1	Influence of liquid-to-biogas ratio and alkalinity on the biogas
2	upgrading performance in a demo scale algal-bacterial
3	photobioreactor
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## 16 ABSTRACT

The influence of the liquid-to-biogas ratio (L/G) and alkalinity on methane quality was 17 evaluated in a 11.7 m<sup>3</sup> outdoors horizontal semi-closed tubular photobioreactor 18 19 interconnected to a 45-L absorption column (AC). CO2 concentrations in the upgraded 20 methane ranged from <0.1 to 9.6% at L/G of 2.0 and 0.5, respectively, with maximum CH<sub>4</sub> concentrations of 89.7% at a L/G of 1.0. Moreover, an enhanced CO<sub>2</sub> removal 21 (mediating a decrease in CO<sub>2</sub> concentration from 9.6 to 1.2%) and therefore higher CH<sub>4</sub> 22 23 contents (increasing from 88.0 to 93.2%) were observed when increasing the alkalinity of the AC cultivation broth from  $42\pm1$  mg L<sup>-1</sup> to  $996\pm42$  mg L<sup>-1</sup>. H<sub>2</sub>S was completely 24 removed regardless of the L/G or the alkalinity in AC. The continuous operation of the 25 photobioreactor with optimized operating parameters resulted in contents of CO<sub>2</sub> (<0.1%-26

1.4%), H<sub>2</sub>S (<0.7 mg m<sup>-3</sup>) and CH<sub>4</sub> (94.1%-98.8%) complying with international
regulations for methane injection into natural gas grids.

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Keywords: Algal-bacterial photobioreactor, Alkalinity, Biogas upgrading, Liquid/Gas
ratio, Outdoors conditions.

32

#### 33 **1. Introduction**

The anaerobic digestion (AD) of organic solid waste and sludge from wastewater 34 treatment generates a biogas that represents a potential renewable energy source capable 35 36 of generating electricity and reduce the dependence on fossil fuels (Muñoz et al., 2015). 37 Biogas can be purified and injected into natural gas grids or used as a vehicle fuel, or desulphurised and used for the generation of domestic heat or steam and electricity in 38 39 industry (Andriani et al., 2014; Muñoz et al., 2015). In this regard, a growing contribution of biogas to the EU energy sector has been observed within the past years, with an 40 41 increase in the numbers of biogas producing plants by a factor of 3 (from 6,772 in 2009 to 17,439 by the end of 2016) (European Biogas Association, 2017). The upgrading of 42 43 biogas prior injection into natural gas grids or use as a vehicle fuel is required due to the 44 large number and high concentrations of impurities in raw biogas: CO<sub>2</sub> (15-60%), H<sub>2</sub>S (0.005-2%), O<sub>2</sub> (0-1%), N<sub>2</sub> (0-2%), CO (<0.6%), NH<sub>3</sub> (<1%), siloxanes (0-0.2%) and 45 volatile organic compounds (<0.6%) (Ryckebosch et al., 2011). In this context, most 46 international regulations establish that a methane composition of  $CH_4 \ge 95\%$ ,  $CO_2 \le 2$ -47 4%,  $O_2 \leq 1\%$  and negligible amounts of  $H_2S$  is mandatory for its injection into natural 48 gas grids, while a lower CH<sub>4</sub> content is required when methane is used as a vehicle fuel 49 (Muñoz et al., 2015). The removal of biogas contaminants like H<sub>2</sub>S reduces the corrosion 50 in pipelines, engines and biogas storage structures, while the reduction in CO<sub>2</sub> contributes 51

to increase the calorific value of methane and reduces its transportation costs (Posadas etal., 2015).

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Nowadays, several biological technologies are available to remove CO<sub>2</sub> and H<sub>2</sub>S from 55 biogas. For instance, chemoautotrophic biogas upgrading is used for the removal of CO<sub>2</sub>, 56 57 while biofiltration or in situ micro-aerobic AD are applied for H<sub>2</sub>S removal (Farooq et 58 al., 2018; Marín et al., 2018a; Muñoz et al., 2015). The removal of only one biogas contaminant at a time represents the main disadvantage associated to these biological 59 technologies, resulting in the need of implementing two-stage biological upgrading 60 61 processes. Likewise, several physical-chemical technologies are commercially available 62 to remove CO<sub>2</sub> and H<sub>2</sub>S from biogas. Membrane separation, pressure swing adsorption or chemical/water/organic scrubbing are applied for CO<sub>2</sub> removal, while *in-situ* chemical 63 64 precipitation or adsorption onto activated carbon or metal ions provide satisfactory levels of H<sub>2</sub>S removal (Marín et al., 2018a; Muñoz et al., 2015; Toledo-cervantes et al., 2017). 65 Two sequential stages are also necessary for a complete biogas upgrading, which entails 66 an increase in investment and operational costs. In this context, algal-bacterial 67 photobioreactors can be engineered as an environmentally friendly and cost-effective 68 69 technology due to their capacity to simultaneously remove  $CO_2$  and  $H_2S$  in a single stage 70 process (Bahr et al., 2014).

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Algal-bacterial processes have emerged as a cost-competitive technology capable of removing CO<sub>2</sub> and H<sub>2</sub>S from biogas in a single stage at low environmental impacts (Bahr et al., 2014; Muñoz et al., 2015). Biogas upgrading in algal-bacterial photobioreactors is based on the simultaneous photosynthetic fixation of CO<sub>2</sub> by microalgae and the oxidation of H<sub>2</sub>S to SO<sub>4</sub><sup>2-</sup> by sulfur oxidizing bacteria promoted by the high dissolved

oxygen (DO) concentration present in the cultivation broth as a result of photosynthesis 77 78 (Posadas et al., 2017, 2015; Toledo-Cervantes et al., 2016). Photosynthetic biogas 79 upgrading has been recently evaluated indoors in high rate algal ponds (HRAPs) interconnected to a biogas absorption column (AC) under artificial illumination. Bahr et 80 al. (2014) demonstrated for the first time the capability of microalgal-bacterial processes 81 for the simultaneous removal of CO<sub>2</sub> and H<sub>2</sub>S from biogas. Serejo et al. (2015) studied 82 83 the influence of the liquid/biogas (L/G) ratio on the composition of the upgraded biogas. Posadas et al. (2016) optimized the biogas upgrading process in a HRAP using centrate 84 as a source of nutrients under laboratory conditions, while Rodero et al. (2018) evaluated 85 86 the influence of alkalinity and temperature on the photosynthetic biogas upgrading efficiency in an indoor HRAP. In addition, Posadas et al. (2017) evaluated the 87 simultaneous biogas upgrading and centrate treatment in a HRAP operated under 88 89 outdoors conditions during summer, while Marín et al. (2018a,b) investigated the influence of the yearly variations of environmental conditions on the biogas upgrading 90 91 performance. Nevertheless, and despite the satisfactory results obtained so far, new photobioreactor configurations should be tested in order to overcome design constraints 92 93 associated to algal ponds such as their high footprint. In this sense, semi-closed or closed 94 tubular photobioreactors have been proposed as a promising alternative to reduce land requirement, while offering higher photosynthetic efficiencies, enhanced biomass 95 productivities and a superior CO<sub>2</sub> mass transfer (Toledo-Cervantes et al., 2018). 96

97

98 This study investigated for the first time the biogas upgrading potential of an outdoors
99 pilot-scale hybrid (semi-closed) horizontal tubular photobioreactor (PBR) interconnected
100 to an external AC. The influence of the L/G ratio and the alkalinity of the cultivation

medium in the AC on the quality of the upgraded biogas was assessed and optimized. In
addition, the PBR-AC was operated continuously under optimized process parameters.

103

104 **2. Materials and methods** 

#### 105 **2.1 Biogas**

The biogas used in this experiment was obtained from the anaerobic digestion of microalgal biomass in a pilot anaerobic digester located at the Agròpolis experimental campus of the Universitat Politécnica de Catalunya-BarcelonaTech (Catalunya, Spain) (García et al., 2018; Uggetti et al., 2018). The average biogas composition was CO<sub>2</sub> (13.7  $\pm$  1.0%), H<sub>2</sub>S (0.1  $\pm$  0.05%) and CH<sub>4</sub> (86.2  $\pm$  1.0%).

111

### 112 2.2 Experimental set-up

113 The experimental set-up was built outdoors at the Agròpolis experimental campus of the Universitat Politécnica de Catalunya-BarcelonaTech (41.29°N, 2.04°E). The horizontal 114 115 hybrid (semi-closed) tubular photobioreactor (PBR) consisted of 2 lateral open tanks 116 made of polypropylene (width=1 m; length=5 m; depth=0.6 m) interconnected by 16 low density transparent polyethylene tubes (length=47 m; diameter=125 mm). The total 117 working volume of the PBR was 11.7 m<sup>3</sup>. The cultivation broth was continuously 118 circulated in each tank by a 6-blade paddlewheel with a rotational speed of 9-12 rpm, 119 120 which resulted in a velocity of the cultivation broth inside the tubes of 0.20-0.25 m s<sup>-1</sup>. 121 This recirculation rate ensured a homogeneous distribution and mixing of the cultivation broth and a turbulent flow inside the tubes, avoiding biomass settling. The different height 122 123 level between the two open tanks caused a gravity flow through 8 tubes from the deep side of one tank to the shallow side of the opposite one (Uggetti et al., 2018). The open 124 tanks supported the release of the DO accumulated along the closed tubes and also 125

126	provided a cooling effect via water evaporation, thus preventing the occurrence of the
127	extremely high temperatures that would be reached in completely closed tubular PBRs.
128	The PBR was interconnected to a separate 45 L bubble AC (internal diameter=12 cm;
129	height=4 m) made of PVC and provided with a ring of seven metallic biogas diffusers of
130	$2 \mu m$ pore size located at the bottom of the column. (Fig. 1).
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132	<figure 1=""></figure>
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134	2.3 Operational conditions and experimental procedure
135	The PBR was inoculated at an initial concentration of 220 mg volatile suspended solids
136	(VSS) L <sup>-1</sup> with a microalgal consortium composed of <i>Chlorella vulgaris, Stigeoclonium</i>
137	tenue, Nitzschia closterium and Navicula amphora, obtained from an outdoors HRAP
138	located at the facilities of the Environmental Engineering and Microbiology Research
139	Group (GEMMA) the Universitat Politécnica de Catalunya-BarcelonaTech (Gutiérrez et
140	al., 2016). The PBR was operated as the third of a set of 3 identical PBRs interconnected
141	in series and treating 2.3 $m^3$ d <sup>-1</sup> of agricultural wastewater with the following
142	composition: total organic carbon (TOC) = $131 \pm 80 \text{ mg L}^{-1}$ , inorganic carbon (IC) = $36$
143	$\pm$ 10 mg L <sup>-1</sup> , total nitrogen (TN) = 15 $\pm$ 7 mg L <sup>-1</sup> and total phosphorus (TP) = 0.9 $\pm$ 1.0
144	mg L <sup>-1</sup> . Three experimental series were conducted as described below:
145	
146	2.3.1 Influence of the liquid-to-biogas ratio in the absorption column on the quality of the
147	upgraded biogas

148 L/G ratios ranging from 0.5 to 5.0 were tested in order to optimize the quality of the 149 upgraded biogas. Biogas was sparged into the AC at  $100 \text{ L} \text{ d}^{-1}$ , while the cultivation broth 150 from the PBR was supplied in co-current mode at different flow rates in order to provide

151	L/G ratios of 0.5, 1.0, 2.0, 3.0, 4.0 and 5.0. The duration of each L/G ratio condition was
152	at least four times the hydraulic retention time (HRT) of the liquid in the AC (Table 1).
153	The ambient and cultivation broth temperatures, the pH, dissolved TOC, IC, TN, N-NH4 <sup>+</sup>
154	and TP concentrations in the cultivation broth of the PBR, and the composition of the raw
155	and upgraded biogas were analyzed in triplicate at the end of each operational condition.
156	
157	<table 1=""></table>
158	
159	2.3.2 Influence of the alkalinity in the cultivation broth on the quality of the upgraded
160	biogas
161	In order to assess the impact of different alkalinities of the cultivation broth in the AC on
162	the upgrading efficiency, a carbonate solution (NaHCO3 and Na2CO3) with a
163	concentration of 16,000 mg L <sup>-1</sup> of IC was injected at the bottom of the AC in co-current
164	mode (Fig. 1, dashed line). Biogas flowrate and L/G ratio were fixed at 100 L d <sup>-1</sup> and 0.5,
165	respectively. Carbonate solution flow rates of 0, 1, 2, 3 and 5 L $d^{-1}$ (corresponding to an
166	IC concentration in the cultivation broth of the AC of $42 \pm 1$ ; $311 \pm 6$ ; $634 \pm 48$ ; $996 \pm$
167	42 and 1,557 $\pm$ 26 mg L <sup>-1</sup> , respectively) were tested in order to optimize the quality of the
168	upgraded biogas. Each carbonate solution flowrate was maintained for at least four times
169	the HRT of the liquid in the AC. The ambient and PBR cultivation broth temperatures,
170	the pH, dissolved TOC, IC, TN, $N-NH_4^+$ and TP concentrations in the cultivation broth
171	of the PBR, and the composition of the raw and the upgraded biogas were analyzed in
172	triplicate at the end of each operational condition.
173	
174	2.3.3 Continuous photosynthetic biogas upgrading operation

Biogas upgrading performance of the demo scale PBR was evaluated throughout 42 days under continuous operation. The optimum operating parameters previously identified were selected: biogas flowrate of 100 L d<sup>-1</sup>, L/G ratio of 0.5 and the supplementation of 2.0 L d<sup>-1</sup> of carbonate solution to the AC. The ambient and cultivation broth temperatures, the pH, dissolved TOC, IC, TN, N-NH4<sup>+</sup> and TP concentrations in the cultivation broth of the PBR, and the composition of the raw and the upgraded biogas were analyzed in duplicate once per week.

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#### **183 2.4 Analytical procedures**

184 The concentration of CH<sub>4</sub>, CO<sub>2</sub>, N<sub>2</sub> and O<sub>2</sub> in biogas and methane were determined using a gas chromatograph (GC) equipped with a thermal conductivity detector (Trace GC 185 186 Thermo Finnigan with Hayesep packed column). Injector, detector and oven temperatures 187 were maintained at 150, 250 and 35 °C, respectively, with helium as a carrier gas. The concentration of H<sub>2</sub>S in the raw biogas was determined using Gastec colorimetric tubes, 188 189 while its concentration in the upgraded methane was analyzed by a Dräger X-am 5000 190 electrochemical sensor (lower detection limit of 0.5 ppm<sub>v</sub>). Temperature and pH were 191 measured in-situ by a pH-meter with temperature sensor (Mettler Toledo, USA). 192 Dissolved TOC, IC and TN concentrations were determined using a C/N analyzer (21005, Analytikjena, Germany). The analysis of TP concentration was performed according to 193 the Ascorbic Acid Method of Standard Methods (APHA, 2005), while N-NH4<sup>+</sup> 194 195 concentration was measured by a colorimetric method according to Solorzano (1969). The determination of the concentration of total suspended solids (TSS) and VSS in the 196 197 PBR was performed according to Standard Methods (APHA, 2005), and the temperature of the cultivation broth was periodically monitored with a temperature sensor (Campbell 198 Scientific Inc., USA). 199

## 201 **3. Results and discussion**

3.1 Influence of the liquid-to-biogas ratio in the absorption column on the quality of the
upgraded biogas

The composition of the methane produced in the PBR-AC varied depending on the L/G 204 205 ratio tested (Fig. 2). At a L/G ratio of 2.0, CO<sub>2</sub> was not detected in the upgraded methane, 206 thus achieving minimum concentrations < 0.1% according to the GC detection limit. On the contrary, a maximum concentration of  $9.6 \pm 0.1\%$  was recorded at a L/G ratio of 0.5 207 208 (Fig. 2). These results were in accordance with Posadas et al. (2017), who recorded the 209 highest concentration of CO<sub>2</sub> in methane at the lowest L/G ratio (≈12.0% at a L/G ratio 210 of 0.5). L/G ratios > 2.0 supported a significant decrease in the CO<sub>2</sub> concentration of the 211 upgraded biogas, which ranged from <0.1 to 1.4% (corresponding to removal efficiencies 212 (REs) between 90.4 and >99.9%). On the other hand, H<sub>2</sub>S was not detected in the upgraded methane regardless of the tested L/G ratio, its complete removal being attributed 213 214 to the high aqueous solubility of this biogas contaminant. An efficient removal of H<sub>2</sub>S 215 from raw biogas in algal-bacterial PBRs with a negligible impact of the L/G ratio has 216 been consistently reported both in outdoors (Posadas et al. 2017) and indoors HRAPs 217 (Serejo et al. 2015).

218

Unfortunately, the concentrations of  $N_2$  and  $O_2$  recorded in the upgraded biogas increased from 3.4% at a L/G ratio of 0.5 to 11.9% at a L/G ratio of 5.0 (Fig. 2), which clearly indicated that the stripping of these gases from the recirculating cultivation broth was promoted at higher liquid flowrates (Sovechles and Waters, 2015). These results were in accordance with Toledo-Cervantes et al. (2016), who reported  $N_2/O_2$  concentrations between 2.5 and 37.0% at L/G ratios ranging from 0 to 40 in a closed tubular

photobioreactor. Likewise, Posadas et al. (2017) also reported an increase in N2 and O2 225 226 concentration in the upgraded biogas from 1.4 to 18.3% when the L/G ratio increased from 0.5 to 5, respectively. Similarly, Rodero et al., 2019 found N<sub>2</sub>/O<sub>2</sub> concentrations 227 ranging from 6.6 and 11.4% at L/G ratios ranging from 1.2 to 3.5 in an outdoors HRAP. 228 229 Finally, a maximum concentration of CH<sub>4</sub> of 89.7% in the upgraded biogas was recorded 230 231 at a L/G ratio = 1 (Fig. 2). Interestingly, although further increases in the L/G ratio resulted in lower CO<sub>2</sub> concentrations, they also mediated a higher desorption of N<sub>2</sub> and 232 O<sub>2</sub>, which negatively impacted the final concentration of CH<sub>4</sub> in the upgraded biogas. 233 234 <Figure 2> 235 236 237 3.2 Influence of the alkalinity in the cultivation broth on the quality of the upgraded biogas 238 239 The supplementation of a carbonate solution to the absorption column resulted in an 240 improved quality of the final methane. In this context, average concentrations of CO<sub>2</sub> of  $9.6 \pm 0.2$ ;  $2.6 \pm 0.2$ ;  $1.3 \pm 0.0$ ;  $1.2 \pm 0.0$  and  $1.1 \pm 0.2\%$  were recorded at IC concentrations 241 in the AC cultivation broth of 42  $\pm$  1; 311  $\pm$  6; 634  $\pm$  48; 996  $\pm$  42 and 1,557  $\pm$  26 mg L  $^{-}$ 242 <sup>1</sup>, respectively (Fig. 3). The increase in CO<sub>2</sub>-REs resulting from the addition of alkalinity 243 (from 24.0  $\pm$  0.2% at 42  $\pm$  1 mg IC L<sup>-1</sup> to 91.9  $\pm$  0.2% at 1,557  $\pm$  26 mg IC L<sup>-1</sup>) was 244 associated to the concomitant increase of pH in the cultivation broth of the AC (from 6.5 245  $\pm$  0.1 at 42  $\pm$  1 mg IC L<sup>-1</sup> up to 9.3  $\pm$  0.0 at 1,557  $\pm$  26 mg IC L<sup>-1</sup>). The beneficial effect 246 of alkalinity on CO<sub>2</sub> removal performance has been previously reported in literature. For 247 instance, Rodero et al. (2018) reported CO<sub>2</sub>-REs of 97.8  $\pm$  0.8, 50.6  $\pm$  3.0 and 41.5  $\pm$  2.0% 248 during the operation of an indoors HRAP interconnected to an AC using a feeding nutrient 249

solution with an average IC concentration of 1,500 mg L<sup>-1</sup>, 500 mg L<sup>-1</sup> and 100 mg L<sup>-1</sup>, 250 251 respectively. On the other hand, the higher solubility of H<sub>2</sub>S compared to that of CO<sub>2</sub> also mediated complete removals of this biogas contaminant regardless of the alkalinity 252 253 of the AC cultivation broth. These results were in accordance with Franco-Morgado et al. (2017), who reported values of H<sub>2</sub>S-REs of 99.5  $\pm$  0.5% throughout the operation of an 254 indoors HRAP interconnected to an AC using a highly carbonated medium at a pH of 9.5. 255 256 Likewise, Rodero et al. (2018) observed H<sub>2</sub>S-REs of 100.0  $\pm$  0.0, 94.7  $\pm$  1.9 and 80.3  $\pm$ 3.9% using a feeding nutrient solution with an average IC concentration of 1,500 mg  $L^{-1}$ , 257 500 mg  $L^{-1}$  and 100 mg  $L^{-1}$ , respectively. 258

259

The N<sub>2</sub> and O<sub>2</sub> concentration in the upgraded biogas increased from 2.4% at an IC concentration of  $42 \pm 1 \text{ mg L}^{-1}$  to 6.1% at 1,557 ± 26 mg IC L<sup>-1</sup> (Fig. 3). This increase was attributed to the enhanced N<sub>2</sub> and O<sub>2</sub> stripped out from the recycling cultivation broth mediated by the increase in medium salinity (which ultimately decreased the solubility of these gases).

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Finally, the lowest concentration of CH<sub>4</sub> in the upgraded biogas (88.0%) was recorded at 266 an IC concentration of  $42 \pm 1 \text{ mg L}^{-1}$ , increasing up to a maximum concentration of 93.2% 267 at  $634 \pm 48 \text{ mg L}^{-1}$  (Fig. 3). Interestingly, higher carbonate supplementation rates did not 268 result in an additional increase in the CH<sub>4</sub> content. The increased CH<sub>4</sub> concentration at 269 higher alkalinity loads was attributed to the limited desorption of N2 and O2 when 270 operating at the optimum L/G ratio and the high absorption efficiency of CO<sub>2</sub> and H<sub>2</sub>S 271 due to the acidic nature of these gases. Similar results were obtained by Rodero et al. 272 (2018), who reported CH<sub>4</sub> contents of  $98.9 \pm 0.2$ ,  $80.9 \pm 0.8$  and  $75.9 \pm 0.7\%$  at average 273

274	IC feed concentrations of 1,500, 500 and 100 mg L <sup>-1</sup> , respectively. Therefore, the results
275	herein obtained confirmed the key role of alkalinity on the methane quality.
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277	<figure 3=""></figure>
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279	3.3 Continuous photosynthetic biogas upgrading operation
280	The optimum operating parameters (i.e. L/G ratio of 0.5 and supplementation of a 16,000
281	mg IC $L^{-1}$ solution to the AC at a flowrate of 2.0 $L d^{-1}$ ) identified in sections 3.1 and 3.2
282	were selected to test the performance of the PBR during the continuous upgrading of raw
283	biogas coupled with the treatment of the mixed wastewater.
284	
285	3.3.1 Biogas upgrading
286	The composition of the methane obtained exhibited a rather constant value along the 42
287	days of operation (Fig. 4). $CO_2$ concentrations ranged between <0.1% and 1.4%,
288	corresponding to REs >91.0% (Fig. 4). The previous optimization of key operating
289	parameters such as the L/G ratio and the alkalinity in the cultivation broth of the AC
290	supported these consistent CO2 removals. Similarly, Marín et al. (2018a) reported values
291	of CO <sub>2</sub> concentration in the upgraded biogas ranging from 0.7 to 1.9% throughout the
292	operation of an outdoors HRAP interconnected to an external AC. It is important to
293	highlight that the CO <sub>2</sub> concentrations here obtained fulfilled most international
294	regulations for methane, which require $CO_2$ concentrations $\leq 2-4\%$ to be acceptable for
295	injection into natural gas grids (Muñoz et al., 2015). Moreover, no H <sub>2</sub> S was detected in
296	the methane during the whole experimental period regardless of the environmental
297	conditions, which agreed with the results previously observed during the optimization
298	assays. Therefore, the resulting methane also complied with the maximum H <sub>2</sub> S levels

enforced by international regulations for methane injection into natural gas grids (< 5 mg</li>
m<sup>-3</sup>) (Muñoz et al., 2015).

301

302 The N<sub>2</sub> and O<sub>2</sub> concentration in the upgraded biogas ranged from 0.9 to 5.9% throughout the entire operating period (Fig. 4), similar concentrations to those recorded by Marín et 303 al. (2018a), who reported N<sub>2</sub> and O<sub>2</sub> contents in the upgraded biogas between 0.5 and 304 305 6.3% during the operation of an outdoors HRAP interconnected to an AC. Likewise, Posadas et al. (2017) also recorded similar N<sub>2</sub> and O<sub>2</sub> concentrations of 1.4-6.1% in the 306 upgraded biogas. Unfortunately, these concentrations exceeded most of the time the 307 308 maximum quality requirements demanded for methane injection into natural gas grids of 309  $\leq$  1% (please note that the GC-TCD method did not allow to quantify separately O<sub>2</sub> and N<sub>2</sub>). O<sub>2</sub> is a hazardous biomethane contaminant based on its associated explosion risks, 310 311 while the presence of N<sub>2</sub> typically lowers the content of the biomethane. Therefore, a further optimization of the technology in order to avoid an active stripping of N<sub>2</sub> and O<sub>2</sub> 312 313 from the cultivation broth into the upgraded methane is still necessary.

314

315 Finally, high CH<sub>4</sub> concentrations in the upgraded biogas ranging from 94.1 to 98.9% were 316 recorded during this continuous assay (Fig. 4), likely due to the high CO<sub>2</sub>-REs, the complete elimination of H<sub>2</sub>S and the limited N<sub>2</sub> and O<sub>2</sub> desorption obtained under these 317 operating conditions. In this regard, the quality of the upgraded methane was similar or 318 319 even higher in terms of CH<sub>4</sub> content than that reported in previous studies. Indeed, Posadas et al. (2017) obtained CH<sub>4</sub> concentrations in the upgraded biogas of 87.0 - 93.0%, 320 while Marín et al. (2018a) achieved values up to 97.8% in a similar outdoors experimental 321 set-up (HRAP-AC) with a L/G ratio of 1.0 and IC concentrations in the cultivation broth 322 of ~ 2,600 mg IC  $L^{-1}$ . Finally, it should be highlighted that process performance here 323

324	recorded in this demo-scale PBR was superior to that recently recorded by Rodero et al.
325	(2019) in an outdoors 10 m <sup>3</sup> HRAP interconnected to an AC, where CH <sub>4</sub> concentrations
326	did not exceed 91% in the upgraded biogas. In this context, a minimum CH4
327	concentrations of $\geq$ 95% must be typically ensured prior injection of the methane into
328	natural gas grids in most international methane regulations (Muñoz et al., 2015).
329	
330	<figure 4=""></figure>
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332	3.3.2 Wastewater Treatment
333	Wastewater treatment performance in the PBR during biogas upgrading was evaluated in
334	terms of dissolved TOC, IC and TN removal (Fig. 5). Dissolved TOC concentrations in
335	the influent and effluent varied throughout the process with values ranging from 69.9 to
336	277.3 mg $L^{-1}$ in the influent and from 90.4 to 217.0 mg $L^{-1}$ in the effluent (Fig. 5). The
337	low TOC-REs recorded were attributed to the low biodegradability of the mixture of
338	agricultural and domestic wastewater used as influent to the PBR. Moreover, the
339	significant water evaporation rates from the cultivation broth in the open tanks and the
340	lysis of the microalgae generated during photosynthetic CO <sub>2</sub> fixation likely contributed
341	to increase the TOC concentration in the effluent in comparison to that of the influent,
342	thus resulting in the negative TOC-REs observed. On the other hand, the dissolved IC
343	concentration in the influent varied from 21.6 to 46.3 3 mg $L^{-1}$ and from 29.8 to 91.8 mg
344	$L^{-1}$ in the effluent (Fig. 5). Although no correlation between the IC concentration in the
345	effluent of the PBR and the addition of the carbonate solution in the AC was found due
346	to the high dilution effect associated to the large volume and short hydraulic retention
347	time of the PBR, the high values of pH in the PBR ranging between 7.9 and 8.9 might
348	have promoted the increase in the IC concentration of the effluent supported by biogas

349	absorption. Finally, no effective TN removal was observed during the entire experimental
350	period, with dissolved TN concentrations in the influent (ranging from 9.1 to 25.0 mg L <sup>-</sup>
351	<sup>1</sup> ) comparable to those recorded in the effluent (ranging from 11.1 to 25.9 mg L <sup>-1</sup> ) (Fig.
352	5).
353	
354	<figure 5=""></figure>
355	
356	4. Conclusions
357	This work constitutes, to the best of our knowledge, the first validation of photosynthetic
358	biogas upgrading in a pilot-scale semi-closed PBR interconnected to an AC under
359	outdoors conditions. Both the L/G ratio and the alkalinity in the AC were identified as
360	key parameters influencing the quality of the final methane, with optimum values of 0.5
361	and 634 $\pm$ 48 mg L <sup>-1</sup> , respectively. The implementation of the optimum operating
362	parameters during continuous operation resulted in a methane with CO <sub>2</sub> concentrations
363	of <0.1%-1.4%, $H_2S$ <0.5ppm <sub>v</sub> and $CH_4$ contents of 94.1-98.9%, which complied with
364	most international regulations for methane injection into natural gas grids.
365	
366	E-supplementary data of this work can be found in online version of the paper.
367	
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## 377 **REFERENCES**

- 378 (1) Andriani, D., Wresta, A., Atmaja, T.D., Saepudin, A., 2014. A Review on
- 379 Optimization Production and Upgrading Biogas Through CO2 Removal Using
- 380 Various Techniques. Appl. Biochem. Biotechnol. 172, 1909–1928.
- 381 doi:10.1007/s12010-013-0652-x
- 382 (2) APHA, 2005. Standard Methods for the Examination of Water and Wastewater,
- 383 21st ed. Public Health Association, Washington DC.
- 384 (3) Bahr, M., Díaz, I., Dominguez, A., González Sánchez, A., Muñoz, R., 2014.
- 385 Microalgal-biotechnology as a platform for an integral biogas upgrading and
- nutrient removal from anaerobic effluents. Environ. Sci. Technol. 48, 573–581.
- 387 doi:10.1021/es403596m
- 388 (4) European Biogas Association, 2017. EBA Statistical Report 2017 [WWW
- 389 Document]. URL http://european-biogas.eu/2017/12/14/eba-statistical-report-
- 390 2017-published-soon/ (accessed 11.13.18).
- 391 (5) Farooq, M., Almustapha, M.N., Imran, M., Saeed, M.A., Andresen, J.M., 2018. In-

392 situ regeneration of activated carbon with electric potential swing desorption

- 393 (EPSD) for the H2S removal from biogas. Bioresour. Technol. 249, 125–131.
- doi:10.1016/j.biortech.2017.09.198
- 395 (6) Franco-Morgado, M., Alcántara, C., Noyola, A., Muñoz, R., González-Sánchez,
- A., 2017. A study of photosynthetic biogas upgrading based on a high rate algal
- 397 pond under alkaline conditions: Influence of the illumination regime. Sci. Total
- 398 Environ. 592, 419–425. doi:10.1016/j.scitotenv.2017.03.077

- 399 (7) García, J., Ortiz, A., Álvarez, E., Belohlav, V., García-Galán, M.J., Díez-Montero,
- 400 R., Álvarez, J.A., Uggetti, E., 2018. Nutrient removal from agricultural run-off in
- 401 demonstrative full scale tubular photobioreactors for microalgae growth. Ecol.
- 402 Eng. 120, 513–521. doi:10.1016/j.ecoleng.2018.07.002
- 403 (8) Gutiérrez, R., Ferrer, I., González-Molina, A., Salvadó, H., García, J., Uggetti, E.,
- 404 2016. Microalgae recycling improves biomass recovery from wastewater treatment
- 405 high rate algal ponds. Water Res. 106, 539–549. doi:10.1016/j.watres.2016.10.039
- 406 (9) Marín, D., Posadas, E., Cano, P., Pérez, V., Blanco, S., Lebrero, R., 2018a.
- 407 Seasonal variation of biogas upgrading coupled with digestate treatment in an
- 408 outdoors pilot scale algal-bacterial photobioreactor. Bioresour. Technol. 263, 58–
- 409 66. doi:10.1016/j.biortech.2018.04.117
- 410 (10) Marín, D., Posadas, E., Cano, P., Pérez, V., Lebrero, R., Muñoz, R., 2018b.
- 411 Influence of the seasonal variation of environmental conditions on biogas
- 412 upgrading in an outdoors pilot scale high rate algal pond. Bioresour. Technol. 255,

413 354–358. doi:10.1016/j.biortech.2018.01.136

- 414 (11) Muñoz, R., Meier, L., Diaz, I., Jeison, D., 2015. A review on the state-of-the-art of
- 415 physical/chemical and biological technologies for biogas upgrading. Rev. Environ.
- 416 Sci. Biotechnol. 14, 727–759. doi:10.1007/s11157-015-9379-1
- 417 (12) Posadas, E., Marín, D., Blanco, S., Lebrero, R., Muñoz, R., 2017. Simultaneous
- 418 biogas upgrading and centrate treatment in an outdoors pilot scale high rate algal
- 419 pond. Bioresour. Technol. 232, 133–141. doi:10.1016/j.biortech.2017.01.071
- 420 (13) Posadas, E., Serejo, M.L., Blanco, S., Pérez, R., García-Encina, P.A., Muñoz, R.,
- 421 2015. Minimization of biomethane oxygen concentration during biogas upgrading
- 422 in algal-bacterial photobioreactors. Algal Res. 12, 221–229.
- 423 doi:10.1016/j.algal.2015.09.002

424	(14) Posadas, E., Szpak, D., Lombó, F., Domínguez, A., Díaz, I., Blanco, S., García-
425	Encina, P.A., Muñoz, R., 2016. Feasibility study of biogas upgrading coupled with
426	nutrient removal from anaerobic effluents using microalgae-based processes. J.
427	Appl. Phycol. 28, 2147–2157. doi:10.1007/s10811-015-0758-3
428	(15) Rodero, M. del R., Lebrero, R., Serrano, E., Lara, E., Arbib, Z., García-Encina,
429	P.A., Muñoz, R., 2019. Technology validation of photosynthetic biogas upgrading
430	in a semi-industrial scale algal-bacterial photobioreactor. Bioresour. Technol. 279,
431	43-49. doi:10.1016/j.biortech.2019.01.110
432	(16) Rodero, M. del R., Posadas, E., Toledo-Cervantes, A., Lebrero, R., Muñoz, R.,
433	2018. Influence of alkalinity and temperature on photosynthetic biogas upgrading
434	efficiency in high rate algal ponds. Algal Res. 33, 284–290.
435	doi:10.1016/j.algal.2018.06.001
436	(17) Ryckebosch, E., Drouillon, M., Vervaeren, H., 2011. Techniques for
437	transformation of biogas to biomethane. Biomass and Bioenergy 35, 1633–1645.
438	doi:10.1016/j.biombioe.2011.02.033
439	(18) Serejo, M.L., Posadas, E., Boncz, M.A., Blanco, S., García-Encina, P., Muñoz, R.,
440	2015. Influence of biogas flow rate on biomass composition during the
441	optimization of biogas upgrading in microalgal-bacterial processes. Environ. Sci.
442	Technol. 49, 3228–3236. doi:10.1021/es5056116
443	(19) Solorzano, L., 1969. Determination of ammonia in natural waters by the
444	phenolhypochlorite method. Limnol. Oceanogr. 799–801.
445	doi:10.4319/lo.1969.14.5.0799
446	(20) Sovechles, J.M., Waters, K.E., 2015. Effect of ionic strength on bubble coalescence
447	in inorganic salt and seawater solutions. AICHE 61, 2489–2496.
448	doi:doi.org/10.1002/aic.14851

- 449 (21) Toledo-cervantes, A., Estrada, J.M., Lebrero, R., Muñoz, R., 2017. A comparative
- 450 analysis of biogas upgrading technologies : Photosynthetic vs physical / chemical
- 451 processes. Algal Res. 25, 237–243. doi:10.1016/j.algal.2017.05.006
- 452 (22) Toledo-Cervantes, A., Morales, T., González, Á., Muñoz, R., Lebrero, R., 2018.
- 453 Long-term photosynthetic CO2 removal from biogas and flue-gas: Exploring the
- 454 potential of closed photobioreactors for high-value biomass production. Sci. Total
- 455 Environ. 640–641, 1272–1278. doi:10.1016/j.scitotenv.2018.05.270
- 456 (23) Toledo-Cervantes, A., Serejo, M.L., Blanco, S., Pérez, R., Lebrero, R., Muñoz, R.,
- 457 2016. Photosynthetic biogas upgrading to bio-methane: Boosting nutrient recovery
- 458 via biomass productivity control. Algal Res. 17, 46–52.
- doi:10.1016/j.algal.2016.04.017
- 460 (24) Uggetti, E., García, J., Álvarez, J.A., García-Galán, M.J., 2018. Start-up of a
- 461 microalgae-based treatment system within the biorefinery concept: From
- 462 wastewater to bioproducts. Water Sci. Technol. 78, 114–124.
- 463 doi:10.2166/wst.2018.195
- 464

### 465 FIGURE CAPTIONS

- 466 Figure 1. Schematic diagram of the experimental set-up used for the continuous467 photosynthetic upgrading of biogas.
- **Figure 2.** Concentration of CO<sub>2</sub> ( $\blacksquare$ ), N<sub>2</sub> + O<sub>2</sub> ( $\blacklozenge$ ) and CH<sub>4</sub> ( $\circ$ ) in the upgraded biogas at
- different L/G ratios.
- 470 **Figure 3.** Concentration of CO<sub>2</sub> ( $\blacksquare$ ), N<sub>2</sub> + O<sub>2</sub> ( $\blacklozenge$ ) and CH<sub>4</sub> ( $\circ$ ) in the upgraded biogas at
- 471 different IC concentration in the cultivation broth of the AC.
- 472 **Figure 4.** Time course of the concentration of CO<sub>2</sub> ( $\blacksquare$ ), N<sub>2</sub> + O<sub>2</sub> ( $\blacklozenge$ ) and CH<sub>4</sub> ( $\circ$ ) in the
- 473 upgraded biogas during continuous process operation.
- 474 Figure 5. Time course of the influent (solid symbols) and effluent (empty symbols)
- 475 concentrations of total organic carbon (squares), inorganic carbon (diamonds) and total
- 476 nitrogen (circles) throughout the continuous operation of the PBR.

Figure 1. Schematic diagram of the experimental set-up used for the continuous photosynthetic upgrading of biogas.





**Figure 2.** Concentration of CO<sub>2</sub> (**•**), N<sub>2</sub> + O<sub>2</sub> ( $\blacklozenge$ ) and CH<sub>4</sub> ( $\circ$ ) in the upgraded biogas at different L/G ratios.

**Figure 3.** Concentration of CO<sub>2</sub> (**•**),  $N_2 + O_2$  (**•**) and CH<sub>4</sub> ( $\circ$ ) in the upgraded biogas at different IC concentration in the cultivation broth of the AC.





Figure 4. Time course of the concentration of CO<sub>2</sub> (■), N<sub>2</sub> + O<sub>2</sub> (♦) and CH<sub>4</sub> (○) in the upgraded biogas during continuous process operation.

**Figure 5.** Time course of the influent (solid symbols) and effluent (empty symbols) concentration of total organic carbon (squares), inorganic carbon (diamonds) and total nitrogen (circles) throughout the continuous operation of the PBR



L/G ratio	Liquid flowrate (L d <sup>-1</sup> )	Biogas flowrate (L d <sup>-1</sup> )	Biogas HRT (h)
0.5	50	100	10.8
1.0	100	100	5.4
2.0	200	100	2.7
3.0	300	100	1.8
4.0	400	100	1.4
5.0	500	100	1.1

**Table 1.** Operational parameters during the evaluation of the influence of the L/G ratio in the AC.

1	Influence of liquid-to-biogas ratio and alkalinity on the biogas
2	upgrading performance in a demo scale algal-bacterial
3	photobioreactor
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# **1. Influence of alkalinity on the quality of the upgraded biogas**

- 19 Figure S1. Influence of the IC concentration in the cultivation broth of the AC on the
- 20 pH of the cultivation broth in the ( $\circ$ ) PBR and ( $\blacksquare$ ) AC.



Figure S2. Influence of the carbonate solution flowrate on the IC concentration in the
cultivation broth of the (•) PBR and (•) AC.



# 27 **2. Continuous Operation**

- **Figure S3.** Time course of the pH in the cultivation broth of the PBR during the
- 29 continuous biogas upgrading assay.



## 31 **3.** Temperature and biomass concentration

30

**Table S1.** Temperature and biomass concentration in the PBR during the evaluation of

the influence of the L/G ratio on the quality of the upgraded biogas.

L/G ratio	Temperature in the AC (°C)	Temperature in the PBR (°C)	TSS (g L <sup>-1</sup> )
0.5	29.9	27.4	0.11
1.0	30.3	28.0	0.15
2.0	33.4	37.2	0.20
3.0	33.5	36.9	0.45
4.0	32.3	35.0	0.37
5.0	32.4	22.1	0.27

- **Table S2.** Temperature and biomass concentration in the PBR during the evaluation of
- the influence of the alkalinity of the cultivation broth in AC on the quality of the

IC Concentration (mg L <sup>-1</sup> )	Temperature in the AC (°C)	Temperature in the PBR (°C)	TSS (g L <sup>-1</sup> )
42	31.7	28.4	0.40
311	32.5	27.4	0.43
634	31.3	29.4	0.49
996	35.0	32.8	0.19
1557	31.3	30.1	0.25

upgraded biogas.

37

**Table S3.** Temperature and biomass concentration in the PBR during the continuous

39 photosynthetic biogas upgrading experiment.

Day	Temperature in the AC (°C)	Temperature in the PBR (°C)	TSS (g L-1)
1	32.4	26.1	0.98
8	31.6	21.2	1.27
14	33.4	21.0	0.77
28	31.3	19.2	0.68
35	30.5	16.1	1.27
42	31.2	18.2	0.98