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**Temporal patterns of CO₂ and CH₄ in a rural area in northern Spain described by
a harmonic equation over 2010-2016**

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1 **Abstract**

2 The present paper seeks to improve our knowledge concerning the evolution of CO₂ and
3 CH₄ in terms of monthly trends, growth rate and seasonal variations in the lower
4 atmosphere. Dry continuous measurements of CO₂ and the CH₄ mixing ratio were
5 carried out over five and a half years (from 15 October 2010 to 29 February 2016) by
6 multi-point sampling at 1.8, 3.7 and 8.3 m, using a Picarro analyzer at a rural site in the
7 Low Atmosphere Research Centre (CIBA), on the upper Spanish plateau. Data were
8 divided into diurnal and nocturnal records. The mathematical equation proposed to
9 analyze the overall data was a harmonic one, comprising a polynomial (trend) and a
10 series of harmonics (seasonal cycle). Amplitude was considered as a constant and
11 variable term over time. Quite different behaviour was found between day and night
12 measurements in both climate forcing agents. CO₂ showed an accelerating trend in
13 autumn, whereas CH₄ trends were higher during the winter. Increasing growth rates
14 were reported for CO₂ and CH₄ over the whole study period. Nocturnal CO₂ amplitudes
15 are higher than diurnal ones except in winter for both gases, and also in the autumn for
16 CH₄.

17

18 **Key words:** upper Spanish plateau, greenhouse gases, series of harmonics, daytime,
19 night-time.

20

21 **1. Introduction**

22 Two of the major trace constituents warming the planet are carbon dioxide (CO₂) and
23 methane (CH₄), commonly known as “greenhouse gases” (GHGs). Any change in the
24 atmospheric amount of greenhouse gases results in the redistribution of the energy in
25 the atmosphere surface system and leads to a change in surface temperature,
26 contributing to global climate change, and posing one of the major challenges facing
27 humankind today (Haszpra et al., 2008). The contribution of radiative forcing from one
28 CH₄ molecule is 25 times higher than CO₂ (Fang et al., 2013b). However, mean
29 tropospheric lifetime is around 10 years for CH₄ and 30 for CO₂ (García et al., 2016).

30 Over the last ten years or so, fossil fuel combustion and industrial activities have been
31 the major CO₂ sources yielding 9.3 ± 0.5 GtC year⁻¹ (Global Carbon Project, 2016).
32 Respiration processes from plants and soils (Sreenivas et al., 2016) as well as land use
33 changes (Li et al., 2014) must also be considered. In contrast, the main CO₂ sinks for
34 the same period of time, have been terrestrial ecosystems (3.1 ± 0.9 GtC year⁻¹) and
35 oceans (2.6 ± 0.5 GtC year⁻¹) (Global Carbon Project, 2016).

36 As stated before, terrestrial ecosystems can act as important CO₂ sources or sinks. Plants
37 mainly take carbon out of the atmosphere through photosynthesis in summer, thereby
38 decreasing CO₂ concentrations, and release it in autumn and spring, thus increasing
39 concentrations (Barlow et al., 2015). The type of ecosystem plays a very important role
40 in this cycle, as well as, meteorological conditions (Fang et al., 2014).

41 Barlow et al. (2015) reported increases in the amplitude of seasonal exchange of CO₂
42 since the 1950s, particularly at mid-to high northern latitudes (Graven et al., 2013). CO₂
43 growth rates at different stations in the Northern Hemisphere have varied by around 1–3
44 ppm over the past decade (Huang et al., 2015).

45 As regards CH₄, it should be remembered that around 60-70% of emissions are from
46 anthropogenic sources (Fang et al., 2013a; Haszpra et al., 2008), with approximately
47 two thirds originating from the Northern Hemisphere as reported by Lelieveld (2006)
48 (Sánchez et al., 2014). Anthropogenic CH₄ sources include fermentation in livestock
49 and manure management (106 Tg year⁻¹), gas (79 Tg year⁻¹), landfills and waste water
50 treatment (59 Tg year⁻¹), coal mining (42 Tg year⁻¹) and biomass burning (30 Tg year⁻¹)
51 when incomplete combustion occurs (Global Carbon Project, 2016; Haszpra et al.,
52 2008; Sánchez et al., 2014). The main natural CH₄ sources are principally wetlands
53 (185 Tg year⁻¹), followed by peatlands, wild animals, digestion processes in termites,
54 microorganisms living in the oceans, forest fires, CH₄ carbohydrates and permafrost
55 (García et al., 2016; Global Carbon Project, 2016; Sánchez et al., 2014).

56 On the other hand, oxidation of this gas by OH radicals in the troposphere is the major
57 sink and accounts for 90% of CH₄ loss in the atmosphere (Fang et al., 2013a). Due to
58 the photochemical nature of OH radicals, reactions are particularly enhanced in summer
59 (Sánchez et al., 2014). Therefore, a maximum in the OH sink drives the summer
60 minima in CH₄ (Buchholz et al., 2016). Overlaid on the cycle induced by OH is the
61 seasonality of source contributions (Buchholz et al., 2016). Topp and Pattey (1997)
62 point to microbial uptake by soils and reactions in the stratosphere as other sinks
63 (Sánchez et al., 2014).

64 Global atmospheric CH₄ mole fractions went unnoticed from 1999 to 2006 (increasing
65 by 6 ppb) and then increased rapidly from 2007 to 2011 (29 ppb) (Fang et al., 2013a).
66 The reasons for the new increase are not yet perfectly understood (Fang et al., 2013a).
67 The insufficient number of monitoring stations, especially in rural areas, is one of the
68 reasons underlying this lack of understanding (Sánchez et al., 2005).

69 The main goal of the present study is to quantify the baseline concentration and
70 interpret the results regarding the ecosystem and atmospheric behaviour in terms of
71 annual cycle and inter-annual variability. To achieve this goal, a harmonic equation is
72 employed. The use of harmonic functions is not new in trace gases analyses. Nakazawa
73 et al. (1997) suggested that the trend of the data must be fitted by a polynomial term of
74 suitable degree, whereas the seasonal cycle may be described by a series of harmonic
75 terms. Anderson-Cook (2000) used a second harmonic model for cylindrical data
76 involving wind direction and temperature. Chamard et al. (2003) and Artuso et al.
77 (2009) employed a function consisting of an exponential term and two harmonics in line
78 with Fourier analyses. A fourth-order polynomial term with one harmonic was
79 developed by Inoue et al. (2006). Sánchez et al. (2008) employed one harmonic
80 function, considering the amplitude fixed and variable with time, to analyze the ozone
81 trend at the CIBA station for a six-year period. Sánchez et al. (2010) applied a model
82 comprising a linear polynomial term plus two harmonics, only one of which was
83 variable over time, expressing seasonality. Wu et al. (2012) and Liu et al. (2015) fitted
84 the data to a function with a linear polynomial and two harmonic functions considering
85 the variable term but in a simplified manner. Others, such as Fang et al. (2013b), have
86 used a quadratic polynomial plus the sum of four harmonics, without considering the
87 variable term.

88 In line with the previous literature, many authors have applied harmonic functions to
89 describe the behaviour of pollutants in the troposphere. However, researchers who have
90 employed four harmonics have not considered the amplitude variable over time for the
91 seasonal cycle, resulting in a worse fit of the data, since CO₂ and CH₄ concentrations
92 are time-dependent variables. Those who have developed a harmonic model in which
93 amplitude depends on time have not considered four harmonics, which involves a loss

94 of annual and interannual information in the cycle. For these reasons, this paper
95 develops a function taking into account both considerations. Recorded data were fitted
96 into a harmonic function, producing a prediction pattern for each instant that was the
97 sum of a linear trend based on a third-degree polynomial plus four harmonics, each
98 made up of a constant and a variable part along the time series. This enhancement
99 provides a better data fit and gives more accurate results.

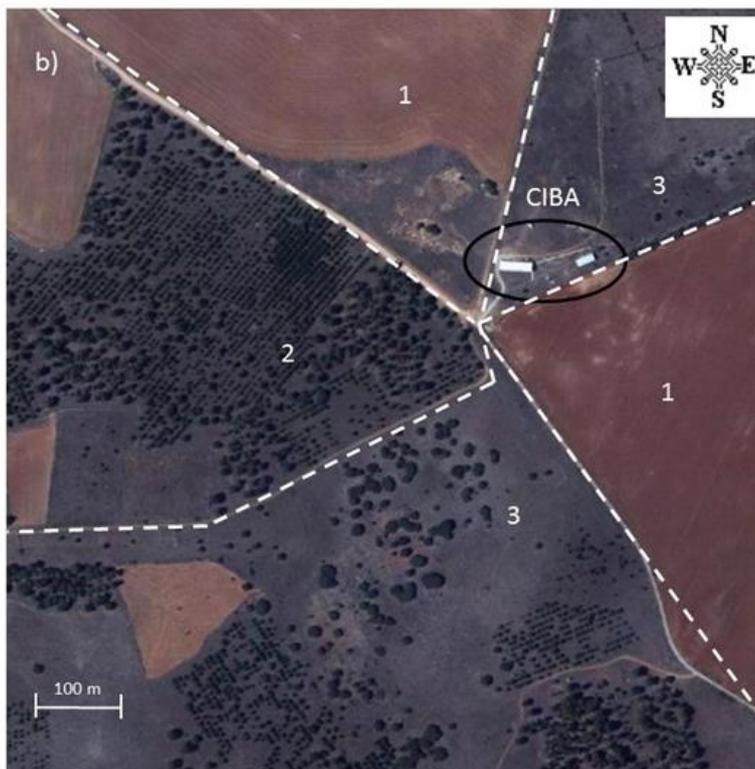
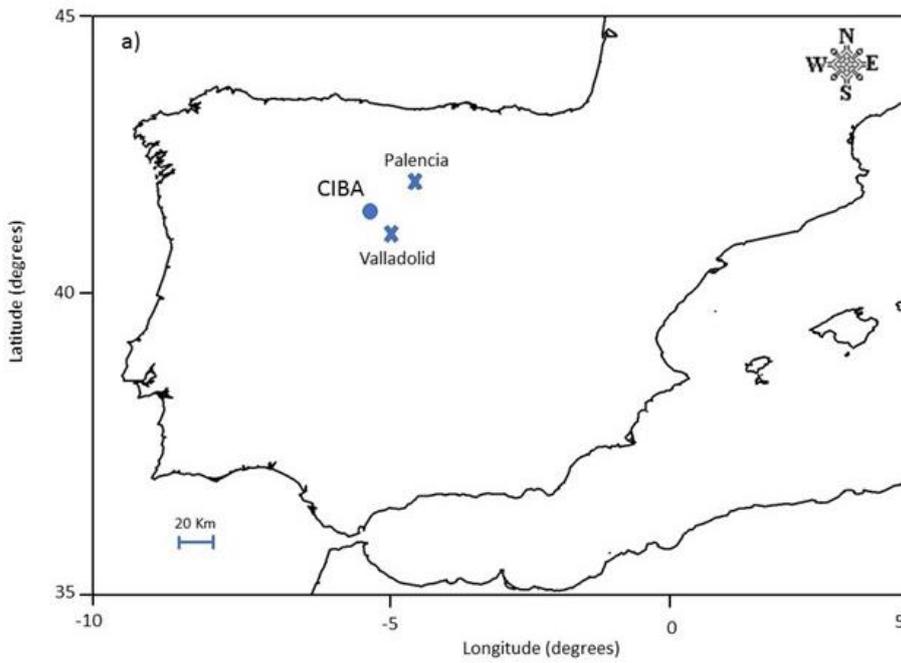
100 Previous studies on the measurement site have not considered the evolution of these
101 gases, but analyzed the distributions of the observations (Pérez et al., 2013; 2014). This
102 paper focuses on the differences between daytime and night-time, while other studies of
103 this research group examined total observations without differentiating between daytime
104 and night-time (Pérez et al., 2016; 2017). The findings of this study may improve
105 current knowledge concerning the evolution of CO₂ and CH₄ in a rural area of Southern
106 Europe over the last five years.

107

108 **2. Materials and methods**

109 *2.1 Site description*

110 Dry continuous atmospheric CO₂ and CH₄ measurements were carried out from 15
111 October 2010 to 29 February 2016 at the Low Atmosphere Research Centre (CIBA) in
112 the centre of the upper Spanish plateau (lat: 41°48'49"N, long: 4°55'59"W, alt: 845 m).
113 Figure 1 shows the location of the CIBA station and its surrounding area.



114

115 Fig.1. a) Location of the CIBA station on the upper Spanish plateau. b) PNOA image
 116 courtesy of © ign.es showing the land use around the sampling site, surrounded by a
 117 black line. Number 1 represents agricultural crops, 2 coniferous forest, and 3 a scrub-
 118 grassland association.

119 The monitoring station is located in a rural area (Fig. 1a) away from major industrial
120 sources. The closest largest cities are Valladolid (304,000 inhabitants) 24 km to the
121 southeast and Palencia (80,000 inhabitants) 40 km to the northeast. The surrounding
122 area (Fig. 1b) is characterized by non-irrigated crops and grass (Pérez et al., 2009a) and
123 a small area of coniferous forest. There are no relief elements, such that horizontal
124 homogeneity can be assumed (Pérez et al., 2009c).

125 The site is under a Mediterranean continental climate (García et al., 2012) with four
126 distinct seasons. Maximum temperatures are recorded in summer, July-August, and
127 minimum in winter, December-January (Sánchez et al., 2003). Annual precipitation in
128 the centre of the plateau is around 450 mm (Sánchez et al., 2014), with an evenly
129 distributed pattern throughout the year (García et al., 2012), peaking in autumn and
130 spring (Sánchez et al., 2014).

131

132 *2.2 Instrumentation*

133 A high precision Cavity Ring-Down Spectroscopy (CRDS) analyzer (G1301,
134 PICARRO Inc) is used for continuous atmospheric measurements at the CIBA station.
135 The response presents high linearity with concentrations and the signal is very stable
136 (Fang et al., 2014) and accurate and requires little maintenance (Sánchez et al., 2014). A
137 detailed discussion of this instrument and the necessary calibration has been addressed
138 by Buchholz et al. (2016) or Haszpra et al. (2008) among others.

139 The analyzer software includes a valve sequencer to automatically control external
140 solenoid valves to measure at 1.8, 3.7 and 8.3 m (García et al., 2016).

141 *2.3 Dataset*

142 Data collection is mostly complete except for sporadic gaps caused by instrument
143 malfunction (e.g., stream selection valves shifted to a wrong position), calibration,
144 maintenance or power failure. Boundary layer height values at the CIBA station were
145 obtained by METeoro logical data EXplorer web-based systems (Zeng et al., 2010) for
146 the whole study period. To gain a better understanding of the atmospheric patterns
147 between day and night the overall series has been divided taking the GMT hour from
148 the National Geographic Institute of Spain as a reference. Concentrations were analyzed
149 based on semi-hourly means.

150 The Fisher method was applied to the dataset in order to study the statistical differences
151 between levels. Using this procedure, if the intervals of a pair of samples do not overlap
152 in the vertical direction, their means could be considered different.

153 *2.4 Mathematical expression*

154 To evaluate the global evolution of CO₂ and CH₄ mole fractions, a harmonic equation
155 has been applied, which can simultaneously separate the long-term trend and the
156 seasonal cycle while retaining information about changes in amplitude.

157 The fitted curve used in this paper may thus be written as follows:

$$158 \quad y = \sum_{i=0}^3 a_i t^i + \sum_{j=1}^4 \sum_{k=0}^1 (b_{jk} t^k \cos(j2\pi t) + c_{jk} t^k \sin(j2\pi t)) \quad (1)$$

159 In the above expression, “y” represents the CO₂ or CH₄ mole fraction and time (*t*) is
160 expressed as a fraction of year based on the number of data between the start date (15
161 October 2010) and the end date (29 February 2016).

162 The unknown coefficients of the multiple harmonic regression were linearly estimated
163 running MATLAB[®] software. Independent variables were time (*t*, *t*², *t*³) and the series of
164 harmonics *t*^k *cos*(*j*2π*t*), *t*^k *sin*(*j*2π*t*).

165 The polynomial supplies information about the data trend. A third-degree polynomial
166 has been considered since Nakazawa et al. (1997) did not specify the suitable degree
167 and other authors such as Fang et al. (2013b) and Inoue et al. (2006) have adopted
168 similar orders.

169 The sequence of four harmonics (j) provides information about the yearly cycle. As the
170 main goal of this paper is to describe baseline concentrations of both gases in terms of
171 annual cycle and inter-annual changes, we need to employ four harmonics. The two first
172 harmonics refer to annual behaviour. The first ($j=1$), expresses annual behaviour and the
173 second ($j=2$) reinforces this information, sharpening the peaks and troughs. The third
174 harmonic ($j=3$) refers to four month changes and the fourth ($j=4$) to quarterly
175 information, that is say, both harmonics give the seasonal evolution.

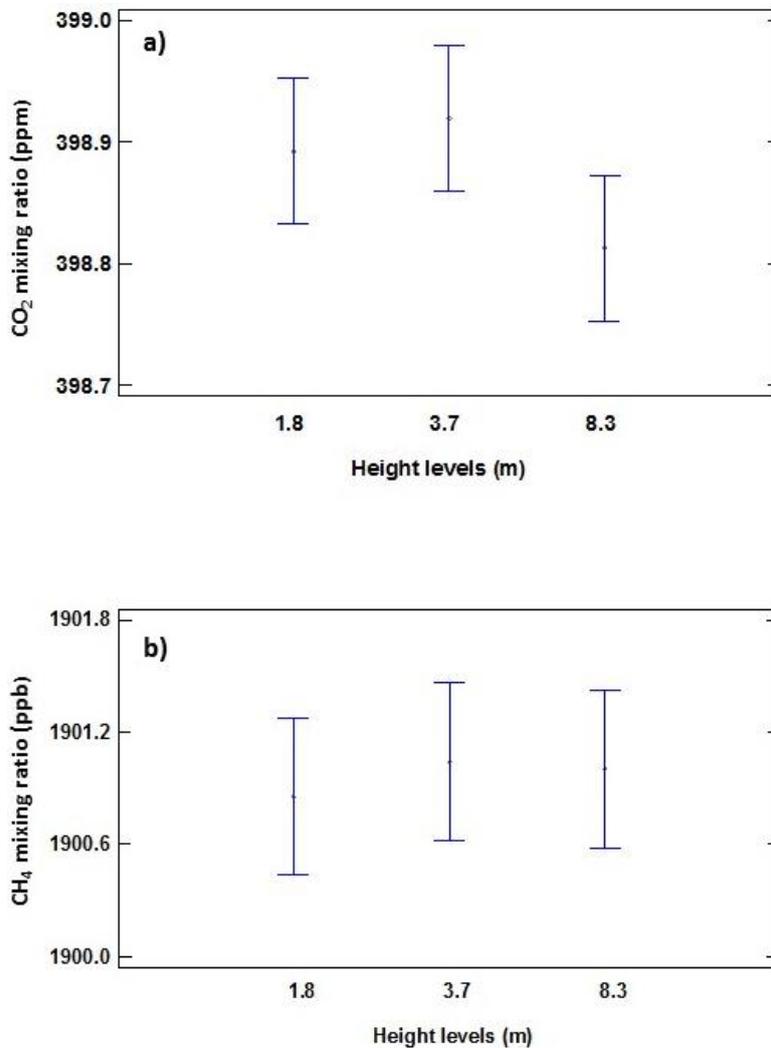
176 The main contribution of Eq (1) is to consider amplitude as a constant and a variable
177 term along the time series. As mentioned in the Introduction section, to the best of the
178 author's knowledge, the variable term has not been widely studied. Contributions of
179 other authors including Inoue et al. (2006) or Wu et al. (2012) have considered the
180 variable term but in a simple way. We strongly believe that a further evaluation of its
181 influence in the seasonal cycle is required. The term in which the amplitude is constant
182 with time represents the general features of the seasonal cycle. According to Sánchez et
183 al. (2010), this term explains around 87% of the seasonal cycle. However, the term in
184 which the amplitude is variable along the time series reinforces the information of the
185 fixed term, offering more accurate results and endowing the data with greater flexibility.
186 The index k reflects the term in which amplitude is constant with time ($k=0$) and the
187 other in which it is variable ($k=1$) over time.

188

189 **3. Results and discussion**

190 *3.1. Measures at the three levels*

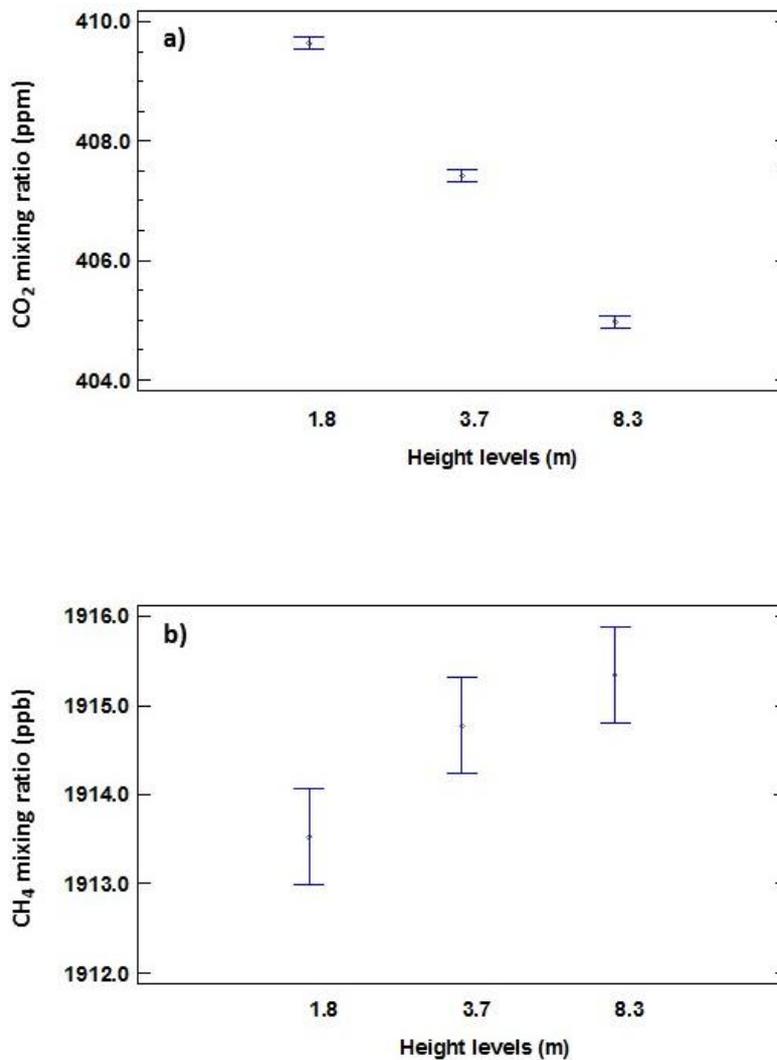
191 We found no statistical differences between the three levels during daytime (Fig. 2) in
192 line with Pérez et al. (2012a), due to thermal turbulence in the lower atmosphere, which
193 causes an intensive vertical mixture as suggested by said author. Figure 2 reflects the
194 least significant difference between samples in accordance with the Fisher method with
195 a confidence interval of 95%.



196

197 Fig. 2. Least significant difference (LSD) intervals for CO₂ and CH₄ mixing ratio during
198 daytime for the three levels measured: a) CO₂ results. b) CH₄ results.

199 Controversially, we found small differences in the nocturnal data (Fig. 3) attributed to
200 vegetation respiration intensified by the stability increase, as suggested by Pérez et al.
201 (2012a). As the height of the measurement increases, the CO₂ concentration decreases
202 (Fig. 3a) as reported by other authors such as Fang et al. (2013a; 2014) and Lee et al.
203 (2012). This decrease may be due to the lower dispersion linked to lower wind speed as
204 reported by Pérez et al. (2012a). With regard to CH₄, concentrations grow as height
205 increases (Fig.3b), in contrast to certain previous studies such as Fang et al. (2013a;
206 2014) and Lee et al. (2012), although in agreement with the conclusions of García et al.
207 (2016).



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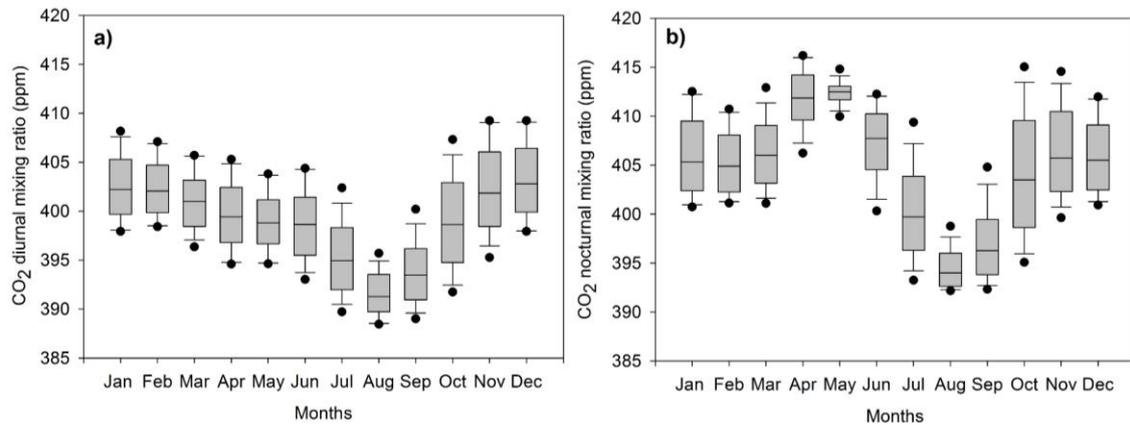
209 Fig. 3. Least significant difference (LSD) intervals for CO₂ and CH₄ mixing ratio at
210 night-time for the three levels measured: a) CO₂ results. b) CH₄ results.

211 No strong differences could be inferred between the three height levels (Figs. 2 and 3).
212 The highest (8.3 m) was thus used to describe the CO₂ and CH₄ cycle in the following
213 sections.

214 3.2. *Global evolution*

215 Global evolution of CO₂ and CH₄ have been obtained from Eq (1). The annual average
216 concentration at the CIBA station from 2010 to 2016 was 399.24 ± 5.05 ppm for CO₂
217 and 1906.63 ± 25.70 ppb for CH₄ during daytime, and 404.83 ± 6.28 ppm (CO₂) and
218 1913.27 ± 24.32 ppb (CH₄) during the night-time. These values were consistent with
219 earlier research in the same study area reported by García et al. (2008). The mean
220 concentrations found for CO₂ closely follow the global values, 394.90 ppm, considering
221 monthly data (NOAA, 2016). CH₄ concentrations at CIBA are higher than global mean
222 values, which represent 1816.49 ppb (NOAA, 2016).

223 Figure 4 depicts the variation of regional CO₂ concentration recorded at the CIBA
224 station. Each box represents the interquartile range where the median is represented by a
225 line. Whiskers extend from the 10th to the 90th percentile. Outliers correspond to the 5th
226 and 95th percentiles.



227

228 Fig. 4. Annual evolution of CO₂ measurements over the five and a half years of study at
 229 the CIBA station: a) daytime data. b) night-time data.

230 Several points should be noted from Fig. 4. First, it has been evidenced that time of year
 231 affects gas concentrations, and second, CO₂ concentrations during the day (Fig. 4a) are
 232 higher in winter (maximum in December with 403.17 ppm) and lower in summer
 233 (August 391.28 ppm). These results are supported by García et al. (2012). A decreasing
 234 pattern from February to August followed by an increase in mean values should be
 235 noted.

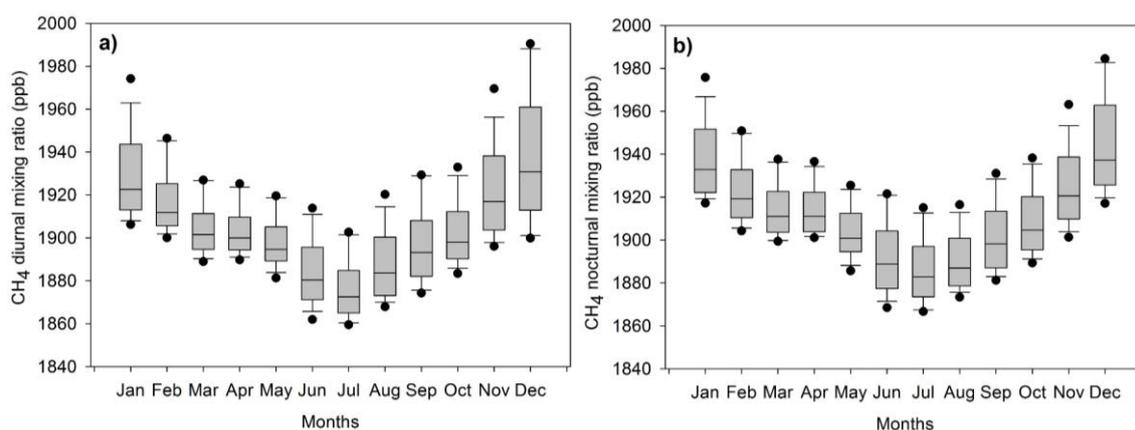
236 Analyzing CO₂ nocturnal records (Fig. 4b), we detected two peaks as did other
 237 researchers such as Sánchez et al. (2010). Sánchez et al. (2010) attributed these
 238 temporal variations to strong temperature inversions at the rural site which contributed
 239 to trapping CO₂ during night-time especially during the growing season and in autumn
 240 (García et al., 2012).

241 The first peak is located in spring (April: 411.72 ppm), corresponding to the period of
 242 maximum vegetation growth (Sánchez et al., 2005). In that period, the moisture content
 243 of soils and their respiration process play an important role as vegetation becomes
 244 photosynthetically active (Lee et al., 2012). Pérez et al. (2009a) have documented the
 245 influence of the Valladolid plume over the site in spring, increasing gas concentration.

246 A decrease in concentration can be observed from May to August. The same behaviour
247 has been explained by García et al. (2008), Pérez et al. (2009a) and Sánchez et al.
248 (2003; 2005). This decrease might be attributed to the lack of vegetation (the harvest
249 season), higher temperatures and low soil moisture, factors that may diminish the
250 respiration process (García et al., 2010). Lower anthropogenic emissions during the
251 summer due to the absence of heating and the reduction in traffic and industrial
252 activities must also be considered (García et al., 2012). The second maximum occurs in
253 autumn (November, 405.74 ppm) as reported by Sánchez et al. (2003). These authors
254 attribute this fact to local soil disturbances caused by ploughing the soil while preparing
255 the land for seeding and to an increase in the amount of precipitations, as also pointed
256 out by García et al. (2008). According to Sun et al. (2014), this peak seems to be linked
257 to increased ecosystem productivity as well as soil microbial activity and respiration
258 processes (Kirschke et al., 2013). Moreover, at this time of the year the Palencia plume
259 has an impact on the site (Pérez et al., 2009a), although its emissions are four times
260 lower than Valladolid emissions (Pérez et al., 2009b). Due to the smaller population of
261 Palencia and the greater distance from the CIBA station, its plume has less of an impact
262 on the measuring point than the Valladolid plume (Pérez et al., 2009b; 2012a).

263 The nocturnal CO₂ minimum is also located in summer (August 394.01 ppm) as occurs
264 during the daytime, indicating the strong absorption by land biosphere in the middle of
265 the year (Fang et al., 2016). These results are in agreement with the CO₂ cycle in the
266 Northern Hemisphere, which is mainly dominated by photosynthetic activity and which
267 evidences rapid decreases from June to August and large increments from September to
268 December (Fang et al., 2014). Transport processes during night-time described by Pérez
269 et al. (2012b) showed low concentrations of CO₂ associated with the NE, WSW and
270 SSW wind prevalence.

271 The CH₄ cycle is simpler, showing peaks in winter and troughs in summer (Fig. 5). This
 272 pattern is in line with other research in the Northern Hemisphere. Some examples
 273 include the conclusions of Fang et al. (2015) and García et al. (2016). The box plots in
 274 Fig. 5 illustrate the annual CH₄ cycle. Outliers, as reported by Sánchez et al. (2014),
 275 may be attributed to high emissions from Valladolid and Palencia and to the impact of
 276 fugitive emissions from a landfill, located near the monitoring station, which releases
 277 around 7.11 kt year⁻¹ into the atmosphere. The local influence of livestock in the region
 278 must also be mentioned. This makes it a major source of CH₄ emissions with
 279 approximately 197 kt year⁻¹. In line with García et al. (2016), from a long range
 280 standpoint, the main pollutant sources come from European trajectories, specifically, in
 281 our case from the southeast sector where the urban landfill is located.



282
 283 Fig. 5. Annual CH₄ mixing ratio at the CIBA station over the whole study period: a)
 284 diurnal measurements. b) nocturnal measurements.

285 The highest mole fractions occur in December, average 1937.86 ppb during the daytime
 286 (Fig. 5a) and 1944.36 ppb during the night-time (Fig. 5b). The low presence of radicals
 287 during winter leads to increased CH₄ in this season (Fang et al., 2013a). At the same
 288 time, biomass burning begins around October in the upper Spanish plateau and might
 289 influence the cycle as Fang et al. (2013a) has suggested for other sites.

290 The minimum concentration was found in July and was 1877.22 ppb for daytime data
291 and 1886.74 ppb for the nocturnal dataset. The main reason for the minimum point
292 results from the higher temperatures, which dry the croplands, thereby reducing plant
293 activity and increasing CH₄ oxidation, which is considered the principal removal
294 process (García et al., 2016).

295 Lower mixing ratios of both greenhouse gases in summertime have been detected. Al-
296 Anzi et al. (2016) explain this behaviour because of the higher planetary boundary layer
297 that reinforces pollutant transport and the dispersion process. During this season, solar
298 radiation is also higher, producing good dilution from surface and thus a decrease in
299 CO₂ and CH₄ concentrations (García et al., 2012; Pérez et al., 2009b). The opposite is
300 found for the winter season. Moreover, Atlantic air masses (cleaner masses) are more
301 frequent in summer as are continental air masses (more pollutants) in spring (García et
302 al., 2010).

303 The CIBA station reveals sharp differences between day and night (Figs. 4 and 5).
304 Higher concentrations at night and lower during the day have been detected in
305 agreement with Martins et al. (2016) and Northern Hemisphere mean values (Pu et al.,
306 2014). These diurnal changes are most marked in the growing season (Pérez et al.,
307 2009c). Newman et al. (2013) found that diurnal variations in CO₂ and CH₄ were
308 mostly influenced by the planetary boundary layer through convection (Sreenivas et al.,
309 2016).

310 During the daytime, at the CIBA station, the boundary layer extends to a mean of 550
311 metres (METEX, 2017). Its maximum height in the middle of the day and the stronger
312 vertical mixing (Fang et al., 2014), produce more dilution in the air (García et al., 2016).

313 CO₂ uptake and higher photochemical reactions are also considered important causes of
314 the low CO₂ and CH₄ concentrations (Lee et al., 2012).

315 At night, radiation loss at ground level leads to a shallow stable boundary layer (Fang et
316 al., 2013a) reaching 190 metres at CIBA (METEX, 2017). Atmospheric mixing is low
317 (Haszpra et al., 2008) and turbulent processes decrease. For these reasons, dispersion
318 confines emissions in a narrow mixing layer (García et al., 2012) contributing to trap
319 surface emissions (Fang et al., 2013a). Thermal inversions which are greater at night
320 (Pérez et al., 2009b) and the respiration process also impact on higher concentrations at
321 night (Lee et al., 2012). As regards CH₄, higher concentrations might be explained by
322 the prevalence of easterly and northerly winds in the area at night. Bearing in mind that
323 Palencia is in the northeast, Valladolid in the southeast and the urban landfill in the SSE
324 sector, the highest CH₄ concentrations during night-time are in consonance with the
325 main CH₄ sources in the surrounding area (García et al., 2008; 2016; Sánchez et al.,
326 2014).

327 *3.3. Monthly trend analysis*

328 To study the trend, datasets were deseasonalized in order to analyze only the polynomial
329 term of Eq (1). Monthly mean trend values were then calculated and used to apply
330 simple linear regressions (mixing ratio-time), obtaining the monthly trend equations for
331 CO₂ and CH₄ (Tables 1 and 2). Positive and almost linear trends for CH₄ and CO₂ were
332 found, as reported previously by other researchers such as Buchholz et al. (2016).

333 (Table 1)

334 The outcomes from Table 1 are in consonance with those given by ul-Haq et al. (2017),
335 showing slopes around 2 ppm. The increasing trend is greater in autumn than in spring.

336 Piao et al. (2008) linked this to higher autumn temperatures caused by climate change,
337 which leads to increased respiration rates (Zhang and Zhou., 2013).

338 (Table 2)

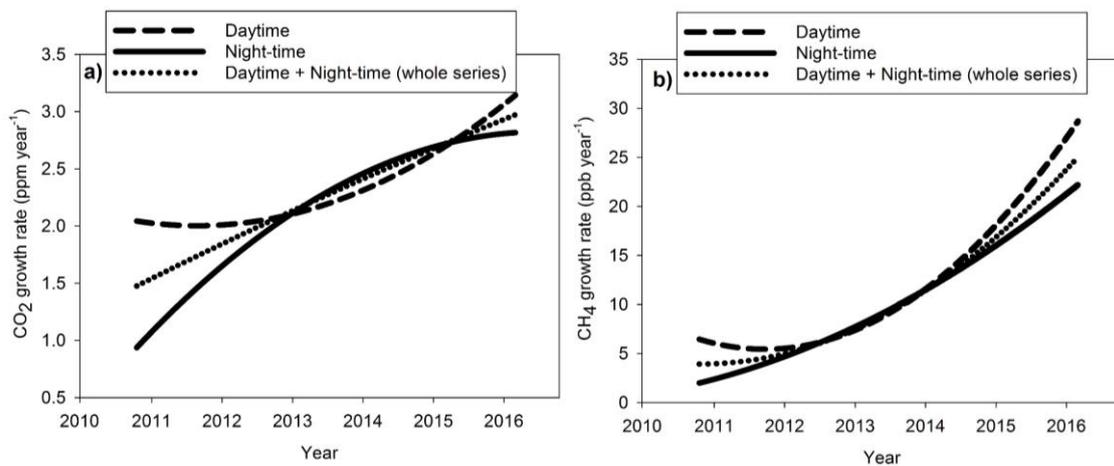
339 In the case of CH₄ (Table 2), anthropogenic causes have an impact on autumn and
340 winter emissions.

341 The satisfactory correlation of the linear fit proves the goodness of the equation. This
342 adjustment was better for CO₂ results. In both cases R² was above 0.96.

343 3.4. Growth rate

344 According to Barlow et al. (2015), periods of more than a year and a half are good
345 indicators of the growth rate. In this case, a database of five and a half years has been
346 studied.

347 Figure 6 shows the growth rates of CO₂ and CH₄ which were calculated by the first
348 derivative of trend curves from Eq (1), as other authors, such as Fang et al. (2016), have
349 done previously.



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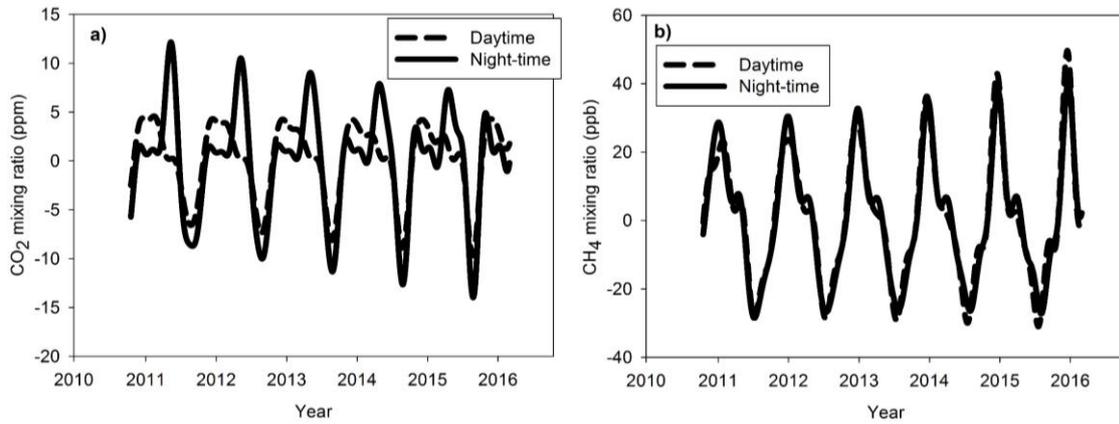
351 Fig. 6. Growth rates. The dotted line represents the whole data series, the dashed line
352 corresponds to daytime observations, and the continuous line to night-time data a) CO₂
353 results. b) CH₄ results.

354 The previous CO₂ results yielded a positive linear increase of 2.32 ppm year⁻¹ at the
355 station during daytime conditions, and 2.15 ppm year⁻¹ during night-time; with a net
356 growth rate (whole series) of 2.25 ppm year⁻¹ (Fig. 6a), slightly higher than the global
357 values of 2.06 ppm year⁻¹ for the past decade, (WMO Greenhouse Gas Bulletin, 2015),
358 but lower than those presented by Sánchez et al. (2010) at the same location. These
359 results concur with others shown by García et al. (2010) and Haszpra et al. (2008). The
360 positive CO₂ diurnal growth rate might be derived from vehicle emissions and industrial
361 activities of the nearest cities. The nocturnal increment could be partially attributed to a
362 rise in global mean temperatures over the last five years. This increase may anticipate
363 the beginning of autumn, increasing the respiration rate at nights.

364 For CH₄ (Fig. 6b), a mean growth rate of 11.90 ppb year⁻¹ for diurnal data, and 10.32
365 ppb year⁻¹ for nocturnal outputs has been detected, meaning a net growth rate of 10.95
366 ppb year⁻¹ considering all the data series. These results double Mauna Loa's results (4.7
367 ppb year⁻¹ for the last decade), but are in agreement with those presented by Fang et al.
368 (2016) and Sánchez et al. (2014). The latter author suggested that the contribution of the
369 large amount of livestock in the region may be one reason why the mean growth rate is
370 greater at the CIBA station compared to some non-disturbed areas. Both, daytime and
371 night-time series seem to be mainly influenced by the increase in local anthropogenic
372 emissions from industrial activities, fuel burning and emissions from the urban landfill.

373 *3.5. Seasonal cycle*

374 In this section, data were detrended (harmonic data minus polynomial data), working
375 only with the series of harmonics from Eq (1). These are depicted in Fig. 7.



376

377 Fig. 7. Seasonal cycle: a) CO₂ results. b) CH₄ results.

378 The difference between the maximum and minimum value (peak-to-peak) is defined as
379 the amplitude of a cycle. This parameter may be considered as a biological activity
380 indicator (Barlow et al., 2015), but is also influenced by local sources/sinks (Li et al.,
381 2014).

382 CO₂ amplitudes (Fig.7a) were 1.84, 4.59, 6.13 and 2.45 ppm for spring, summer,
383 autumn, and winter, respectively during the day. When analyzing nocturnal series, the
384 results were 7.36, 8.91, 9.97 and 1.59 ppm, respectively. Amplitudes are higher for
385 nocturnal data except in winter due to the greater emissions from heating systems and
386 the lower assimilation rates during cold months.

387 The highest amplitudes were found in autumn, which could be representative of the
388 increase in rainfall during this season as suggested by Buermann et al. (2007). This
389 maximum seems to be in accordance with Lee et al. (2012). The lowest variability was
390 found in spring (daytime) and in winter (night-time). Huang et al. (2015) and Sánchez et

391 al. (2005) have also found lower amplitudes in winter, linking this to the weakest
392 respiratory process by vegetation and roots.

393 An increase in the amplitude of the CO₂ autumnal maximum over night-time data
394 collection was also apparent. These results disagree with those found by Sánchez et al.
395 (2014). Since the amplitude is linked to biological processes (Barlow et al., 2015), this
396 increment could be attributed to an increase in temperatures, stimulating vegetation
397 growth. This leads to an increase in CO₂ emissions through respiration process. This
398 peak occurred earlier over the whole study period in line with the results found by
399 Barlow et al. (2015). The extended growing season caused by higher temperatures in
400 recent years might explain this behaviour in the sampling area. Further studies are
401 needed to confirm such a hypothesis.

402 Seasonal amplitudes for CH₄ (Fig.7b) were 12.90, 12.17, 21.97 and 31.64 ppb during
403 daytime (spring, summer, autumn and winter). For the nocturnal series, results were
404 24.69, 12.69, 17.51 and 22.90 ppb. The highest values in winter (daytime) might be
405 explained by biomass burning. During night-time, maximum values appear in spring
406 and winter, which could be attributed to the strong temperature inversions in these
407 seasons at the CIBA station. The lowest amplitudes occur in summer in both scenarios,
408 which may be associated to photochemical reactions.

409 An increasing trend in seasonal amplitude has been identified in agreement with other
410 observations in the Northern Hemisphere (Sánchez et al., 2010). The annual amplitude
411 growth rate was linear for diurnal values, being 0.7373 ppm ($R^2= 0.9995$) for CO₂ and
412 5.9 ppb ($R^2= 0.9474$) for CH₄.

413

414 **4. Conclusions**

415 CO₂ and CH₄ behaviour in the low atmosphere was analyzed in this paper and showed a
416 seasonal pattern from 2010 to 2016. Strong differences were found between day and
417 night-time measurements, with lower values being in evidence during the daytime. The
418 CO₂ cycle revealed sharper differences between day and night compared with the CH₄
419 cycle. During the day, a maximum occurs for CO₂ in December (403.17 ppm) and a
420 minimum in August (391.28 ppm), whereas at night two maximums were observed; the
421 first in April (411.72 ppm) and the second in November (405.74 ppm). The lowest point
422 was also located in August (394.01 ppm). CH₄ peaks were found in December for both
423 scenarios, reaching 1937.86 ppb in daytime conditions, and 1944.36 ppb for night-time.
424 Troughs were found in July, and reached 1877.22 ppb during the daytime and 1886.74
425 ppb during the night. These differences were the result of many factors mainly related
426 with biosphere and biological ecosystem changes (soil humidity, respiration process,
427 amount of precipitations), anthropogenic local sources (plumes, the urban landfill and
428 surrounding agricultural activities) and atmospheric patterns (height of the planetary
429 boundary layer, the prevalence of southeast winds or turbulent processes). Monthly
430 trends led to very stable slopes (around 2 ppm year⁻¹ for CO₂ and 10 ppb year⁻¹ for CH₄)
431 showing no evidence of stronger effects in certain months. Growth rates were far more
432 accelerated, with higher values than those found at Mauna Loa. Differences were
433 greater during the day (2.32 ppm year⁻¹ for CO₂ and 11.90 ppb year⁻¹ for CH₄) than
434 during the night (2.15 ppm year⁻¹ for CO₂ and 10.32 ppb year⁻¹ for CH₄). Seasonal
435 variations due to ecosystem activity and atmospheric behaviour were greater in autumn
436 for CO₂ observations (6.13 ppm for daytime and 9.97 ppm for night-time
437 measurements), in contrast with the CH₄ results that found this maximum during the
438 winter season (31.64 ppb) for the daytime measurements and in spring (24.69 ppb) for
439 the nocturnal data recorded. With all this information, the effectiveness of the harmonic

440 equation for evaluating temporal trends and seasonal cycles of CO₂ and CH₄ employing
441 a large dataset from a rural area in Southern Europe was demonstrated.

442 The results obtained might prove useful to understand the processes that govern CO₂
443 and CH₄ cycles and their future trend, which is essential for designing effective air
444 pollution control and dealing with climate change.

445 **Conflict of interest**

446 The authors declare that there is no conflict of interest regarding publication of this
447 paper.

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452 **References**

- 453 Al-Anzi, B., Abusam, A., Khan, A.R., 2016. Evaluation of temporal variations in
454 ambient air quality at Jahra using multivariate techniques. *Environmental Technology*
455 *and Innovation*. 5, 225-232.
- 456 Anderson-Cook, C.M., 2000. A second order model for cylindrical data. *Journal of*
457 *Statistical Computation and Simulation*. 66, 51-65.
- 458 Artuso, F., Chamard, P., Piacentino, S., Sferlazzo, D.M., De Silvestri, L., di Sarra, A.,
459 Meloni, D., Monteleone, F., 2009. Influence of transport and trends in atmospheric CO₂
460 at Lampedusa. *Atmospheric Environment*. 43, 3044-3051.
- 461 Barlow, J.M., Palmer, P.I., Bruhwiler, L.M., Tans, P., 2015. Analysis of CO₂ mole

462 fraction data: first evidence of large-scale changes in CO₂ uptake at high northern
463 latitudes. *Atmospheric Chemistry and Physics*. 15, 13739-13758.

464 Buchholz, R.R., Paton-Walsh, C., Griffith, D.W.T., Kubistin, D., Caldow, C., Fisher, J.
465 A., Deutscher, N.M., Kettlewell, G., Riggenbach, M., Macatangay, R., Krummel, P.B.,
466 Langenfelds, R.L., 2016. Source and meteorological influences on air quality (CO, CH₄
467 and CO₂) at a Southern Hemisphere urban site. *Atmospheric Environment*. 126, 274-
468 289.

469 Buermann, W., Lintner, B.R., Koven, C.D., Angert, A., Pinzon, J.E., Tucker, C.J., Fung,
470 I.Y., 2007. The changing carbon cycle at Mauna Loa Observatory. *Proceedings of the*
471 *National Academy of Sciences of the United States of America*. 104, 4249-4254.

472 Chamard, P., Thiery, F., di Sarra, A., Ciattaglia, L., De Silvestri, L., Grigioni, P.,
473 Monteleone, F., Piacentino, S., 2003. Interannual variability of atmospheric CO₂ in the
474 Mediterranean: measurements at the island of Lampedusa. *Tellus, Series B: Chemical*
475 *and Physical Meteorology*. 55, 83-93.

476 Fang, S.X., Zhou, L.X., Masarie, K.A., Xu, L., Rella, C.W., 2013a. Study of
477 atmospheric CH₄ mole fractions at three WMO/GAW stations in China. *Journal of*
478 *Geophysical Research: Atmospheres*. 118, 4874-4886.

479 Fang, Z., Lingxi, Z., Lin, X., 2013b. Temporal variation of atmospheric CH₄ and the
480 potential source regions at Waliguan, China. *Science China. Earth Sciences*. 56, 727-
481 736.

482 Fang, S.X., Zhou, L.X., Tans, P.P., Ciais, P., Steinbacher, M., Xu, L., Luan, T., 2014. In
483 situ measurement of atmospheric CO₂ at the four WMO/GAW stations in China.
484 *Atmospheric Chemistry and Physics*. 14, 2541-2554.

485 Fang, S.X., Tans, P.P., Steinbacher, M., Zhou, L.X., Luan, T., 2015. Comparison of the
486 regional CO₂ mole fraction filtering approaches at a WMO/GAW regional station in
487 China. *Atmospheric Measurement Techniques*. 8, 5301-5313.

488 Fang, S.X., Tans, P.P., Dong, F., Zhou, H., Luan, T., 2016. Characteristics of
489 atmospheric CO₂ and CH₄ at the Shangdianzi regional background station in China.
490 *Atmospheric Environment*. 131, 1-8.

491 García, M.Á., Sánchez, M.L., Pérez, I.A., De Torre, B., 2008. Continuous carbon
492 dioxide measurements in a rural area in the upper Spanish plateau. *Journal of the Air
493 and Waste Management Association*. 58 (7), 940-946.

494 García, M.Á., Sánchez, M.L., Pérez, I.A., 2010. Synoptic weather patterns associated
495 with carbon dioxide levels in Northern Spain. *Science of the Total Environment*. 408,
496 3411-3417.

497 García, M.Á., Sánchez, M.L., Pérez, I.A., 2012. Differences between carbon dioxide
498 levels over suburban and rural sites in Northern Spain. *Environmental Science and
499 Pollution Research*. 19, 432-439.

500 García, M.Á., Sánchez, M.L., Pérez, I.A., Ozores, M.I., Pardo, N., 2016. Influence of
501 atmospheric stability and transport on CH₄ concentrations in northern Spain. *Science of
502 the Total Environment*. 550, 157-166.

503 Global Carbon Project. 2016. Carbon Budget 2016.
504 <http://www.globalcarbonproject.org/carbonbudget/index.htm> (accessed 2.16.17).

505 Global Carbon Project. 2016. Methane Budget 2016.
506 <http://www.globalcarbonproject.org/methanebudget/index.htm> (accessed 2.16.17).

507 Graven, H.D., Keeling, R.F., Piper, S.C., Patra, P.K., Stephens, B.B., Wofsy, S.C.,

508 Welp, L.R., Sweeney, C., Tans, P.P., Kelley, J.J., Daube, B.C., Kort, E.A., Santoni,
509 G.W., Bent, J.D., 2013. Enhanced seasonal exchange of CO₂ by Northern ecosystems
510 since 1960. *Science*. 341, 1085-1089.

511 Haszpra, L., Barcza, Z., Hidy, D., Szilágyi, I., Dlugokencky, E., Tans, P., 2008. Trends
512 and temporal variations of major greenhouse gases at a rural site in central Europe.
513 *Atmospheric Environment*. 42, 8707-8716.

514 Huang, X., Wang, T., Talbot, R., Xie, M., Mao, H., Li, S., Zhuang, B., Yang, X., Fu, C.,
515 Zhu, C., Zhu, J., Huang, X., Xu, R., 2015. Temporal characteristics of atmospheric CO₂
516 in urban Nanjing, China. *Atmospheric Research*. 153, 437-450.

517 Inoue, H.Y., Matsueda, H., Igarashi, Y., Sawa, Y., Wada, A., Nemoto, K., Sartorius, H.,
518 Schlosser, C., 2006. Seasonal and Long-Term Variations in Atmospheric CO₂ and ⁸⁵Kr
519 in Tsukuba, Central Japan. *Journal of the Meteorological Society of Japan*. 84(6), 959-
520 968.

521 Kirschke, S., Bousquet, P., Ciais, P., Saunois, M., Canadell, J.G., Dlugokencky, E.J.,
522 Bergamaschi, P., Bergmann, D., Blake, D.R., Bruhwiler, L., Cameron-Smith, P.,
523 Castaldi, S., Chevallier, F., Feng, L., Fraser, A., Heimann, M., Hodson, E.L.,
524 Houweling, S., Josse, B., Fraser, P.J., Krummel, P.B., Lamarque, J.-F., Langenfelds,
525 R.L., Le Quéré, C., Naik, V., O'Doherty, S., Palmer, P.I., Pison, I., Plummer, D.,
526 Poulter, B., Prinn, R.G., Rigby, M., Ringeval, B., Santini, M., Schmidt, M., Shindell, D.
527 T., Simpson, I.J., Spahni, R., Steele, L.P., Strode, S.A., Sudo, K., Szopa, S., van der
528 Werf, G.R., Voulgarakis, A., Weele, M., Weiss, R.F., Williams, J.E., and Guang, Z.,
529 2013. Three decades of global methane sources and sinks. *Nature Geoscience*. 6, 813-
530 823.

531 Lee, T.R., De Wekker, S.F.J., Andrews, A.E., Kofler, J., Williams, J., 2012. Carbon

532 dioxide variability during cold front passages and fair weather days at a forested
533 mountaintop site. *Atmospheric Environment*. 46, 405-416.

534 Lelieveld, J., 2006. Climate change: a nasty surprise in the greenhouse. *Nature*. 443,
535 405-406.

536 Li, C., Zhou, L., Qin, D., Liu, L., Qin, X., Wang, Z., Ren, J., 2014. Preliminary study of
537 atmospheric carbon dioxide in a glacial area of the Qilian Mountains, west China.
538 *Atmospheric Environment*. 99, 485-490.

539 Liu, M., Wu, J., Zhu, X., He, H., Jia, W., Xiang, W., 2015. Evolution and variation of
540 atmospheric carbon dioxide concentration over terrestrial ecosystems as derived from
541 eddy covariance measurements. *Atmospheric Environment*. 114, 75-82.

542 Martins, C.S.C., Macdonald, C.A., Anderson, I.C., Singh, B.K., 2016. Feedback
543 responses of soil greenhouse gas emissions to climate change are modulated by soil
544 characteristics in dryland ecosystems. *Soil Biology and Biochemistry*. 100, 21-32.

545 METEX (METeorological data EXplorer), 2017.
546 <http://db.cger.nies.go.jp/metex/trajectory.html> (accessed 1.18.17).

547 Nakazawa, T., Ishizawa, M., Higuchi, K., Trivett, N.B.A., 1997. Two curve fitting
548 methods applied to CO₂ flask data. *Environmetrics*. 8, 197-218.

549 Newman, S., Jeong, S., Fischer, M.L., Xu, X., Haman, C.L., Lefer, B., Alvarez, S.,
550 Rappenglueck, B., Kort, E.A., Andrews, A.E., Peischl, J., Gurney, K.R., Miller, C.E.,
551 Yung, Y.L., 2013. Diurnal tracking of anthropogenic CO₂ emissions in the Los Angeles
552 basin megacity during spring 2010. *Atmospheric Chemistry and Physics*. 13, 4359-
553 4372.

554 NOAA (National Oceanic and Atmospheric Administration, Global Monitoring
555 Division), 2016. <http://www.esrl.noaa.gov/gmd/> (accessed 2.16.17).

556 Pérez, I.A., Sánchez, M.L., García, M.Á., De Torre, B., 2009a. Daily and annual cycle
557 of CO₂ concentration near the surface depending on boundary layer structure at a rural
558 site in Spain. *Theoretical and Applied Climatology*. 98, 269-277.

559 Pérez, I.A., Sánchez, M.L., García, M.Á., De Torre, B., 2009b. CO₂ transport by urban
560 plumes in the upper Spanish plateau. *Science of the Total Environment*. 407, 4934-4938.

561 Pérez, I.A., Sánchez, M.L., García, M.A., De Torre, B., 2009c. Boundary layer structure
562 and stability classification validated with CO₂ concentrations over the Northern Spanish
563 Plateau. *Annales Geophysicae*. 27, 339-349.

564 Pérez, I.A., Sánchez, M.L., García, M.Á., Pardo, N., 2012a. Analysis and fit of surface
565 CO₂ concentrations at a rural site. *Environmental Science and Pollution Research*. 19,
566 3015-3027.

567 Pérez, I.A., Sánchez, M.L., García, M.Á., Pardo, N., 2012b. Spatial analysis of CO₂
568 concentration in an unpolluted environment in northern Spain. *Journal of*
569 *Environmental Management*. 113, 417-425.

570 Pérez, I.A., Sánchez, M.L., García, M.Á., Pardo, N., 2013. Carbon dioxide at an
571 unpolluted site analysed with the smoothing kernel method and skewed distributions.
572 *Science of the Total Environment*. 456-457, 239-245.

573 Pérez, I.A., Sánchez, M.L., García, M.Á., Ozores, M., Pardo, N., 2014. Analysis of
574 carbon dioxide concentration skewness at a rural site. *Science of the Total Environment*.
575 476-477, 158-164.

576 Pérez, I.A., Sánchez, M.L., García, M.Á., Pardo, N., 2016. Features of the annual
577 evolution of CO₂ and CH₄ in the atmosphere of a Mediterranean climate site studied
578 using a nonparametric and a harmonic function. *Atmospheric Pollution Research*. 7,
579 1013-1021.

580 Pérez, I.A., Sánchez, M.L., García, M.Á., Pardo, N., 2017. Trend analysis of CO₂ and
581 CH₄ recorded at a semi-natural site in the northern plateau of the Iberian Peninsula.
582 *Atmospheric Environment*. 151, 24-33.

583 Piao, S., Ciais, P., Friedlingstein, P., Peylin, P., Reichstein, M., Luysaert, S., Margolis,
584 H., Fang, J., Barr, A., Chen, A., Grelle, A., Hollinger, D.Y., Laurila, T., Lindroth, A.,
585 Richardson, A.D., Vesale, T., 2008. Net carbon dioxide losses of northern ecosystem in
586 response to autumn warming. *Nature*. 451, 49-52.

587 Pu, J.J., Xu, H.H., He, J., Fang, S.X., Zhou, L.X., 2014. Estimation of regional
588 background concentration of CO₂ at Lin'an Station in Yangtze River Delta, China.
589 *Atmospheric Environment*. 94, 402-408.

590 Sánchez, M.L., Ozores, M.I., López, M.J., Colle, R., De Torre, B., García, M.A., Pérez,
591 I., 2003. Soil CO₂ fluxes beneath barley on the central Spanish plateau. *Agricultural and*
592 *Forest Meteorology*. 118, 85-95.

593 Sánchez, M.L., García, M.Á., De Torre, B., Pérez, I., 2005. O₃ and CO₂ concentrations
594 in a rural area in central Spain, in: Brebbia, C.A. (Ed.), *Air Pollution XIII*. WIT
595 Transactions on Ecology and the Environment. UK. 82, 401-410.

596 Sánchez, M.L., García, M.A., Pérez, I.A., de Torre, B., 2008. Evaluation of surface
597 ozone measurements during 2000–2005 at a rural area in the upper Spanish plateau.
598 *Journal of Atmospheric Chemistry*. 60, 137-152.

599 Sánchez, M.L., Pérez, I.A., García, M.A., 2010. Study of CO₂ variability at different
600 temporal scales recorded in a rural Spanish site. *Agricultural and Forest Meteorology*.
601 150, 1168-1173.

602 Sánchez, M.L., García, M.Á., Pérez, I.A., Pardo, N., 2014. CH₄ continuous
603 measurements in the upper Spanish plateau. *Environmental Monitoring and Assessment*.
604 186, 2823-2834.

605 Sreenivas, G., Mahesh, P., Subin, J., Lakshmi, A., Narasimha, P.V., Kumar, V., 2016.
606 Influence of meteorology and interrelationship with greenhouse gases (CO₂ and CH₄) at
607 a suburban site of India. *Atmospheric Chemistry and Physics*. 16, 3953-3967.

608 Sun, Y., Bian, L., Tang, J., Gao, Z., Lu, C., Schnell, R.C., 2014. CO₂ monitoring and
609 background mole fraction at Zhongshan Station, Antarctica. *Atmosphere*. 5, 686-698.

610 Topp, E., Pattey, E., 1997. Soils as sources and sinks for atmospheric methane.
611 *Canadian Journal of Soil Science*. 77, 167-178.

612 ul-Haq, Z., Tariq, S., Ali, M., 2017. Spatiotemporal assessment of CO₂ emissions and
613 its satellite remote sensing over Pakistan and neighboring regions. *Journal of*
614 *Atmospheric and Solar-Terrestrial Physics*. 152-153, 11-19.

615 WMO Greenhouse Gas Bulletin, 2015. The state of greenhouse gases in the atmosphere
616 based on global observations through 2014. World Meteorological Organization.
617 <http://www.wmo.int/pages/prog/arep/gaw/ghg/GHGbulletin.html> (accessed 12.12.16).

618 Zeng, J., Matsunaga, T., Mukai, H., 2010. METEX-A flexible tool for air trajectory
619 calculation. *Environmental Modelling and Software*. 25, 607-608.

620 Zhang, F., Zhou, L.X., 2013. Implications for CO₂ emissions and sinks changes in
621 western China during 1995-2008 from atmospheric CO₂ at Waliguan. *Tellus, Series B*:

- 622 *Chemical and Physical Meteorology*. 65, 1-14.
- 623 Wu, J., Guan, D., Yuan, F., Yang, H., Wang, A., Jin, C., 2012. Evolution of
- 624 atmospheric carbon dioxide concentration at different temporal scales recorded in a tall
- 625 forest. *Atmospheric Environment*. 61, 9-14.

Tables

Table 1

Monthly linear regression for CO₂ (CO₂ = tx + b) at the CIBA station over the whole study period.

	CO ₂ Daytime			CO ₂ Night-time		
	t (ppm year ⁻¹)	b (ppm)	R ²	t (ppm year ⁻¹)	b (ppm)	R ²
January	2.2971	391.22	0.9975	2.2174	397.24	0.9935
February	2.3171	391.33	0.9974	2.2480	397.31	0.9941
March	2.2010	391.80	0.9987	2.1267	397.74	0.9943
April	2.2160	391.93	0.9986	2.1581	397.83	0.9947
May	2.2317	392.07	0.9985	2.1887	397.91	0.9951
June	2.2482	392.22	0.9984	2.2184	397.99	0.9955
July	2.2655	392.36	0.9983	2.2474	398.09	0.9958
August	2.2838	392.50	0.9982	2.2760	398.19	0.9961
September	2.3026	392.63	0.9981	2.3034	398.29	0.9964
October	2.2644	392.99	0.9978	2.1828	398.95	0.9924
November	2.2630	393.21	0.9979	2.1582	399.25	0.9924
December	2.2795	393.36	0.9977	2.1880	399.34	0.9930

t represents the time expressed in years, starting in 2010.

Table 2

Monthly linear regression for CH₄ (CH₄ = tx + b) at the CIBA station over the whole study period.

CH ₄ Daytime	CH ₄ Night-time
-------------------------	----------------------------

	t (ppb year ⁻¹)	b (ppb)	R^2	t (ppb year ⁻¹)	b (ppb)	R^2
January	11.3291	1867.59	0.9618	10.3677	1877.48	0.9678
February	11.7308	1867.41	0.9614	10.7056	1877.06	0.9689
March	9.3674	1873.47	0.9752	8.9775	1881.81	0.9749
April	9.6674	1873.45	0.9742	9.2678	1881.66	0.9757
May	9.9839	1873.36	0.9733	9.5630	1881.65	0.9764
June	10.3166	1873.27	0.9726	9.8627	1881.57	0.9771
July	10.6659	1873.36	0.9719	10.1672	1881.37	0.9777
August	11.0376	1873.18	0.9714	10.4816	1881.52	0.9783
September	11.4198	1873.10	0.9710	10.7955	1881.36	0.9790
October	10.7317	1876.72	0.9647	9.9236	1885.78	0.9672
November	10.6401	1878.50	0.9638	9.7580	1887.52	0.9660
December	10.9735	1878.74	0.9627	10.0580	1887.68	0.9669

t represents the time expressed in years, starting in 2010.