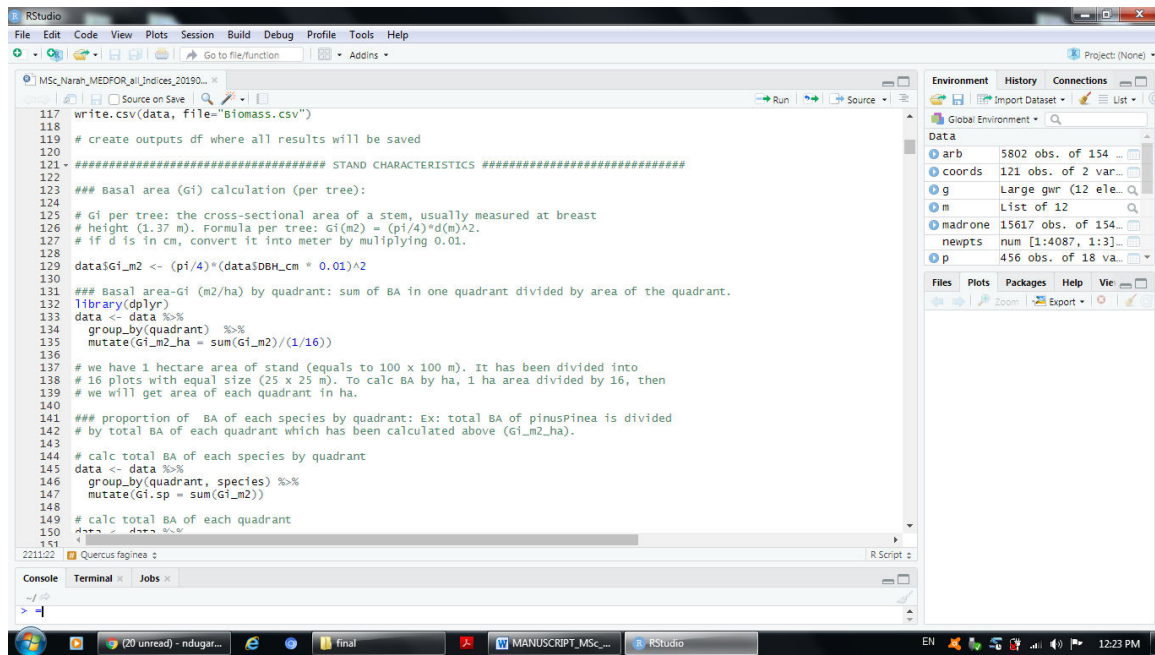
RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Narar_MEDFOR_all_Indices_20190...
49 data <- read.csv("data_martel\loscoptes_Marugan.csv", header = TRUE)
50
51 ##### BIOMASS CALCULATION #####
52
53 # https://www.r-bloggers.com/a-wrapper-around-nested-ifelse/
54
55 ##### Use ifelse to calculate biomass for each compartment of tree
56
57 # define variables that are used for calculation
58 d <- data$DBH_cm
59 h <- data$total_height_m
60 species <- data$species
61
62 ### Biomass weight (kg) for stem (bs) according to different diameter of different species:
63
64 data$bs_kg <- ifelse(species == "Juniperus thurifera", 0.0132*d^2*h + 0.217*d*h,
65 ifelse(species=="Pinus pinea", 0.0224 * d^1.923 * h^1.0193,
66 ifelse(species=="Quercus faginea", 0.154 * d^2,
67 ifelse(species=="Quercus ilex", 0.143 * d^2,
68 ifelse(species=="Pinus nigra", 0.0403 * d^1.838 * h^0.945,
69 ifelse(species=="Pinus sylvestris", 0.0154 * d^2 * h, 0))))))
70
71 ### Biomass weight for thin branch (b2) according to different diameter of different species:
72
73 data$b2_kg <- ifelse(species == "Juniperus thurifera", 0.273*d*h,
74 ifelse(species == "Pinus pinea", 21.927 + 0.0707*d^2 - 2.827*h,
75 ifelse(species == "Quercus faginea", 0.0726*d^2 - 0.00275*d^2*h,
76 ifelse(species == "Quercus ilex", 0.0824 * d^2,
77 ifelse(species == "Pinus nigra", 0.0720 *d^2,
78 ifelse(species=="Pinus sylvestris", 0.530*d^2.199*h^(-1.153),
79 ifelse(species == "Quercus pyreniaca", 0.898*d-0.445*h, 0))))))
80
81 ### Biomass weight for medium branch (b2_7) according to different diameter of different species:
82
83
2211:22 Quercus faginea
R Script

```

```

RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Narar_MEDFOR_all_Indices_20190...
81 ### Biomass weight for medium branch (b2_7) according to different diameter of different species:
82
83 data$b2_7_kg <- ifelse(species == "Juniperus thurifera", 0.00792*d^2*h,
84 ifelse(species == "Pinus pinea", 0.0525*d^2,
85 ifelse(species=="Quercus faginea", 0.127*d^2 - 0.00598*d^2*h,
86 ifelse(species=="Quercus ilex", 0.0898*d^2,
87 ifelse(species=="Pinus nigra", 0.0521*d^2,
88 ifelse(species=="Pinus sylvestris", 0.0295*d^2.742 * h^(-0.899),
89 ifelse(species=="Quercus pyreniaca", (-0.0260*d^2 + 0.536*h + 0.
90
91
92
93 ### Biomass weight for thick branch (b7) according to different diameter of different species:
94
95 data$b7_kg <- ifelse(species=="Juniperus thurifera" & d>22.5, (0.107*(d-22.5)^2),
96 ifelse(species=="Quercus ilex" & d>12.5, (0.0684*(d-12.5)^2*h),
97 ifelse(species=="Pinus pinea" & d>22.5, (0.247*(d-22.5)^2),
98 ifelse(species=="Quercus faginea", 0.0861*d^2,
99 ifelse(species=="Pinus nigra" & d>32.5, (0.228*(d-32.5)^2),
100 ifelse(species=="Pinus sylvestris" & d>37.5, (0.540*(d-37.5)^2-0.0119*(d-
101
102
103 ### Biomass weight for roots according to different diameter of different species:
104
105 data$brt_kg <- ifelse(species == "Juniperus thurifera", 0.0767*d^2,
106 ifelse(species == "Pinus pinea", 0.117*d^2,
107 ifelse(species=="Quercus faginea", 0.169*d^2,
108 ifelse(species == "Quercus ilex", 0.254*d^2,0)))
109
110 ### Biomass weight for stem and thick branch of quercus pyreniaca according to different diameter of different species:
111
112 data$bs_b7_q_p_kg <- ifelse(species=="Quercus pyreniaca", 0.0261*d^2*h, 0)
113 # !! it is 0 in output because there is no Quercus pyreniaca in Marugan plot
114
115 ### Total Biomass weight calculation for a whole tree (per tree):
116
2211:22 Quercus faginea
R Script

```



```

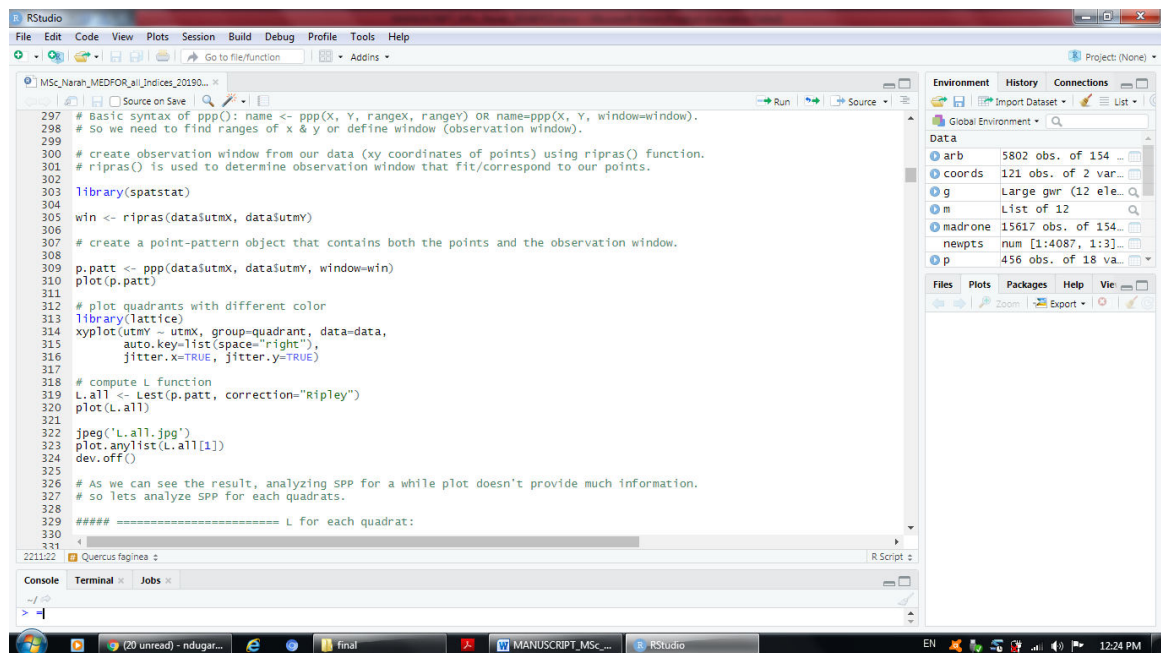
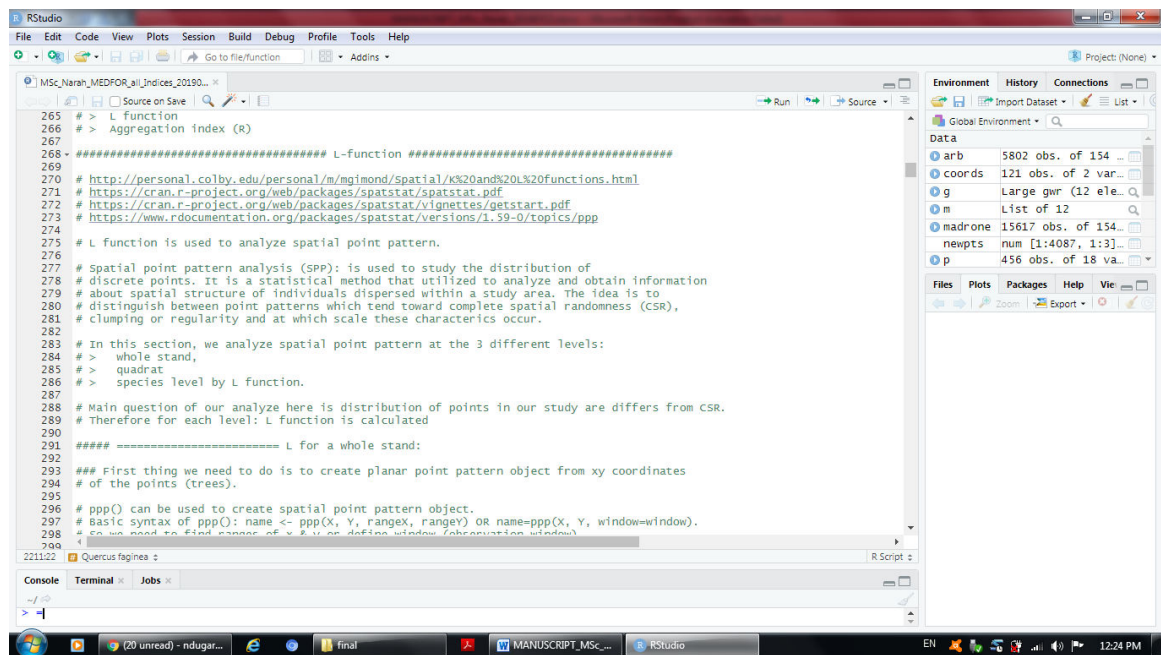
RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Narah_MEDFOR_all_indices_20190...
209 ## define functions that are going to be applied for each quadrant:
210
211 # functions of some indices have already been developed in vegan package. some are not.
212 # So lets write function for those that don't have pre-defined function.
213
214 # function for Evenness index:
215 E <- function(x){
216   H <- diversity(x) # total number of individuals in dataset (shannon = H)
217   S <- specnumber(x) # rowsums does the same...
218   E <- H/Log(S) # Evenness index
219   return(E)
220 }
221
222 # function for berger-parker index:
223 D <- function(x){
224   N <- length(x) # total number of individuals in dataset
225   n <- table(x) # number of individuals of each species
226   Nmax <- max(n) # max number of the most abundant species
227   D <- Nmax/N # calculate berker parker index
228   recip.D <- 1/D
229   return( recip.D)
230 }
231
232
233 # check functions for quadrant 1 whether they are correct or not.
234 for(i in 1:nrow(q1)){
235   q1simp <- diversity(q1$spnum, "simpson")
236   q1shan <- diversity(q1$spnum, "shannon")
237   q1E <- E(q1$spnum)
238   q1D <- D(q1$spnum)
239 }
240 head(q1)
241
242 ## apply all the functions for all quadrats simultaneously:
243
2211:22 Quercus faginea
R Script 2
Console Terminal Jobs
> |

```

```

RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Narah_MEDFOR_all_indices_20190...
237 q1E <- E(q1$spnum)
238 q1D <- D(q1$spnum)
239 }
240 head(q1)
241
242 ## apply all the functions for all quadrats simultaneously:
243
244 outDiver <- lapply(data.list, function(x){
245   for(i in 1:nrow(x)){
246     x1simp <- diversity(x1$spnum, "simpson")
247     x1shannon <- diversity(x1$spnum, "shannon")
248     x1Even <- E(x1$spnum)
249     x1D <- D(x1$spnum)
250   }
251   x
252 }
253
254 # reconstruct initial dataframe back
255 outputDiver <- bind_rows(outDiver)
256
257 # combine result into original data
258 data <- merge(data, outputDiver, all.x = FALSE)
259
260 write.csv(outputDiver, file="diverInd.csv")
261
262 ##### SPATIAL POINT PATTERN ANALYZE #####
263
264 # we will analyze spatial point pattern by:
265 # > L function
266 # > Aggregation index (R)
267
268 ##### L-function #####
269
270 # https://personal.colby.edu/personal/e/mcismond/teaching/8%20and%20%20function.html
271
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R Script 2
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> |

```



```

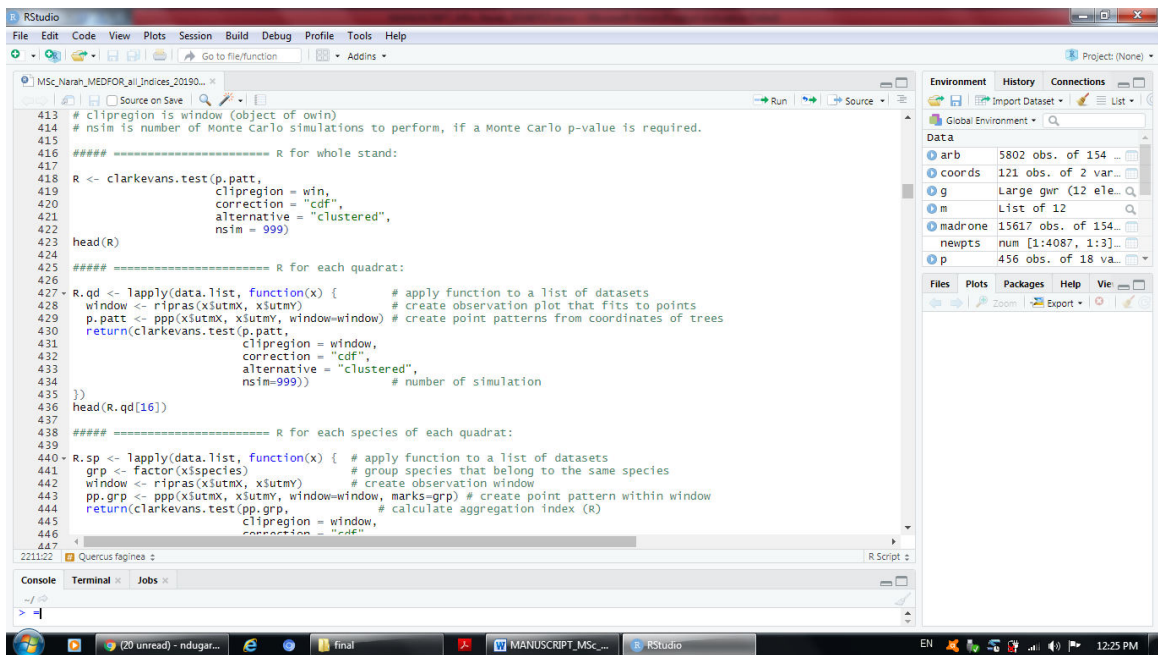
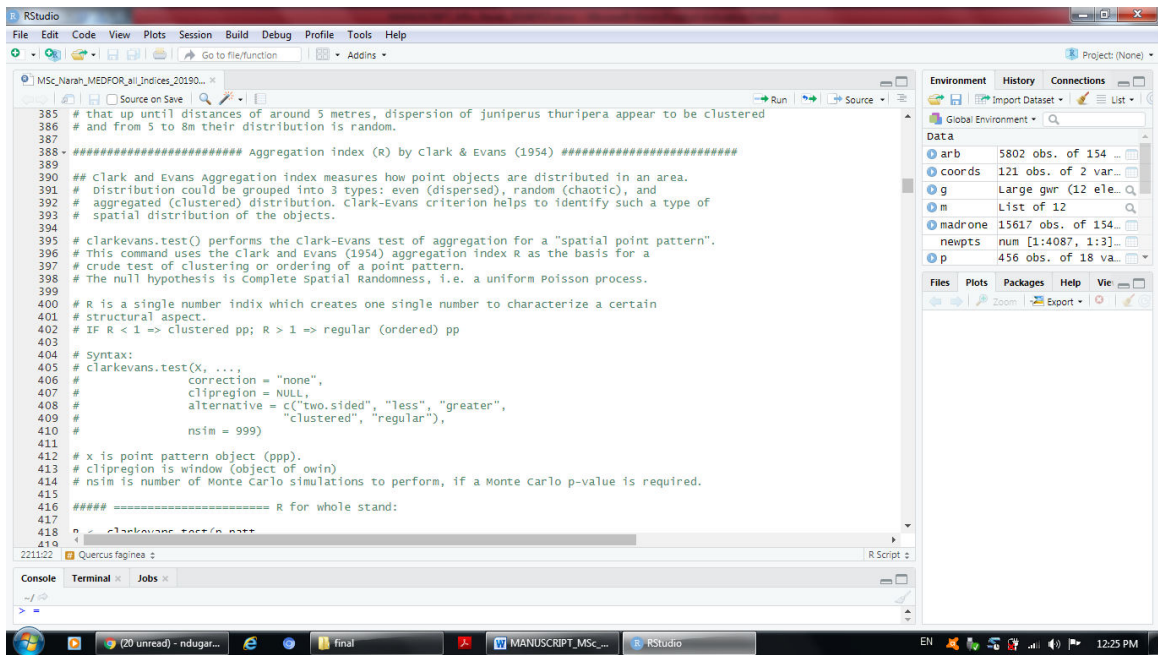
RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Naran_MEDFOR_all_indices_20190...
329 ##### L for each quadrat:
330
331 # apply L.fun() for a list of datasets:
332 L.qd <- lapply(data.list, function(x) { # apply function to a list of datasets
333   window <- rripas(x$utmX, x$utmY) # create observation plot that fits to points
334   p.patt <- ppp(x$utmX, x$utmY, window=window) # create point patterns from coordinates of trees
335   return(Lest(p.patt, correction="Ripley")) # calculate L function
336 })
337
338
339 jpeg("L.qd.jpg")
340 plot.anylist(L.qd[1:2], main="L for quadrant 1 and 2")
341 dev.off()
342
343 # run cut(dev.off()) if plot doesn't appear and then run plot() function again.
344
345 # there is a warning message: data contain duplicated points. we can find duplicated points
346 # remove them. Either remove them or not are the same. They don't effect on analyze!
347 # duplicated(L.qd)
348
349 ##### L for each species type in each quadrat:
350
351 # In this section, we analyze within & between interactions of different species in each quadrats.
352
353 # Main processes for this analyze:
354 # 1. group species by their type with factor() function.
355 # 2. create observation window
356 # 3. create point pattern object within observation window
357 # 4. split each species type
358 # 5. compute L function
359 # 6. check CSR by randomisation test using envelope() function.
360
361 # I combined the processes from 1 to 5 are in a single function and applied this function to
362 # the list of datasets (quadrats) as below
363
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R Script 2

```

```

RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Naran_MEDFOR_all_indices_20190...
357 # 4. split each species type
358 # 5. compute L function
359 # 6. check CSR by randomisation test using envelope() function.
360
361 # I combined the processes from 1 to 5 are in a single function and applied this function to
362 # the list of datasets (quadrats) as below.
363
364 L.sp <- lapply(data.list, function(x) { # apply function to a list of datasets
365   grp <- factor(x$species) # group species that belong to the same species together
366   window <- rripas(x$utmX, x$utmY) # create observation window
367   pp.grp <- ppp(x$utmX, x$utmY, window=window, marks=grp) # create point pattern within window
368   split.grp <- split(pp.grp) # split species by their group
369   return(alltypes(pp.grp, "L")) # compute L function for all types of species
370 })
371
372 jpeg("L.sp.jpg")
373 plot.anylist(L.sp[1])
374 dev.off()
375
376 ## Explanation:
377 # the diagonal plots show the result of L function to examine within and between species interaction
378 # in quadrant 1. These indicate that there are some variation of clustering (repulsion) at
379 # different distances.
380 # the kpois(r) line in blue is the theoretical value of L for each distance window (r) under
381 # a Poisson assumption of complete Spatial Randomness (CSR). The other line is the estimated
382 # values of L accounting...
383 # wherever the value of L falls above the line, species appear to be clustered at that distance.
384 # wherever the value of L is below the line, the data are dispersed. From the graph, we can see
385 # that up until distances of around 5 metres, dispersion of Juniperus thurifera appear to be clustered
386 # and from 5 to 8m their distribution is random.
387
388 ##### Aggregation index (R) by Clark & Evans (1954) #####
389
390 # Clark and Evans Aggregation index measure how point objects are distributed in an area
391
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R Script 2

```



```

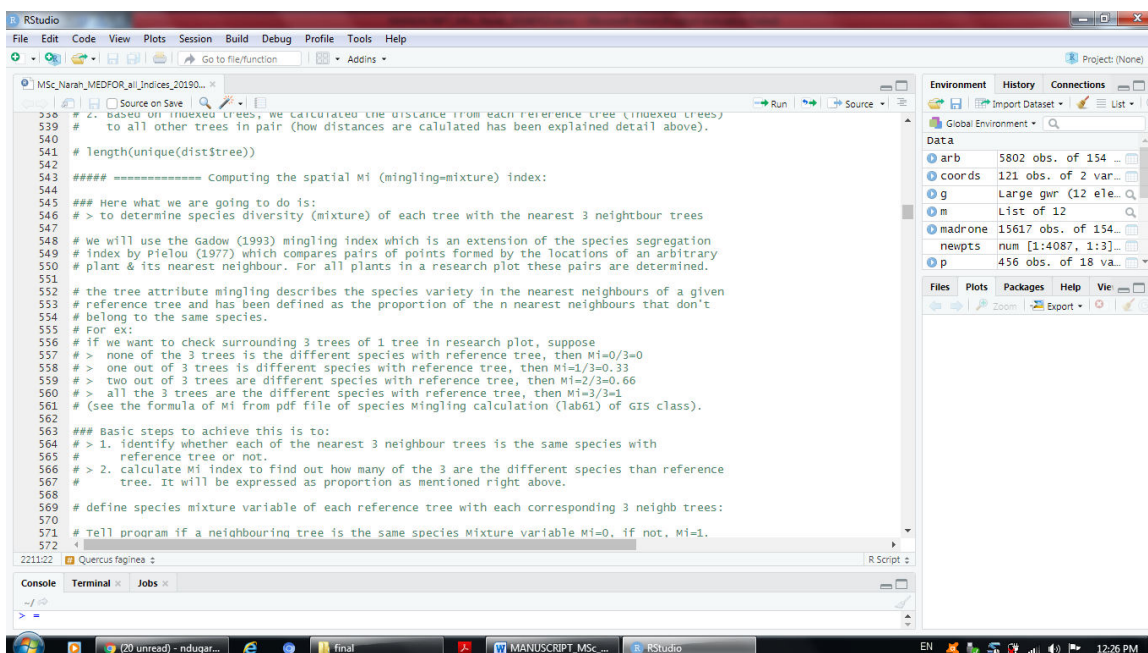
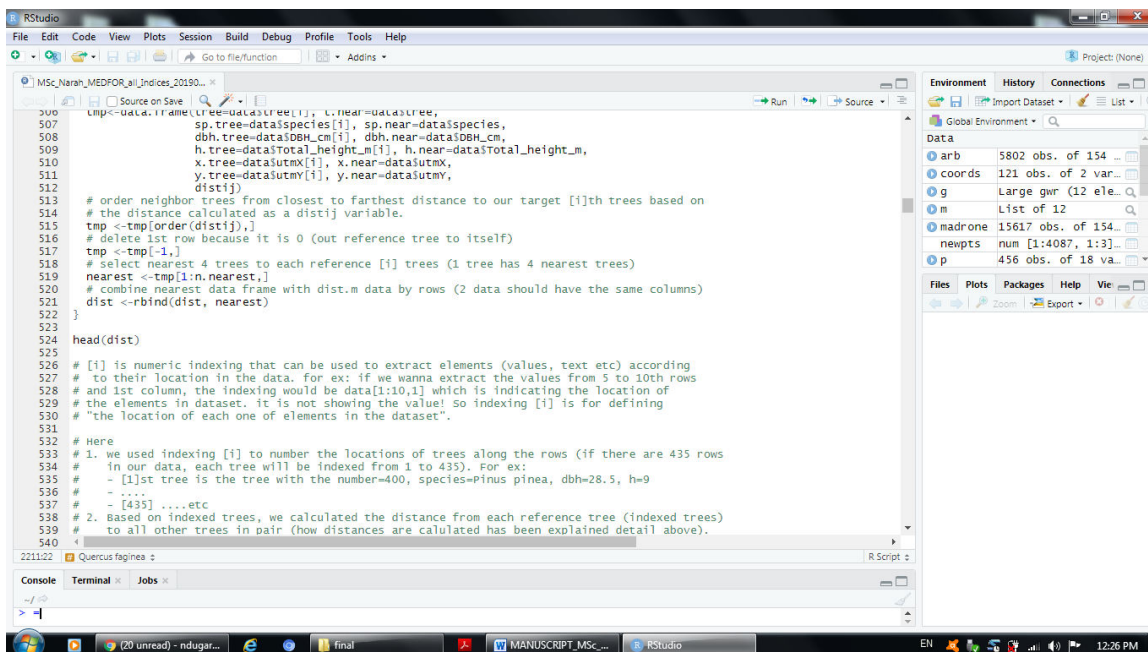
RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Narah_MEDFOR_all_indices_20190...
451 head(R.sp[1:4])
452
453 ##### SPECIES INTERMINGLING INDICES (indices at tree level) #####
454
455 ##### Species mingling (mixture) index (Mi) #####
456
457 # > calculate species mingling-mixture (Mi) index:
458 # a. calculate the distance (Euclidean distance) between trees first and then
459 # b. calculate mingling index (Mi)
460
461 ### a) calculating the distances from reference tree to each the nearest neighbor trees:
462 # > obtain the Euclidean distance between trees using the xy coordinates of every trees
463 # using basic Euclidean distance equation and then
464 # > select the nearest [k] trees (n.nearest) for a reference tree based on distance
465 # that determined by Euclidean distance equation.
466
467 ### For calculation of Euclidean distance:
468 # with the Euclidean distance equation, we will calculate distances of every trees from each
469 # reference tree. Every trees are the reference tree which means if there are 10 trees in plot
470 # with the indices [i] of 1st to 10th, distances from 1st tree to other 9 trees are calculated
471 # in pairs. First each tree will be numbered according to their locations in row. And then
472 # distances will be calculated from each tree to all other trees. For ex:
473 # - from [1]st tree to other 9 trees: [1]st to [2]nd, [3]rd...[10]th. [1]st tree is ref tree here
474 # - from [2] tree to other 9 trees: [2] to [1],[2],[3]...[10]. [2]nd tree is ref tree here
475 # - ...
476 # - from [10] to other 9 trees: [10] to [1],[2],[3]...[10].
477 # Totally 100 (10 * 10 = 100) distances are calculated.
478
479 ### For selecting the nearest [k] trees (n.nearest) for a reference tree
480 # nearest neighbor trees of each tree will be selected based on distances from closest to
481 # farthest after distances are ordered, if we choose 3 nearest neighbour trees to each
482 # reference trees, then 30 distance (10 trees * 3 neighbours = 30) will be selected.
483
484
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R Script 2
Console Terminal Jobs
> |

```

```

RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Narah_MEDFOR_all_indices_20190...
482 # reference trees, then 30 distance (10 trees * 3 neighbours = 30) will be selected.
483
484 #! Some other indices in this script will be calculated based on this distance calculation.
485
486 ### Basic steps are:
487 # > 1. Define the number of nearest trees (n.nearest)
488 # > 2. Use a loop to
489 # - calculate the distance between all the trees in the dataset (by Euclidean dist formula)
490 # - order the results from closest to farthest one
491 # - keep only the nearest [j]=3, 4 or 5 neighbour trees (n.nearest)
492
493 ##### calculate the distances to each neighbour:
494
495 # create empty file where we'll save our results
496 dist <- data.frame()
497
498 # defining the number of neighbour trees which is 3 in this case
499 n.nearest <- 4
500
501 # create a loop
502 for(i in 1:nrow(data)){
503 # obtaining the distance from tree i (reference or target tree) to every tree by eucl eqn.
504 dist[i] = ((data$utmX[i]-data$utmX)^2 + (data$utmY[i]-data$utmY)^2)^(1/2)
505 # now we identify the tree [i] and its neighbours and create temporary file
506 tmp<-data.frame(tree=data$tree[i], t.near=data$tree,
507 sp.tree=data$species[i], sp.near=data$species,
508 dbh.tree=data$DBH_cm[i], dbh.near=data$DBH_cm,
509 h.tree=data$total_height_m[i], h.near=data$total_height_m,
510 x.tree=data$utmX[i], x.near=data$utmX,
511 y.tree=data$utmY[i], y.near=data$utmY,
512 dist[i])
513 # order neighbor trees from closest to farthest distance to our target [i]th trees based on
514 # the distance calculated as a dist[i] variable.
515 tmp <-tmp[order(dist[i],)]
516
2211:22 Quercus faginea
R Script 2
Console Terminal Jobs
> |

```

```

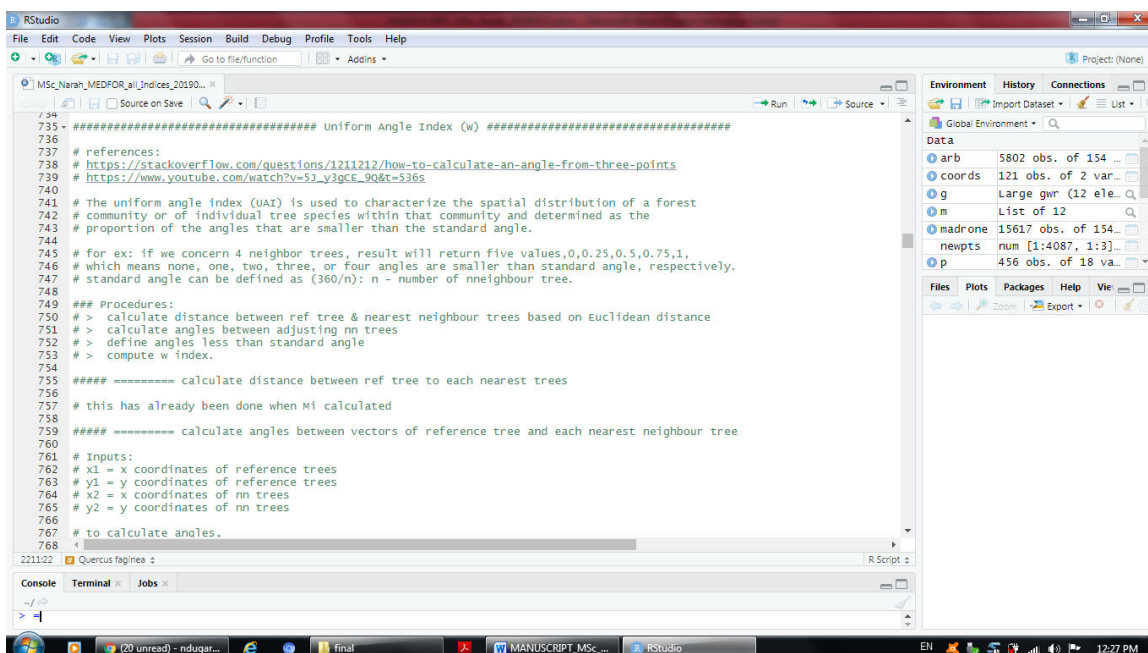
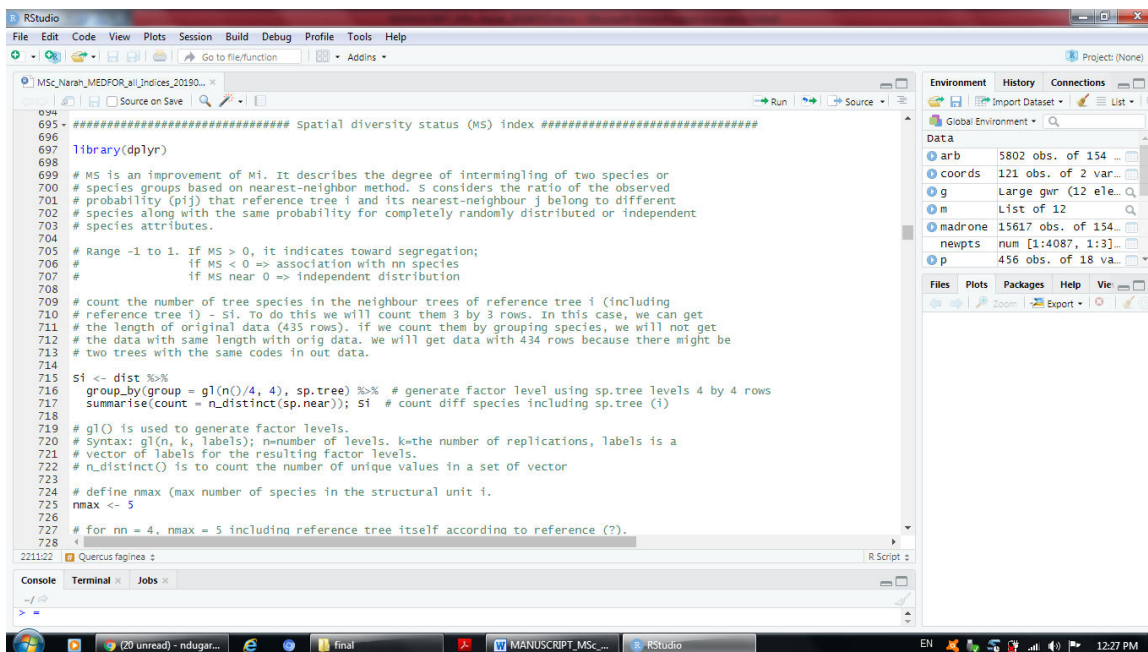
RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Nararh_MEDFOR_all_indices_20190...
370
371 # Tell program if a neighbouring tree is the same species Mixture variable M1=0, if not, M1=1.
372 dist$M1 <- ifelse((dist$sp.tree == dist$sp.near), 0, 1)
373
374 # save data as a file in the working directory.
375 # write.csv(dist, file="D:/MEDFOR_classes/MScThesis_MEDFOR_Uva/outputs/dist.csv")
376
377 # now we will calculate the M1. Here we can calc it by mean of M1ij for 3 nearest neighbors
378 # a reference tree. To do this: we
379 # > selected reference tree variables from dist data
380 # > grouped the data by 3 by 3 rows
381 # > calc-d mean of M1ij within each group
382 # > select i reference tree info from other rows
383
384 outputM1 <- dist %>%
385 select(tree=sp.tree, dbh.tree, h.tree, x.tree, y.tree, distij, M1ij) %>%
386 group_by(nn = gl(n()/4, 4), sp.tree) %>%
387 mutate(M1 = mean(M1ij), dbh.tree = dbh.tree[1], h.tree = h.tree[1],
388        x.tree=x.tree[1], y.tree = y.tree[1], distij = distij[1]) %>%
389 slice(1)
390
391 # gl(): generate factors by specifying the pattern of their levels.
392 # it has been used to generate new factors (grouping) variable.
393 # n()/3: define number of levels. original dist data has 1305 rows. Every 3 rows leveled
394 # as a same numbers. So the number of level is 435 starting from 1.
395 # 3: number of replication. A level is replicated by 3 for every 3 rows.
396 # sp.tree: define a vector of labels for the resulting factor levels.
397
398 # slice(1): slice data by 1st row of each group
399
400 # to save all reference tree variables from dist data into outputM1 data, we should define
401 # them not only inside of select() but also inside of mutate() by telling R to choose only
402 # the first value ([1]) to be the value for that group. otherwise selected variables
403 # were not in output data. [1] is to select only 1 value for that group.
404
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R Script 2

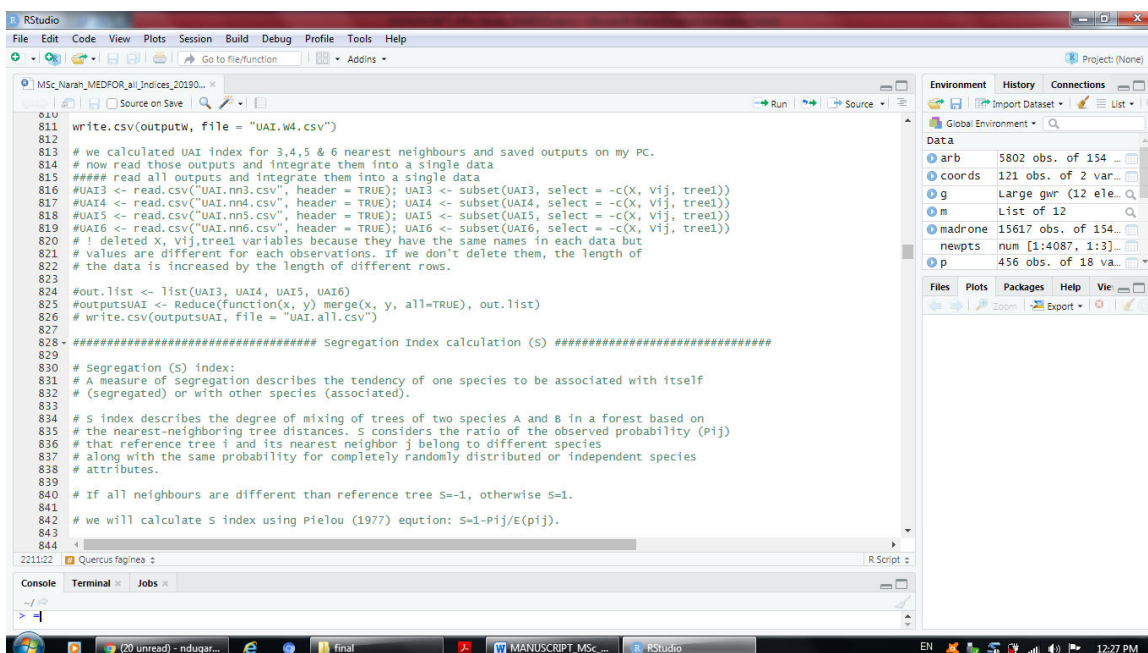
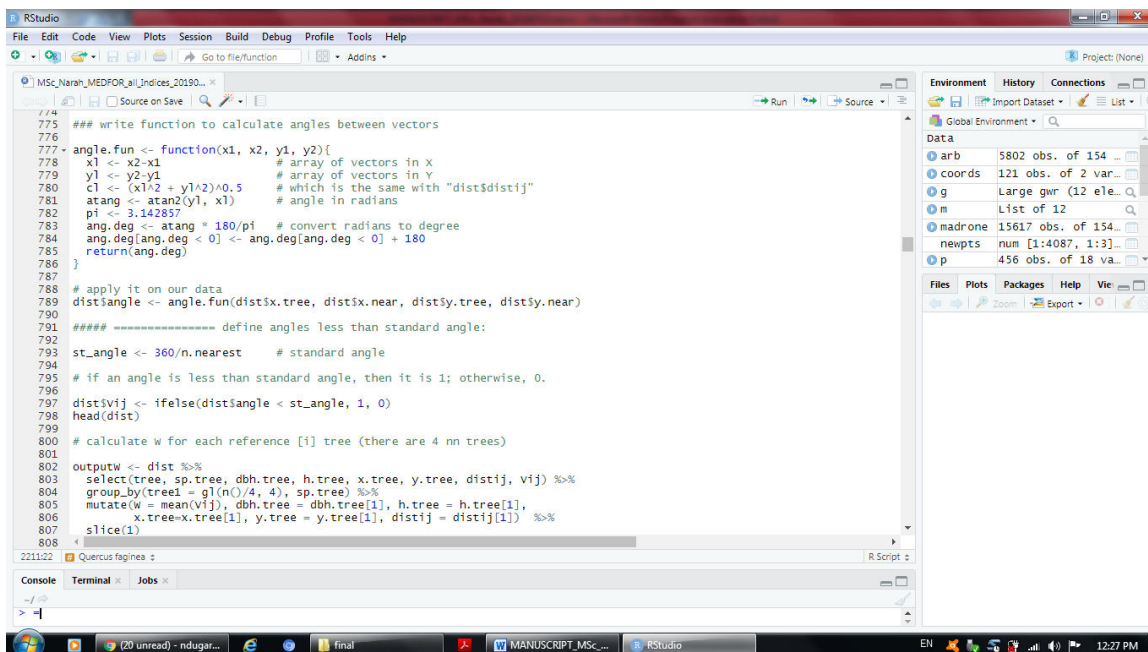
```

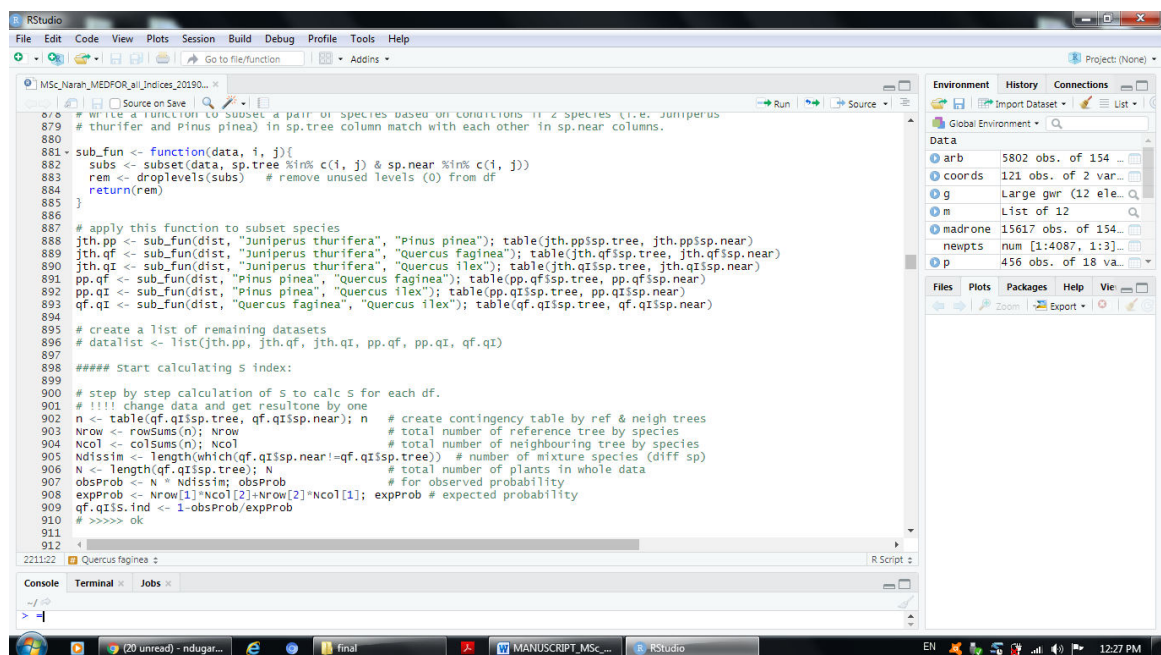
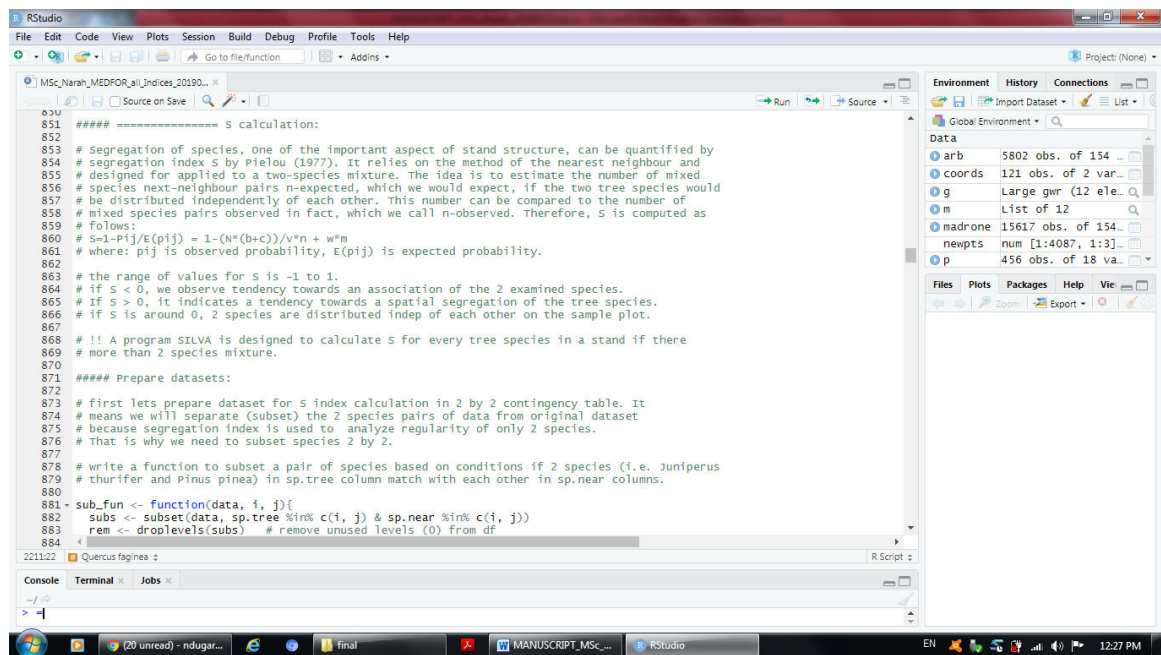
```

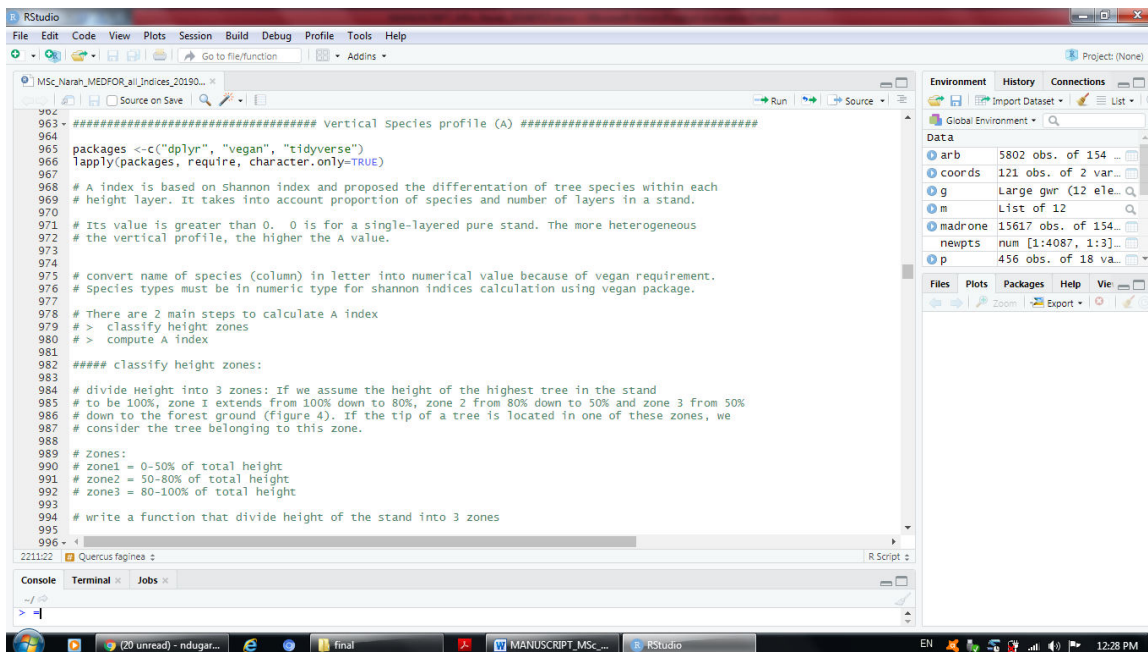
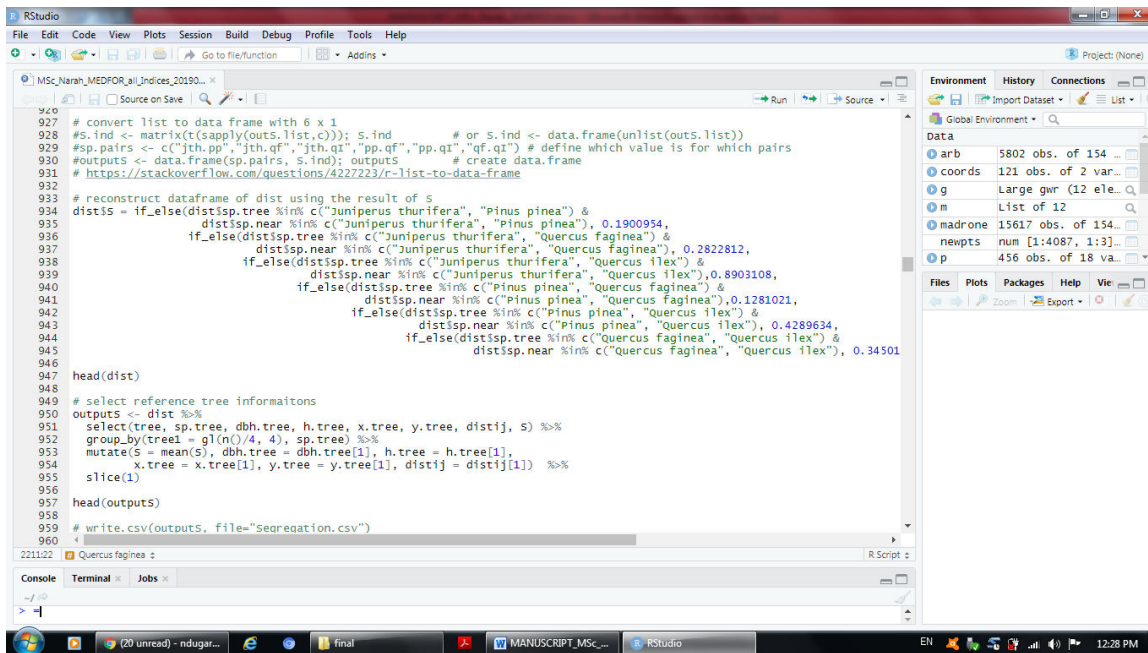
RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Nararh_MEDFOR_all_indices_20190...
000
607 write.csv(outputM1, file="mingling4.csv")
608
609 # we calculated M1 for other 4,5,6 neighbours and saved results as .csv files
610 # now read and integrate them into one dataset
611 #M13 <- read.csv("mingling.n3.csv", header = TRUE)
612 #M14 <- read.csv("mingling.n4.csv", header = TRUE)
613 #M15 <- read.csv("mingling.n5.csv", header = TRUE)
614 #M16 <- read.csv("mingling.n6.csv", header = TRUE)
615 #out.list <- list(M13, M14, M15, M16)
616 #outputM1 <- Reduce(function(x, y) merge(x, y, all=TRUE), out.list)
617
618 ## !!!! Following is another way to calc M1 for each ref tree. The result from both method
619 ## give the same result!. Back up it without discarding!!!
620
621 ## we calculate the mingling index of each reference [i] tree for the nearest neighbours
622 ## file =seq(1,length(dist[,1]), by=n.nearest)
623
624 ## we created sequential values according to location of reference trees along the rows
625 ## to identify location of reference tree from the row.
626 ## Values start from 1 to the last row of dist.m3 by its length of row [,1] with the
627 ## increment of 3.
628 ## output is numbers of locations like 1,4,7,..870 where reference tree info are.
629 ## the reason we doing this is that we are going to calculate M1 for each of these trees.
630
631 ## seq(from, to, by=...) is for creating values that starting and end (maximal) values in a
632 ## sequence by increment number.
633 ## length(dist.m[,1]) is length of rows (all rows) of dist.m data
634
635 # result = c()
636
637 ## c() is for create a vector, here creating an empty vector to save the
638 ## result of M1 calculation.
639
640
2211:22 Quercus faginea
R Script 2

```









```

RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Narah_MEDFOR_all_indices_20190...
# write a function that divide height of the stand into 3 zones
996- zones.fun <- function(h){
997- # determine break points that are going to be 50%, 80%, 100% values of max height
998- H100 <- max(h)
999- H80 <- H100 * 0.8
1000- H50 <- H100 * 0.5
1001- # divide the stand into 3 zones according to their height
1002- zones <- cut(h, breaks = c(0, H50, H80, H100),
1003-             labels = c("zone1", "zone2", "zone3"),
1004-             right=TRUE) # cutting values are included in right hand labels
1005- return(zones)
1006- }
1007-
1008- ##### ----- A index for a whole stand:
1009-
1010- data$hzones.st <- zones.fun(data$total_height_m)
1011-
1012- data <- data %>%
1013-   group_by(hzones.st) %>%
1014-   mutate(Atmp = diversity(spnum, "shannon"))
1015-
1016- # Standardization of A which can be done by dividing A by max value of A. Amax=log(S^2).
1017- Z <- 3 # number of height zones in the quadrant
1018- S <- length(unique(data$species)); s # number of species in a stand
1019- data$A.st <- data$Atmp/log(S^Z)
1020-
1021- ##### ----- A index for each quadrats:
1022-
1023- # define zones in each quadrant
1024- data <- data %>%
1025-   group_by(quadrant) %>%
1026-   mutate(hzones.qd = zones.fun(total_height_m))
1027-
1028-
2211:22 Quercus faginea
R Script 2
Console Terminal Jobs

```

```

RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Narah_MEDFOR_all_indices_20190...
# define zones in each quadrant
1024- data <- data %>%
1025-   group_by(quadrant) %>%
1026-   mutate(hzones.qd = zones.fun(total_height_m))
1027-
1028- # subset each quadrant separately
1029- q1 <- data[data$quadrant==1,]
1030- q2 <- data[data$quadrant==2,]
1031- q3 <- data[data$quadrant==3,]
1032- q4 <- data[data$quadrant==4,]
1033- q5 <- data[data$quadrant==5,]
1034- q6 <- data[data$quadrant==6,]
1035- q7 <- data[data$quadrant==7,]
1036- q8 <- data[data$quadrant==8,]
1037- q9 <- data[data$quadrant==9,]
1038- q10 <- data[data$quadrant==10,]
1039- q11 <- data[data$quadrant==11,]
1040- q12 <- data[data$quadrant==12,]
1041- q13 <- data[data$quadrant==13,]
1042- q14 <- data[data$quadrant==14,]
1043- q15 <- data[data$quadrant==15,]
1044- q16 <- data[data$quadrant==16,]
1045-
1046- # create list of all the data
1047- data.list <- list(q1,q2,q3,q4,q5,q6,q7,q8,q9,q10,q11,q12,q13,q14,q15,q16)
1048-
1049- # write a function to compute A for each quadrant
1050- A.fun <- function(data, hzones.qd, spnum) {
1051-   data %>%
1052-     group_by_at(hzones.qd) %>%
1053-     mutate(A = diversity(!rlang::sym(spnum), "shannon"))
1054- }
1055-
1056-
2211:22 Quercus faginea
R Script 2
Console Terminal Jobs

```

```

1069 # write.csv(data, file="VerProfInd.csv")
1070
1069 ##### Height difference index (TH) #####
1070
1071 packages <- c("dplyr")
1072 lapply(packages, require, character.only = TRUE)
1073
1074 # it is a single tree variable and be calculated for each pair of neighbours trees and reference/t.
1075 # also it can be calculated at a whole stand level (subpopulation) & at species level by
1076 # summing up TH values & divided it by the number of trees or individuals of the subpopulation.
1077
1078 # Formula is: Th = 1-min(Hi, Hj)/max(Hi, Hj); where Hi is height of reference tree
1079 # Hj is height of neighbour tree. For each pair the one with min value is divided by the one
1080 # with max value.
1081
1082 # 0 < TH < 1:
1083 # if TH=1, neighbour trees have high differentiation in height,
1084 # if TH=0, neighbour trees have equal height.
1085
1086 # Procedures:
1087 # > define nearest neighbour trees
1088 # > calculate TH index
1089
1090 # we have already calculated distances between reference tree and nearest neighbour trees.
1091 # using that, lets calculate TH
1092
1093 ##### Calculate height differ index TH:
1094
1095 ### compute TH for each pairs of reference tree and nearest neighbors. For example:
1096 # if we choose 3 neighbour trees, it will be calculated like T11, T12, T13 which mean that
1097 # i is reference tree to 1st neighbour
1098 # i reference tree to 2nd,
1099 # i reference tree to 3rd neighbour
1100
1101
1102
1103
1104
1105
1106
1107
1108
1109
1110
1111
1112
1113
1114
1115
1116
1117
1118
1119
1120
1121
1122
1123
1124
1125
1126
1127
1128
1129
1130
1131
1132
1133
1134
1135
1136

```

```

1102 # pmin () and pmax() returns parallel min or max of two or more input vectors
1103
1104
1105
1106 names(dist)
1107
1108
1109 outputTH <- dist %>%
1110   select(tree, sp.tree, dbh.tree, h.tree, x.tree, y.tree, distij, THij) %>%
1111   group_by(tree1 = gl(n()/4, 4), sp.tree) %>%
1112   mutate(TH = mean(THij), dbh.tree = dbh.tree[1], h.tree = h.tree[1],
1113          x.tree=x.tree[1], y.tree = y.tree[1], distij = distij[1]) %>%
1114   slice(1)
1115
1116
1117 ### the average of T by species
1118 TH.sp <- outputTH %>%
1119   select(sp.tree, tree1, TH) %>%
1120   group_by(sp.tree) %>%
1121   summarise(TH.sp = mean(TH)); TH.sp
1122
1123
1124 ### merge all outputs into one single data
1125
1126 # delete variables that have the same name but with different values. otherwise
1127 # when merged datas, the length increases a lot.
1128 outputM1 <- subset(outputM1, select = -c(mvij))
1129 outputs <- subset(outputs, select = -c(tree1))
1130 outputTH <- subset(outputTH, select = -c(tree1, THij))
1131
1132 out.list <- list(data, outputM1, outputW, outputs, outputTH)
1133 results <- Reduce(function(x, y) merge(x, y, all = TRUE), out.list)
1134
1135 # delete some midd calc variables and just keep main variables
1136 final.results <- subset(results, select = -c(sp.tree, dbh.tree, h.tree, x.tree, y.tree, distij,
1137                                            Hzones.st, Atmp, A.st, Hzones.qd, G1.sp, G1.qd,
1138                                            nm, vij, tree1))

```



```
RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Naran_MEDFOR_all_indices_20190...
write.csv( final_results, file="all_indices.csv")
##### STATISTICAL ANALYSIS #####
##### statistical analysis at stand level #####
# https://rpubs.com/Marcelobn/many_regressions
# https://stackoverflow.com/questions/37395059/running-several-linear-regressions-from-a-single-dataframe-in-r
# https://stackoverflow.com/questions/25036007/linear-regression-loop-for-each-independent-variable-individually-against-d
# https://sebastiansauer.github.io/multiple-lm-purrr/
# https://stackoverflow.com/questions/3922335/fitting-polynomial-model-to-data-in-r
# https://stackoverflow.com/questions/49344507/save-individual-model-output-from-a-list-of-model-summary
# https://sejdemyr.github.io/r-tutorials/basics/tables-in-r/
packages <- c("MASS", "dplyr", "broom", "purrr", "tidyverse", "stargazer", "sjplot", "splines", "nortest")
lapply(packages, require, character.only = TRUE)
setwd("D:/MEDFOR_classes/MScThesis_MEDFOR_UVA/data/processed")
data11 <- read.csv("all_indices.csv", header = TRUE)
# subset variables that will be used for corr analyze
data <- data11[,c("btot_kg", "species", "DBH_cm", "Total_height_m", "v_m3", "gi_m2",
"gi_m2_ha", "gi_pr.sp.qd", "simpson", "shannon", "Even", "D", "M1", "MS",
"W", "S", "A", "TH")]
names(data) <- c("B", "species", "d", "h", "v", "G1_m2", "G1_qd", "G1_pr_qd",
"Sm", "Sn", "E", "D", "M1", "MS", "W", "S", "A", "TH")
# Main steps are:
# > fit several different types of models for each variable
# > select best fitted models based on
# > 1. significance of independent variable (P value < 0.05)
# > 2. Biological meaning & Parameterization (parameters should not be less than 0)
# > 4. Normality of residual, Homociducty
Quercus faginea
Console Terminal Jobs
> |
```

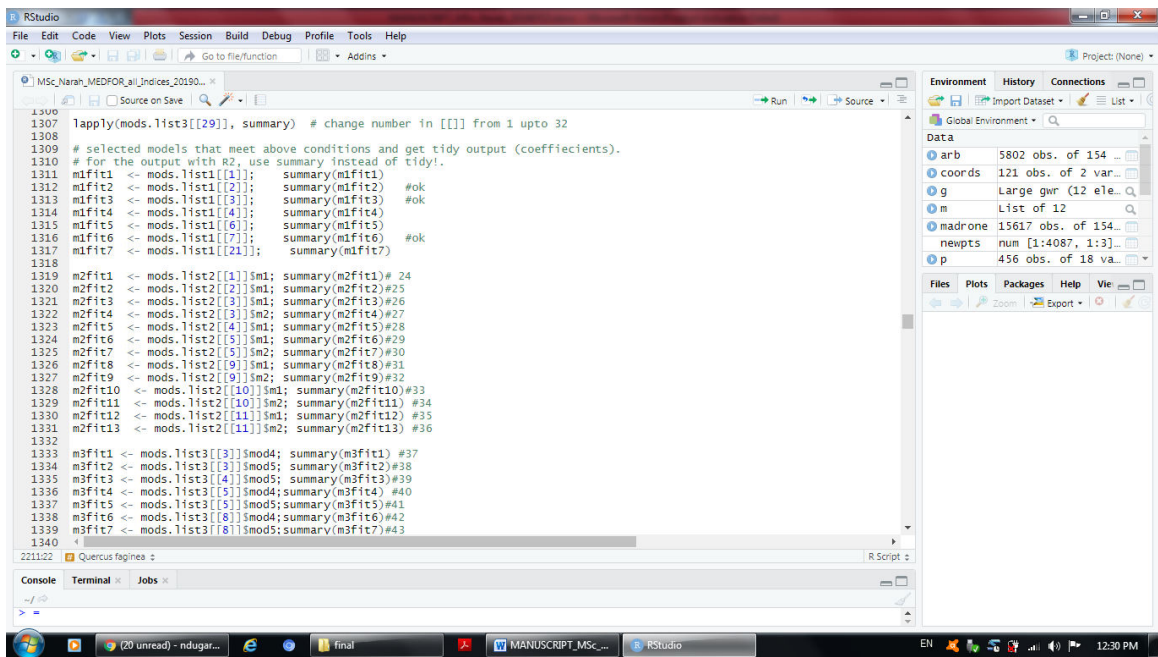
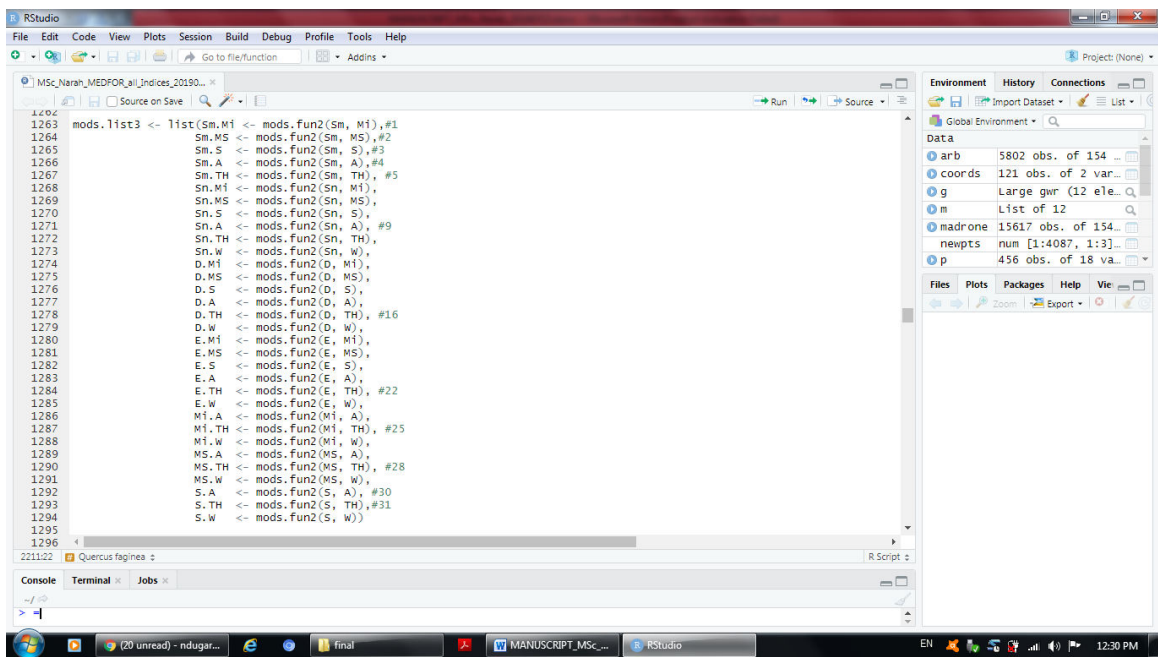
```
RStudio
File Edit Code View Plots Session Build Debug Profile Tools Help
MSC_Naran_MEDFOR_all_indices_20190...
names(data) <- c("B", "species", "d", "h", "v", "G1_m2", "G1_qd", "G1_pr_qd",
"Sm", "Sn", "E", "D", "M1", "MS", "W", "S", "A", "TH")
# Main steps are:
# > fit several different types of models for each variable
# > select best fitted models based on
# > 1. significance of independent variable (P value < 0.05)
# > 2. Biological meaning & Parameterization (parameters should not be less than 0)
# > 4. Normality of residual, Homociducty
# > Rank models by AIC, R2 & BIC etc
# define variables with log transformation
logB <- log(data11$btot_kg)
logG1 <- log(data11$gi_m2_ha) # G1 for quadrant
logd <- log(data11$DBH_cm)
logh <- log(data11$Total_height_m)
logv <- log(data11$v_m3)
logSm <- log(data11$simpson)
logSn <- log(data11$shannon)
logE <- log(data11$Even)
logG <- log(data11$G)
logM1 <- log(data11$M1+1)
logMS <- log(data11$MS+1)
logW <- log(data11$W+1)
logS <- log(data11$S+1)
logA <- log(data11$A+1)
logTH <- log(data11$TH)
mods.list1 <- list(m1 <- lm(logB ~ logG1 + logTH),
m2 <- lm(logB ~ logG1 + logTH + logSm),
m3 <- lm(logB ~ logG1 + logTH + logSn),
m4 <- lm(logB ~ logG1 + logTH + logd),
m5 <- lm(logB ~ logG1 + logTH + logE).
Quercus faginea
Console Terminal Jobs
> |
```

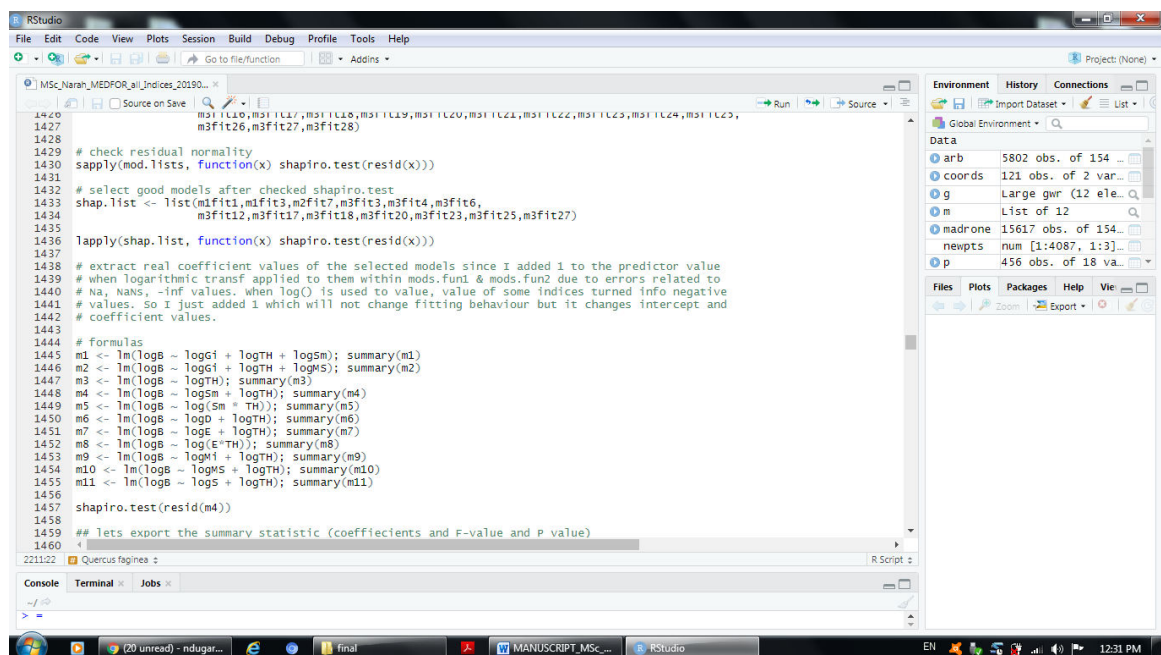
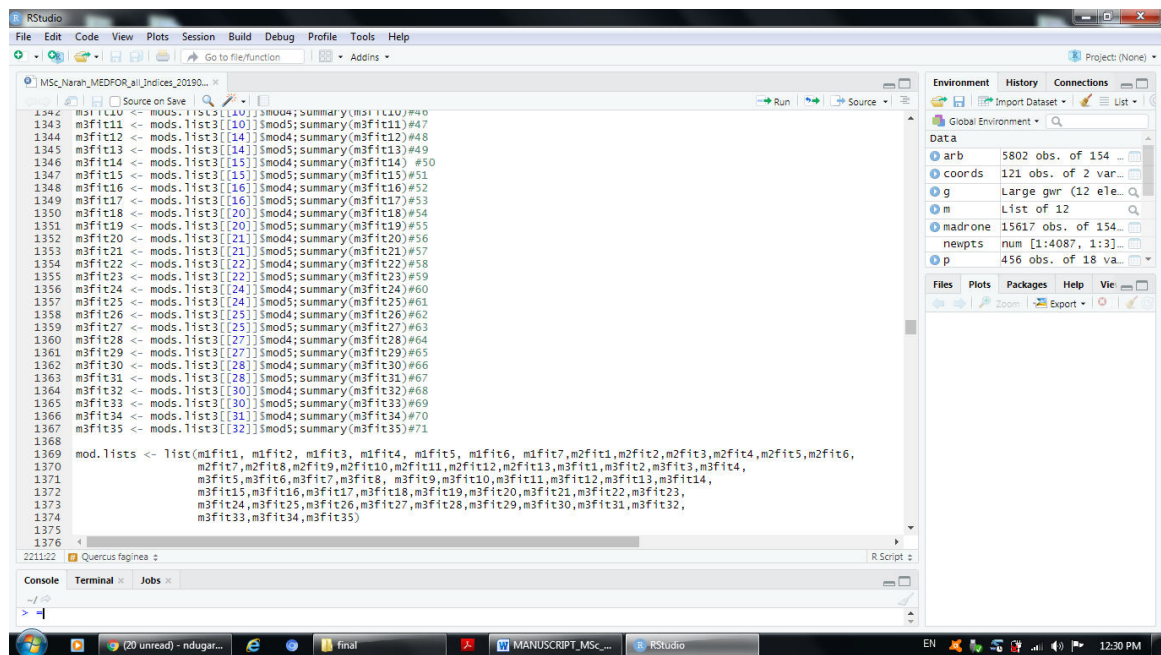
```

1191 mods.list1 <- list(m1 <- lm(logB ~ logG1 + logTH),
1192 m2 <- lm(logB ~ logG1 + logTH + logSm),
1193 m3 <- lm(logB ~ logG1 + logTH + logSn),
1194 m4 <- lm(logB ~ logG1 + logTH + logD),
1195 m5 <- lm(logB ~ logG1 + logTH + logE),
1196 m6 <- lm(logB ~ logG1 + logTH + logW1),
1197 m7 <- lm(logB ~ logG1 + logTH + logWS),
1198 m8 <- lm(logB ~ logG1 + logTH + logS),
1199 m9 <- lm(logB ~ logG1 + logTH + logA),
1200 m10 <- lm(logB ~ logG1 + logTH + logW),
1201 m11 <- lm(logB ~ logG1 + logA),
1202 m12 <- lm(logB ~ logG1 + logA + logSm),
1203 m13 <- lm(logB ~ logG1 + logA + logSn),
1204 m14 <- lm(logB ~ logG1 + logA + logD),
1205 m15 <- lm(logB ~ logG1 + logA + logE),
1206 m16 <- lm(logB ~ logG1 + logA + logW1),
1207 m17 <- lm(logB ~ logG1 + logA + logWS),
1208 m18 <- lm(logB ~ logG1 + logA + logS),
1209 m19 <- lm(logB ~ logG1 + logA + logW),
1210 m20 <- lm(logB ~ logSm + logW1 + logA),
1211 m21 <- lm(logB ~ logD + logWS + logTH),
1212 m22 <- lm(logB ~ logE + logS + logA),
1213 m23 <- lm(logB ~ logA + logTH + logW1),
1214 m24 <- lm(logB ~ logH + logG1 + logTH),
1215 m25 <- lm(logB ~ logSm + logS + logTH),
1216 m26 <- lm(logB ~ logS + logW + logG1),
1217 m27 <- lm(logB ~ logTH + logG1 + logW + logA),
1218 m28 <- lm(logB ~ logG1 + logS + logA + logWS),
1219 m29 <- lm(logB ~ logG1 + logSm + logW + logE))
1221
1222 # define variables without log
1223 B <- data1$tot_kq
1224
1221:22 Quercus faginea
    
```

```

1234 B <- data1$B
1235 A <- data1$A
1236 TH <- data1$TH
1237
1238 # write a function for 2 models that takes a single input
1239 mods.fun1 <- function(x) {
1240 m1 <- lm(log(B) ~ I(log(x+1)))
1241 m2 <- lm(log(B) ~ I(log(x-1)) + I(log(x+1)^2))
1242 list <- list(m1, m2)
1243 names(list) <- paste0("m", 1:2)
1244 return(list)
1245 }
1246
1247 # apply mod1.fun to each predictor variable (14 predictor variable)
1248 mods.list2 <- lapply(data[-c(1:6,8)], function(x) mods.fun1(x))
1249 summary(mods.list2) # this is a list of 13 models that include 5 models inside.
1250
1251 # write function taking 2 predictor variable (without d & h)
1252 mods.fun2 <- function(x, y) {
1253 m1 <- lm(B ~ x + y)
1254 m2 <- lm(B ~ I(x^2) + I(y^2))
1255 m3 <- lm(B ~ I(x*y))
1256 m4 <- lm(log(B) ~ I(log(x+1)) + I(log(y+1)))
1257 m5 <- lm(log(B) ~ I(log(x*y+1)))
1258 list <- list(m1, m2, m3, m4, m5)
1259 names(list) <- paste0("mod", 1:5)
1260 return(list)
1261 }
1262
1263 mods.list3 <- list(Sm.M1 <- mods.fun2(Sm, M1), #1
1264 Sm.MS <- mods.fun2(Sm, MS), #2
1265 Sm.S <- mods.fun2(Sm, S), #3
1266 Sm.A <- mods.fun2(Sm, A), #4
1267 Sm.TH <- mods.fun2(Sm, TH), #5
1268
    
```





```

1483 fit0.st %>% mutate(model = 0),
1484 fit7.st %>% mutate(model = 7),
1485 fit8.st %>% mutate(model = 8),
1486 fit9.st %>% mutate(model = 9),
1487 fit10.st %>% mutate(model = 10),
1488 fit11.st %>% mutate(model = 11))
1489
1490 # Export
1491 # write.csv(sel.models, file = "signifmods.st.csv")
1492
1493 # AIC
1494 mod.list <- list(m1,m2,m3,m4,m5,m6,m7,m8,m9,m10,m11)
1495
1496 lapply(mod.list, function(x) shapiro.test(resid(x)))
1497
1498 sapply(mod.list, AIC)
1499
1500 # export result
1501 stargazer(m1, m2, m3, m4, m5,m6,type = "text", single.row = TRUE,
1502           align=TRUE, no.space = TRUE, out="11signifmods1sec.st.txt")
1503 stargazer(m7, m8, m9, m10,m11, type = "text", single.row = TRUE,
1504           align=TRUE, no.space = TRUE, out="11signifmods2sec.st.txt")
1505
1506 # plot both residual & qqnorm plots together
1507 par(mfrow=c(1,2))
1508 plot(m1, which=c(1,2))
1509
1510 # according to residual normality test, fit 8 and 10 are not good. So eliminate them.
1511
1512
1513
1514
1515 ##### calculate metrics of each model
1516
2211:22
Quercus faginea
R Script 2

```

```

1702
1703
1704
1705 ##### by species #####
1706
1707 ##### Pinus pinea #####
1708
1709 packages <- c("MASS","dplyr","broom","purrr","tidyverse","stargazer","sjPlot","splines")
1710 lapply(packages, require, character.only = TRUE)
1711
1712 setwd("D:/MEDFOR_classes/MScThesis_MEDFOR_UVA/data/processed")
1713 dataAll <- read.csv("all_indices.csv", header = TRUE)
1714
1715 # subset variables that will be used for corr analyze
1716 data <- dataAll[,c("Btot_kg","species","DBH_cm","Total_height_m","V_m3","Gi_m2",
1717                  "Gi_m2_ha","Gi_pr_sp_qd","simpson","shannon","Even","D","M1","MS",
1718                  "W","S","A","TH")]
1719
1720 names(data) <- c("B","species","d","h","v","Gi_m2","Gi_qd","Gi_pr_qd",
1721                 "sm","sn","E","D","M1","MS","w","s","A","TH")
1722
1723 # subset data by species
1724 data.pp <- subset(data, species=="Pinus pinea")
1725 data.jth <- subset(data, species=="Juniperus thurifera")
1726 data.qf <- subset(data, species=="Quercus faginea")
1727 data.qi <- subset(data, species=="Quercus ilex")
1728 data.qr <- subset(data, species %in% c("Quercus faginea","Quercus ilex"))
1729
1730
1731 # define variables with log transformation
1732 logB <- log(data.pp$B)
1733 logGi <- log(data.pp$Gi_pr_qd) # Gi proportion of species at quadrant
1734 logd <- log(data.pp$d)
1735 looh <- log(data.pp$h)
1736
1513:1
statistical analysis at stand level
R Script 2

```

